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Modification in light absorption cross section of laboratory-generated Black Carbon-Brown Carbon particles upon surface reaction and hydration

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Key words: black carbon, brown carbon, light absorption, mixing state, surfacereaction

18 Abstract

Recent reports indicate that the absorption enhancement of black carbon (BC) and 19 brown carbon (BrC) particles is determined by the evolution of morphology and 20 21 mixing state during the atmospheric processing. In this study, laboratory-generated 22 BC-BrC mixture particles (BC-BrC) were exposed to sulfuric acid, ammonia/triethylamine, and water vapor sequentially to investigate the alteration in 23 light absorption, morphology and mixing state during simulated atmospheric 24 25 processing. Condensation of sulfuric acid and exposure to ammonia or triethylamine 26 at 5% relative humidity decreased light absorption cross section of BC-BrC by 13%-26%, but subsequent hydration at 85% relative humidity led to an increase by 27 5%-20%. Our results show that surface reactions and hydration can significantly alter 28 29 the light absorption cross section of BC-BrC. Our observation suggests that the 30 restructuring of BC-BrC induced by neutralization reactions on the BC-BrC surface was responsible for the reduction in light absorption cross section at low relative 31 32 humidity. On the other hand, the formation of an aqueous coating through water uptake by the hydrophilic inorganic coating at high relative humidity caused an 33 absorption enhancement. 34

35 **1. Introduction**

Black carbon (BC) has now been accepted as a strong absorber of visible solar 36 radiation, leading to a positive (warming) radiative forcing that is estimated to be 37 38 about one quarter to one half of the magnitude of atmospheric CO_2 (Bond et al., 2013; IPCC, 2007; Ramanathan and Carmichael, 2008). Recently, another class of 39 light-absorbing carbon-containing substances (organic carbon, OC) in atmospheric 40 aerosols, namely brown carbon (BrC), arises much scientific interests because its 41 intense absorption at the shorter visible and UV wavelength may lead to 42 43 underestimation of absorption of solar radiation in the troposphere (Andreae and Gelencsér, 2006). 44

Annually, 8.0 Tg BC and 33.9 Tg OC are estimated to be emitted at a global scale, 45 and the contributions of biomass/biofuel burning to BC and OC are estimated to be 62% 46 47 and 93%, respectively (Bond et al., 2004). Condensable organic species produced from either biomass burning (Lewis et al., 2008) or diffusion flames (Schnaiter et al., 48 2006), characterized by absorption in the visible light range with an absorption 49 50 Ångström exponent (AAE) greater than unity, has been linked to the source of BrC. The presence of both BC and BrC in particulate matters formed from biomass/biofuel 51 burning at flaming and smoldering conditions complicates the estimation of their 52 climate impacts (Chakrabarty et al., 2010; Kirchstetter et al., 2004; Lack et al., 2012). 53

Atmospheric processing poses additional uncertainties in estimating the radiative forcing of BC particles. Once emitted into the atmosphere, BC particles undergo aging processes including condensation of low volatility species, coagulation with pre-existing particles, and potentially, cloud processing. Over a timescale of a few hours to a few days, the aging processes lead to an internal aerosol mixture. Nonrefractory coating in the internally mixed BC particle is believed to act like a lens 60 that enhances the light absorption by directing more light towards to the BC core ("lensing effect"), when compared with an equivalent external mixture (Jacobson, 61 62 2001). Theoretically, the extent to which absorption of BC can be enhanced depends critically on the ratio of the mass of nonrefractory particulate matter ([NR-PM]) to 63 that of BC ([BC]) in the internally mixed particle ($R_{BC} = [NR - PM]/[BC]$) (Bond et 64 al., 2006). A number of ambient studies on optical properties of aged BC (which 65 usually contained amounts of BrC) measured absorption enhancement (Eabs) that 66 differed in a large range from 1.06 to 1.7 with a large scale of R_{BC} (Cappa et al., 2012; 67 Knox et al., 2009; Lack et al., 2012). The discrepancy can be partly explained by the 68 different particle configuration caused by various atmospheric aging processes 69 (Adachi and Buseck, 2013; Schnaiter et al., 2003; Schnaiter et al., 2005). 70

71 Previous laboratory studies focused on absorption amplification of BC upon 72 condensation of sulfuric acid and organic materials on BC surface (Cross et al., 2010; Schnaiter et al., 2005; Shiraiwa et al., 2010; Zhang et al., 2008). However, the 73 condensed materials are subject to further atmospheric processing, such as surface 74 reactions with gaseous species (Biskos et al., 2009; Wang et al., 2010) and hydration 75 and dehydration cycles (Khalizov et al., 2009; Zhang et al., 2008). Water soluble 76 77 coating (such as sulfuric acid and glutaric acid) leads to hydration and morphology 78 modification of BC, both of which further amplify their absorption abilities (Khalizov et al., 2009; Mikhailov et al., 2006). 79

As one of the most abundant secondary chemical components of atmospheric aerosols, sulfate is usually found to be internally mixed with BC. They are believed to alter the morphology, hygroscopicity, and optical properties of BC particles based on the field observation (Fu et al., 2012; Li et al., 2013). However, there were only a few studies on BC internally mixed with ammonium sulfate/bisulfate generated either by 85 coagulation (Schnaiter et al., 2003) or by an "evaporation-condensation" procedure (Stratmann et al., 2010). In the real atmosphere, sulfuric acid efficiently sinks on the 86 BC surface (Riemer et al., 2004; Zhang et al., 2008) and subsequently 87 ammonia/amine is transferred to the particulate phase in an effort to neutralize the 88 acidic components forming ammonium/aminium sulfate (Seinfeld and Pandis, 2006). 89 Neutralization of the acidic coating by base gaseous species bring additional mass into 90 the nonrefractory coating (i.e. increase in R_{BC}), which can lead to absorption 91 amplification of BC. On the other hand, neutralization, i.e. transformation of sulfuric 92 93 acid to ammonium/aminium sulfate/bisulfate could modify the hygroscopicity, density and phase of the overall coating (Chan and Chan, 2012; Qiu and Zhang, 2012; Tang 94 95 and Munkelwitz, 1994). The influence of neutralization following the condensation of 96 sulfuric acid on BC surface at different relative humidity (RH) is still unclear.

97 In this study, BC and BrC mixture particles (BC-BrC) were artificially produced 98 from fuel-rich flames. Polycyclic aromatic hydrocarbons (PAHs), as the main OC components of our BC-BrC, were assigned to BrC because of their light absorption 99 100 nature in line with previous investigations on particles from both fuel-rich flames (Apicella et al., 2004; Schnaiter et al., 2006) and wood combustion (Chen and Bond, 101 2010). We exposed our BC-BrC particles to sulfuric acid (H₂SO₄, SA), ammonia 102 (NH₃, AM) or triethylamine ((CH₃CH₂)₃N, TEA), and water vapor sequentially to 103 investigate the absorption and morphology alteration. The procedure simulated 104 105 transfer of ammonia/amine to the particulate phase by neutralization with condensed sulfuric acid on the BC-BrC surface and potential hygroscopic growth of internally 106 mixied BC-BrC due to hydrophilic inorganic coating in the atmosphere. Transmission 107 electron microscope (TEM), energy dispersive X-ray spectroscopy (EDX), 108 thermal-evaporation, and Mie calculation using a "core-shell" configuration were 109

employed to investigate mechanisms for observed property variation. In addition,
potential bias on absorption enhancement owing to the instrumental uncertainties was
discussed.

113

114 2. Experimental Methods

115 **2.1. Particle generation and characterization**

BC-BrC particles were generated in a laminar diffusion burner by combustion of methane (99.999%) at a flow rate of 0.2 L min⁻¹ and air (99.999%) of 0.5 L min⁻¹. A flow of 0.05 L min⁻¹ was sampled from the chimney and immediately diluted by 20-fold with nitrogen (99.999%), dried to < 0.5% RH by a silica gel diffusion drier and a Nafion tube (Perma pure), and then directed to further experimental usage. Authentic polystyrene latex spheres (PSL, Thermo Scientific) particles were generated using an atomizer (TSI 3670).

The extinction spectra and the chemical composition of BC-BrC were measured using an UV-Vis spectrophotometer and a gas chromatography–mass spectrometry (GC-MS), respectively (more details were described in the supplementary information, SI).

Fresh and processed particles were deposited onto a copper grid coated with a carbon film using a single-stage cascade impactor with a 0.5-mm diameter jet nozzle (Li and Shao, 2010), and analyzed by a transmission electron microscope (Jeol JEM-2100F) at an accelerating voltage of 200KV and by an energy-dispersive X-ray spectrometer (EDX). A collection time of 30 s for each EDX spectrum was employed.

132 2.2. Processing of BC-BrC particles

133 Size distributions, mass, and hygroscopicity of particles were measured using a

system comprising two differential mobility analyzers (DMA, TSI 3081), an aerosol 134 particle mass analyzer (APM, Kanomax 3601), and a condensation particle counter 135 (CPC, TSI 3771), as shown in the supplementary scheme. Briefly, polydisperse 136 particles were brought to charge equilibrium by a Kr-85 neutralizer (TSI 3077), and 137 then size-selected (D_0) by applying a fixed voltage to DMA1 (TSI, 3081). The 138 obtained monodisperse particles (with a geometric standard deviation of 1.07) were 139 exposed to sulfuric acid vapor in a 100 ml reservoir containing 20 ml 86 wt% H₂SO₄ 140 solution (25±0.1 °C), directed to a 50 cm long 10-cm-i.d. residence chamber (growth 141 142 reactor), and then through a potassium hydroxide (KOH) scrubber, in which excess H₂SO₄ vapor was removed to terminate its condensation. The sulfuric acid-coated 143 particles were exposed to 1 ppm NH₃ or (C₂H₅)₃N in another 100 ml residence 144 145 chamber and then to an elevated-RH environment (5%-85% RH) in a multitube Nafion humidifier (Perma Pure), in which the RH was regulated between 5% and 85%. 146 In the tandem DMA mode (TDMA), the hygroscopic growth factor ($G_{f, RH}$), defined as 147 the ratio of the processed particle diameter (D_p) to the initial one (D_0) , was obtained 148 through the inversion of scanning DMA2 data by the TDMA fit program: 149

$$150 \qquad G_{f,RH} = \frac{D_p}{D_0}.$$

-

The sheath flow in both DMAs was maintained at 6.5 L min⁻¹. The aerosol sample flow through DMAs was kept at 1.0 L min⁻¹. The RH was monitored both on the inlet and outlet of DMA2 by two probes (Vaisala HMM22D). The sheath gas in DMA2 was always humidified to an identical RH to the sample stream to prevent a RH gradient.

156 With the presence of a 250 °C thermal-denuder (TD) placed before DMA2, the 157 shrink in particle size was quantified as $D_{p,250^{\circ}C}/D_0$ (referred as "heating

experiment" in the following text). Thermally evaporated organic and inorganic 158 species might re-condense on the particle surface. In our case, re-condensation was 159 expected to be negligible because of the dilution of the sample flow by DMA2 sheath 160 flow and the relatively larger surface areas of the tubing. Additionally, charring of OC 161 was expected insignificant at this temperature (see details in SI). 162

An APM interfaced with DMA and CPC was used to measure the monodisperse 163 particle mass (m_n) that can be derived from the Gauss fit on the mass distribution 164 recorded in APM-CPC. The mass-mobility scaling exponent (D_m) was derived from 165 the mass mobility relationship $m_p \propto d_m^{D_m}$, where d_m was the mobility diameter 166 determined by DMA1. The effective density was derived from the equation 167 $\frac{6m_p}{\pi d_m^3}$.

168
$$\rho_{eff} =$$

Care was taken to minimize adventitious NH₃ in the system in order to prevent 169 uncontrolled formation of ammonium sulfate. Ultrahigh pure N₂ (99.999%) was 170 171 employed for the flow system. The carrier flow of polydisperse aerosol, the sheath 172 flow in DMAs, and the purge flow in Nafion tubes all passed through NH₃-scrubbers, which consisted of a 5 liter-cylindrical reactor filled with granular citric acid. To avoid 173 temperature fluctuation, the entire setup was placed in a temperature controlled box 174 $(25 \pm 0.1 \text{ °C})$, and all the connecting tubing was wrapped by thermal-insulated 175 176 materials.

177 2.3. Optical properties measurement

Optical properties of fresh and processed monodisperse particles with an initial 178 diameter of 150 nm were measured using a combination of a self-made cavity 179 180 ring-down spectroscopy (CRDS) and a nephelometer (TSI 3563) interfaced to a CPC (TSI 3776) (Li et al., 2011; Li et al., 2013). Total extinction coefficient ($\sigma_{\rm ext}$) was 181

determined at 532 nm by the CRDS, and total scattering coefficient was measured at 450, 550, and 700 nm by the nephelometer. The scattering coefficient was corrected for angular nonidealities using the scheme reported in Anderson and Ogren (1998) and converted to 532 nm (σ_{scat}) using the Ångström exponent equation:

186
$$\qquad \mathring{a} = -\frac{\log(\sigma_{scat,532}/\sigma_{scat,550})}{\log(532/550)}$$

In the present study, the method yielded a correction factor of 1.00±0.01. The
scattering Ångström coefficients ranged from 3.43 to 3.56 for fresh and processed
BC-BrC (details in SI).

190 Particle number concentration measured by CPC was used together with 191 extinction and scattering coefficient to calculate the corresponding cross section: 192 $C_{\text{ext/scat}} = \frac{\sigma_{\text{ext/scat}}}{N}$, where *C* was the cross section and *N* was the particle number 193 concentration corrected with the diffusion loss in either CRDS or nephelometer.

The absorption coefficient (σ_{abs}) at 532 nm was then calculated from the 194 difference between the extinction and scattering coefficient. By using this 195 196 "extinction-minus-scattering" method, the absorption coefficient could be overestimated at high relative humidity. With up to 2 °C difference between the inlet 197 and the body of nephelometer, the humidity in the body was estimated to be 78% 198 when the inlet gave a RH of 85%. Since alteration in complex refractive indices was 199 expected to be low over this RH range (78%-85%), the scattering coefficient solely 200 scaled according to the 4th or 5th power with the particle diameter in this size range. 201 The sulfate coatings were expected to be aqueous in 78%-85% RH range since the 202 highest efflorescence RH among the three forms of coating was reported ~35% for 203 ammonium sulfate (Tang and Munkelwitz, 1994). By estimating the particle size 204 decrease from the interpolated hygroscopic growth curve measured by TDMA (see 205

section 3.4), measured scattering (78% RH in the body) was corrected by a factor of

207 1.07-1.09 according to
$$\left(\frac{D_{p,85\%}}{D_{p,78\%}}\right)^5$$
 to derive that at 85% RH. On the other hand, such

temperature difference-induced correction is unnecessary for CRDS measurements.
Hence, absorption cross section and single scattering albedo (SSA) are values at 85%
RH.

211 SSA was calculated according to $SSA = C_{scat}/C_{ext}$. Enhancement of absorption

212 during the aging was obtained according to
$$E_{abs} = \frac{C_{abs, \text{processed BC-BrC}}}{C_{abs, \text{fresh BC-BrC}}}$$

Overestimation of singly charged particles of 150 nm from DMA1 due to doubly charged particles was ~1% as calculated from the particle size distribution (Figure S1) and the charging efficiency. Bias from doubly charged large particles was <7% and <3% in scattering and absorption measurement, respectively. No correction due to multiply charged particles was applied.

218

219 **3. Results and Discussions**

220 **3.1. Basics of BC-BrC particles**

BC-BrC particles in the size range of 30-250 nm (a count median diameter at 69 nm and a geometric standard deviation σ_g of 1.24) were generated from a fuel-rich flame (V_{methane}/V_{air} = 0.4) in a laminar diffusion burner. An extinction Ångström coefficient of 2.9 was determined for the emulsion of BC-BrC in dichloromethane in the wavelength range of 250-800 nm. This significant spectral dependence was mainly attributable to the absorption by the dissolved OC fraction assigned to BrC in accordance to a series of PAHs detected by GC-MS. Similar spectral dependence due to BrC fraction with similar chemical composition has been also observed for woodcombustion aerosol (Chen and Bond, 2010).

Mass of monodisperse BC-BrC particles (40, 100 and 150 nm) was measured by 230 the DMA-APM system. The obtained mass-mobility scaling exponent of 2.9 231 232 suggested near-spherical (3.0 for a sphere) rather than fractal-like shape (Slowik et al., 2004). The effective density was determined to be 1.10-1.27 g/cm³, with smaller 233 values for larger particles. The values are smaller than the recommended BC material 234 density of 1.8 g/cm³ (Bond and Bergstrom, 2006) and typical PAHs density of 1.3 235 g/cm^3 , but larger than 0.8 g/cm^3 for compact soot aggregates obtained through 236 condensation of sulfuric acid and subsequent heating (Khalizov et al., 2009), 237 indicating either that internal voids existed or that the particles comprised a great 238 amount of aliphatics (typical density of 0.8 g/cm³). Since combustion particles 239 produced under similar conditions contains minor amounts (~5%) of aliphatics 240 (Slowik et al., 2004), which was also evidenced from our GC-MS analysis, we 241 speculate that internal voids were present in our BC-BrC particles in spite of their 242 243 near-spherical shape. By applying and not applying a thermal denuder in front of APM, organic species in BC-BrC volatilized at 250 °C in a N2 atmosphere was 244 determined to be 34%-54%, with smaller values for larger particles. 245

246

3.2. Alteration in optical properties

We exposed monodisperse BC-BrC particles (150 nm) to H_2SO_4 vapor, eliminated excess H_2SO_4 vapor using a KOH trap, exposed the H_2SO_4 -coated particles (BC-BrC-SA) to NH₃ or (CH₃CH₂)₃N gas, and modulated the concentration of water vapor to allow any potential hygroscopic growth. Excess base gases were retained in the system to ensure base gases were always available for neutralization during the entire morphology transformation process of BC-BrC-SA. Optical properties of fresh

and processed particles were then measured using a combination of nephelometer and 253 CRDS at dry (5% RH) and wet (85% RH) conditions, respectively (Fig. 1a). Fresh 254 BC-BrC showed a mass absorption efficiency (the ratio of absorption cross section to 255 the mass of BC-BrC particle) of $3.2\pm0.5 \text{ m}^2 \text{ g}^{-1}$ at 532 nm, and a mass scattering 256 efficiency of 2.7±0.3 m² g⁻¹. The mass absorption efficiency of BC-BrC is much lower 257 than that of typical fresh soot because of the presence of organic materials (Schnaiter 258 et al., 2006), resulting in a single scattering albedo of 0.46±0.04 that is significantly 259 larger than fresh soot (0.2-0.3) (Bond and Bergstrom, 2006). 260

When BC-BrC were coated with H₂SO₄, and subsequently completely or partially 261 neutralized with NH₃ or $(C_2H_5)_3N$, the fraction of nonrefractory coating led to the 262 variation in light absorption, which depended on the coating components and the 263 relative humidity (Fig. 1b, and Table S2). At 5% RH, E_{abs} showed values of 0.87, 0.85, 264 265 and 0.74 for three coatings, respectively, suggesting that condensation of inorganic acid and subsequent surface neutralization reactions were able to decrease the light 266 absorption cross section of BC-BrC at dry conditions. The observation was contrary to 267 the theoretical prediction that addition of nonrefractory coating would lead to increase 268 in light absorption. The reduction in light absorption was plausibly attributable to 269 restructuring of BC-BrC caused by the processing. Restructuring of fractal-like BC 270 act either to amplify light absorption owing to enhanced interaction between 271 neighboring spherules, or to reduce absorption due to inner spherules shielded by the 272 273 outer ones (Liu et al., 2008). The shielding effect appears to be significant for particles with volume equivalent diameter >180 nm or >80 nm suggested by 274 theoretical prediction (Liu et al., 2008) or experimental investigation (Khalizov et al., 275 276 2009), respectively. In the present study, condensed OC components might deprive the fractal-like shape but not necessarily compressed the BC inclusion into spheres in 277

fresh BC-BrC particles, which accorded with from Slowik et al. (2007) that
anthracene didn't cause measurable restructuring of fractal like soot aggregates.
Restructuring of BC-BrC and BC inclusion was evidenced by heating experiment
(section 3.5) in spite of their near-spherical shape.

282 An elevation in relative humidity to 85%, however, led to E_{abs} of 1.05, 1.20 and 283 1.18, respectively, in accordance with previous studies that hydration(Schwarz et al., 2008) can enhance light absorption of internally-mixed hydrophilic soot (Khalizov et 284 285 al., 2009; Mikhailov et al., 2006). Condensed sulfuric acid or its reaction products with base gases can uptake amounts of water at 85% RH and form transparent shell 286 over BC-BrC particles, which is expected to enhance the lensing effect that surpassed 287 the shielding effect of BC-BrC restructuring, resulting in a net absorption increase. On 288 the other hand, Lewis et al. (2009) reported a reduction in light absorption of biomass 289 290 burning aerosol upon humidification, which was hypothesized to arise either from the 291 shielding effect due to particle collapse induced from water uptake by intrinsic inorganic components, or from the instrumental artifact due to photoacoustic heat and 292 mass transfer. 293

294 The scattering of BC-BrC particles did not change at 5% RH but increased by ~1.1-fold at 85% RH (Fig. 1a), which was much weaker compared with previous 295 studies on glutaric acid- or H₂SO₄-coated chain BC agglomerates (Khalizov et al., 296 2009; Mikhailov et al., 2006). Fresh BC agglomerates consisted of small primary 297 spherules (15-25 nm) that weakly scattered light, and the restructuring and volume 298 299 growth due to condensation and hydration led to a more compact shape with much stronger scattering (Khalizov et al., 2009), which did not occur for processed BC-BrC 300 301 particles in this study.

302

303 **3.3. TEM and EDX analysis**

TEM and EDX were used to study the morphology of BC-BrC. The majority of fresh 304 BC-BrC indicated a configuration similar to embedded soot that is heavily coated by 305 condensed organic materials emitted from wildfire (China et al., 2013), while a small 306 fraction showed partially coated chain agglomerates (Fig. 2b) as aggregates of 307 primary particles (China et al., 2013). Due to the deformation of particles during the 308 309 collection and under TEM vacuum, the internal voids of fresh BC-BrC cannot be visualized under TEM. The H₂SO₄ coating did not alter particle morphology 310 311 significantly, and the H₂SO₄ coating itself was not quite visualized in TEM images because of the evaporation of H₂SO₄ under vacuum (Fig. 2c). Neutralization by NH₃ 312 or (C₂H₅)₃N, however, brought on identifiable coating on BC-BrC particle surface 313 314 (Figs. 2d and 2e), which was likely at least partially responsible for the modified optical properties of particles. Note that the BC-BrC-SA-AM particle in Fig. 2d was 315 exposed to 85% RH before collection, and hence the circular spot around the particle 316 317 was attributable to a residue of ammonium sulfate after the evaporation of most of the ammonium sulfate solution. The image might reflect an ideal "core-shell" 318 configuration with BC-BrC located at the center and inorganic coating in the 319 surrounding. EDX spectra revealed the presence of various elements (Figs. 2f and 2g). 320 321 Carbon elements mainly came from BC-BrC, while oxygen, silicon, and copper were 322 from the Si-O substrate and the impactor grid (Li and Shao, 2010). In contrast to TEM tests, minor sulfur element was evidenced in EDX spectra for BC-BrC-SA, suggesting 323 the presence of H₂SO₄ residue after evaporation of most condensed H₂SO₄ under 324 325 vacuum, or the presence of other sulfur-containing species. Sulfur element was more abundant in EDX spectra for BC-BrC-SA-AM, since H₂SO₄ was neutralized to 326 327 less-volatile (NH₄)₂SO₄. Restricted by detection limit and resolution, N elements from

ammonium ions were undetectable by EDX (Li and Shao, 2010).

329 **3.4.** Physicochemical properties of hydrophilic inorganic coating

In Figure 3, the size variation of 150 nm BC-BrC was plotted, as well as that of 330 authentic PSL particles as a reference, when processed with H₂SO₄, NH₃ or 331 (CH₃CH₂)₃N, and water vapor. Similar figures for 40 and 100 nm particles are 332 presented in Fig. S4. Except for two cases of 40 nm particles at the highest RH (Fig. 3 333 and Fig. S4), the mobility diameters of processed BC-BrC were always smaller than 334 those of PSL, indicating morphology alteration during the processing including 335 336 condensation, surface reaction, and hydration. Morphology alteration can be caused by combined effects of particle restructuring and a decrease in the dynamic shape 337 factor. The latter was explained by the padding of small cavities by inorganic coating 338 339 on the irregular particle surface. However, the actual cause for the decrease in mobility size of BC-BrC-SA, -SA-AM, and -SA-TEA was concealed because the 340 alteration in effective density of BC-BrC due to different processing was unavailable. 341

As shown in Fig. 3 and in Fig. S4, BC-BrC-SA, -SA-AM and -SA-TEA have 342 shown distinctive, i.e., either continuous or deliquescent, hygroscopic growth curves, 343 which was largely determined by the nature of inorganic coatings. As reported by 344 Stratmann et al. (2010), evaporation-condensation procedure generates ammonium 345 bisulfate coating rather than ammonium sulfate. Without a direct measurement, the 346 chemical composition in the coating was then evaluated by combining the 347 348 hygroscopic growth factors of processed PSL and physicochemical properties of ammonium/aminium sulfates/bisulfates derived from E-AIM model (see details in SI). 349

The amounts of sulfate ions in the coating of processed PSL were identical for PSL-SA at the RH range of 30%-85% and PSL-SA-AM/TEA at 85% RH, *i.e.*, the occurrence of deliquescence and the formation of an ammonium or triethylaminium sulfate (AS or TEAS) solution (Table S3a). The seeming larger number of sulfate ions for NH₃ exposure at 5%, 30%, and 70% RH was rationalized by the AS coating being in fact crystalline and nonuniform under dry condition. In this case, the diameter sized by DMA was larger than that of a sphere equivalent in volume (Gibson et al., 2007), which resulted in an effective density of AS coating less than that of authentic standards (Table S3a). Hence, within experimental uncertainty, NH₃ gas was able to completely neutralize coated H_2SO_4 at all relative humidity used.

360 In contrast, the calculated number of sulfate ions for (CH₃CH₂)₃N exposure at 5%, 30% and 70% RH was smaller than those in the H_2SO_4 coating if complete 361 neutralization was assumed (Table S3b and Fig. S5). This discrepancy suggested that 362 H₂SO₄ was only partially neutralized by (CH₃CH₂)₃N at low humidity, because the 363 initial reactions between condensed H₂SO₄ and gaseous (CH₃CH₂)₃N led to formation 364 365 of triethylaminium bisulfate/sulfate (TEABS/TEAS) on the particle surface and the steric hindrance of ethyl groups in TEABS/TEAS prevented gaseous (CH₃CH₂)₃N, 366 although still available, from diffusing deep into H₂SO₄ liquid layer to achieve 367 complete neutralization. Further uptake gaseous (CH₃CH₂)₃N and complete 368 neutralization did not occur until high relative humidity was reached (Chan and Chan, 369 370 2012).

Hence, the formation of AS on PSL surface is responsible for the deliquescent behavior for PSL-SA-AM, whereas the evolving chemical composition and hydrophobic ethyl group terminal explains the non-growth for PSL-SA-TEA until 85% RH in contrast to continuous hygroscopic growth of atomized particles (Qiu and Zhang, 2012), or droplets from TEAS or TEABS solutions (Chan and Chan, 2012). Once the relative humidity rose to 85%, a solution layer of AS or TEAS was formed on the BC-BrC particle surface and demonstrated mainly its lensing effect.

3.5. Thermal removal of volatile compounds 378

Fresh and processed BC-BrC particles was heated up to 250 °C in a N₂ 379 atmosphere and the variation in particle diameter was measured to further investigate 380 381 the morphology variation of the refractory core of BC-BrC particles during the aging process (Fig. 4). Heating led to removal of semi-volatile and/or low volatility 382 compounds, and the extent of shrinking in size decreased with increasing D_0 . Coating 383 of H₂SO₄ followed by heating, *i.e.*, heating of BC-BrC-SA, however, resulted in larger 384 particles than heated BC-BrC. Since sulfuric acid is expected to evaporate at 250 °C 385 386 (Cross et al., 2010), the relative increase in diameter can be explained by formation of small amounts of refractory sulfur-containing species upon coating of H₂SO₄ or 387 during heating, as suggested by our EDX spectra. 388

In addition, heated BC-BrC-SA were generally larger than heated 389 390 BC-BrC-SA-AM and BC-BrC-SA-TEA, suggesting BC-BrC collapsed to various extent owing to the neutralization by NH₃ and (CH₃CH₂)₃N. The extent of reduction 391 in light absorption cross section for BC-BrC-SA-AM/-TEA coincided with the extent 392 393 of shrinking of refractory core, indicating that collapse of refractory core (i.e. BC inclusion) was likely responsible for the reduction in light absorption cross section at 394 5% RH. Exposure of fresh and coated particles to 85% relative humidity before 395 heating led to further shrinking of particles, however, the extent is rather small 396 comparing to that due to coating. Hence, compacting of BC-BrC exerted a more 397 398 profound role at the initial stage of our simulated aging, *i.e.*, condensation and surface reactions. 399

400

3.6. Mie theory calculation with a core-shell configuration

401 Mie theory with a core-shell configuration has been widely applied to evaluate the absorption alteration of internally mixed BC (Bohren and Huffman, 2007; Bond et 402

al., 2006; Schnaiter et al., 2005). In our study, the effect of different sulfate coatings
on absorption enhancement was investigated using this method. Briefly, the model
utilized a volume equivalent diameter of 132.7 nm estimated from the effective
density and OC fraction, a complex refractive index of 2.0+0.2i for the BC-BrC core
retrieved from Mie "closure" calculations using measured absorption and scattering
coefficients (Chakrabarty et al., 2010), and the complex refractive indices of sulfates
coating according to the volume weighted mixing rule (see more details in SI).

Our calculation results are presented in Fig. 5. Generally, E_{abs} was always larger than 1 and increased monotonously in the entire range of R_{BC-BrC} in line with Bond et al. (2006). Hydration can significantly enhance light absorption by increasing the coating mass; on the other hand, it can narrow the difference in E_{abs} among different forms of sulfates because their water uptake ability was close at high humidity.

Our measured E_{abs} is presented in Fig. 5 for comparison. The nonrefractory 415 coating mass was derived from that of processed PSL (see section 3.4 and Table S3). 416 The overall uncertainty of R_{BC-BrC} was estimated to be 30%, taking into account the 417 418 surface nature and morphology. The uncertainty of E_{abs} was estimated to be equal to the relative errors of σ_{abs} measurement (see more details in SI). Clearly, the 419 "core-shell" configuration calculation fairly reproduce the Eabs at 85% RH but 420 421 overestimates E_{abs} at 5% RH, likely because the model didn't involve other key facts including particle shape alteration and interaction between OC and inorganic coatings. 422 Schnaiter et al. (2005) also reported less well reproducibility of Mie theory calculation 423 for SOA-coated BC than for pure SOA, most likely because of the restructuring and 424 the incomplete enclosure of the porous soot aggregates. 425

426

428 **4.** Conclusion

Our work on the multistep processing of BC-BrC, which utilized condensation of 429 H₂SO₄ and subsequent neutralization and hydration, provides an attempt to mimic the 430 431 real sequence of atmospheric aging events and to investigate the corresponding alteration in BC-BrC's light absorption, as shown in Fig. 6. Our study suggests that 432 surface reactions can modify the mixing state and morphology of BC-BrC. The 433 restructuring of BC-BrC largely determines the light absorption ability at low 434 humidity even if neutralization of SA-coating by NH₃ or (CH₃CH₂)₃N causes mass 435 growth of nonrefractory coating. Heating experiments could not verify he 436 restructuring of BC-BrC induced by condensation of sulfuric acid because of the 437 formation of refractory sulfur-containing species, while it was evident that 438 439 neutralization of NH₃ or (CH₃CH₂)₃N with condensed H₂SO₄ act to compact the BC 440 inclusion. Meanwhile, ammonium sulfate coating originated from reaction between SA and NH₃ on the surface was evidenced to be nonuniform, which was different 441 442 from the core-shell configuration and correspondingly hindered the lensing effect. In addition, the effect of high relative humidity in the evolution of internally mixed BC 443 and/or BrC has been investigated in this work. Above the deliquescence relative 444 humidity, hydrophilic sulfate coating will exert primarily lensing effect that enhances 445 446 absorption although shielding effect is also occurring.

The presence of OC in BC-BrC particles and their interaction with coating materials turned to be a source of the complexities. Henning et al. (2012) recently reported an increase of the hygroscopic growth with increasing OC-content for CAST-soot internally mixed with sulfuric acid, which arose the hypothesis that sulfuric acid reacts with PAHs and form hydrophilic products. The speculated reaction between sulfuric acid and PAHs would modify the light absorption in the UV-Vis 453 range for BC-BrC particles in addition to their hygroscopicity. In our study, the 454 discrepancy between experimental data and Mie theory calculation with a core-shell 455 model configuration might originate from the interaction between OC and inorganic 456 coating under different humidity conditions in addition to the BC-BrC restructuring.

457 It is now accepted that BC and/or BrC, freshly emitted from incomplete combustion processes, will become eventually internally mixed with secondary 458 organic aerosols, sulfates, and nitrates (Moffet and Prather, 2009; Posfai et al., 1999). 459 460 Our study suggested that various chemical processing largely determined the alteration in mixing state, morphology and light absorption of BC-BrC. Hence, other 461 important surface reaction processes, e.g., the formation of ammonium nitrate 462 following the sulfate coating, should be investigated (Riemer et al., 2004). In addition, 463 the extent of alteration in light absorption cross section for BC-BrC should be 464 cautiously extrapolated to their real climate impact because internally mixed BC/BrC 465 usually contained more nonrefractory materials, i.e. large R_{BC}. Clearly, the whole 466 picture of the evolution of BC/BrC is yet to be completely understood. 467

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469 Acknowledgements

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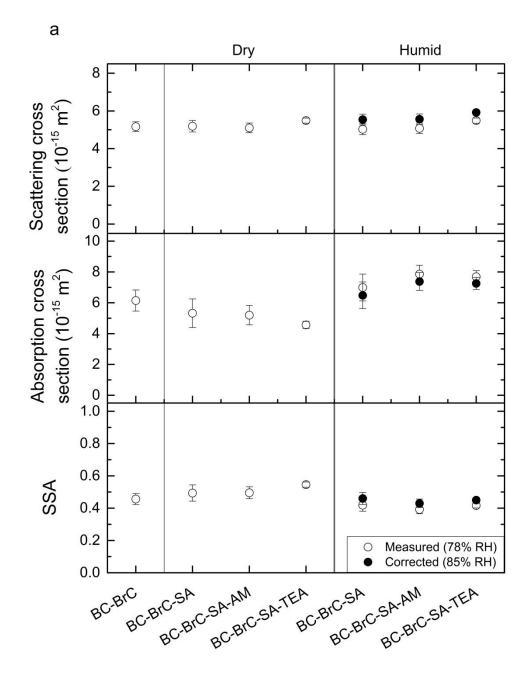
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637 **Figure Captions**

Fig. 1. Optical properties of BC-BrC. (a) scattering and absorption cross sections, and single scattering albedo (SSA) of BC-BrC, BC-BrC-SA, BC-BrC-SA-AM, BC-BrC-SA-TEA particles at 5 % and 85% RH, respectively. The error bars represent one standard deviations (σ). The open circles represent direct measurement, while the solid circles are corrected from 78% to 85% RH. (b) Absorption enhancement (E_{abs}) as a function of RH and R_{BC-BrC}.





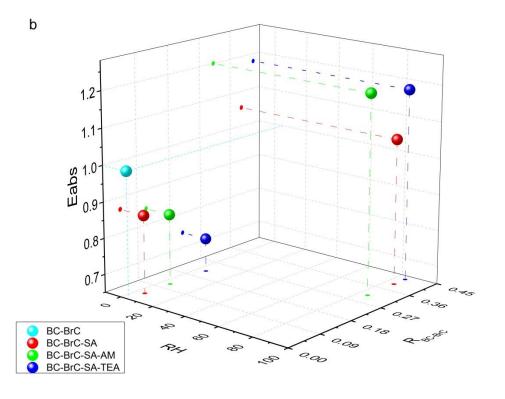




Fig. 2. TEM images of different types of BC-BrC: (a) embedded/heavily coated fresh
BC-BrC that represented the majority of collected particles; (b) partially coated chain
agglomerates that were a minor fraction of collected particles; (c) BC-BrC-SA; (d)
BC-BrC-SA-AM; (e) BC-BrC-SA-TEA. EDX spectra of (f) BC-BrC-SA and (g)
BC-BrC-SA-AM. BC-BrC-SA-AM particles were exposed to 85% RH before
collection, while other particles were sampled at 5% RH.

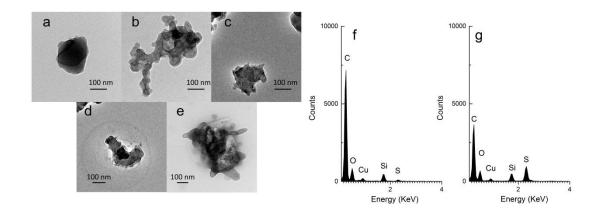
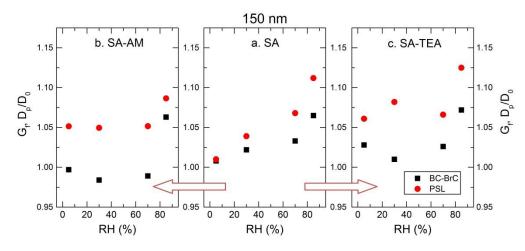


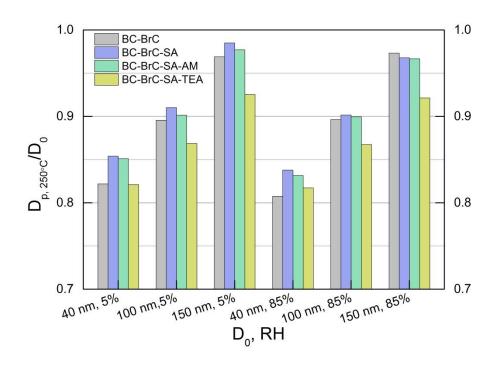
Fig. 3. Hygroscopic growth factor (G_f) of 150-nm BC-BrC. (a), SA, upon condensation of $H_2SO_{4;}$ (b), SA-AM, exposure of H_2SO_4 -coated BC-BrC with $NH_{3;}$ (c), SA-TEA, exposure of H_2SO_4 -coated BC-BrC with (CH_3CH_2)₃N.

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Fig. 4. Shrink in size $(D_{p,250^{\circ}C}/D_0)$ measured by DMA-TD-DMA at 250°C. The TD was set at 250°C to allow complete evaporation of non-refractory components. For high RH experiments, a humidifier was placed between the first DMA and TD to elevate RH up to 85%.



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- **Fig. 5.** E_{abs} for processed BC-BrC under dry (dash line) or wet (solid line) conditions calculated using Mie theory, and experimentally observed E_{abs} for BC-BrC-SA (black squares), BC-BrC -SA-AM (red squares) and BC-BrC-SA-TEA (green squares). The circles at the end of curves correspond to R_{BC-BrC} with ~10-fold sulfate ions of the experimental data in this study.
- 675

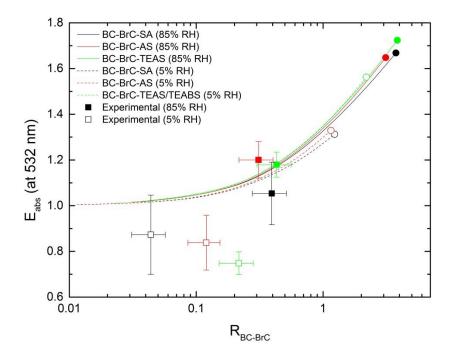
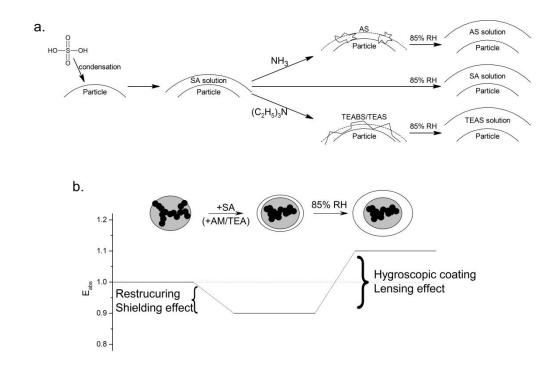


Fig.6. (a), Processing processes for generation of SA, SA-AM and SA-TEA coatings on BC-BrC or PSL particles. (b), Proposed mechanism for alteration in light absorption for internally mixed BC-BrC upon surface reaction and hydration. The black aggregates represent refractory BC inclusion, and the gray and clear circles represent organic and inorganic materials, respectively.



Highlights:

• Condensation of H_2SO_4 , surface neutralization reactions and subsequent hydration of BC-BrC particles have been simulated.

• Restructuring of BC-BrC particles due to surface neutralization reactions causes reduction in light absorption at low relative humidity.

• Hygroscopic growth of processed BC-BrC particles enhances light absorption at high relative humidity.

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Modification in light absorption cross section of laboratorygenerated Black Carbon-Brown Carbon particles upon surface reaction and hydration

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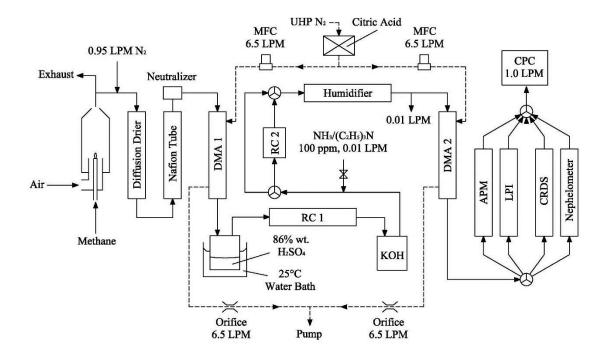
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1. Experimental setup

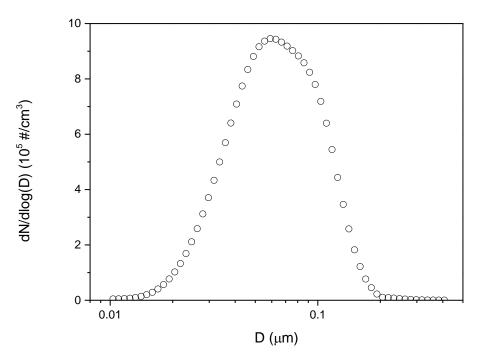
Supplementary Scheme. Schematic layout of BC-BrC generator and the tandem DMA setup in coupled with aerosol particle mass (APM) analyzer, low pressure impactor (LPI), cavity ring-down spectrometer (CRDS), and nephelometer. The two DMAs were performed by a "open loop" setting, whose inlets and outlets of sheath flow were controlled by mass flow controllers (MFC) and orifices respectively. Condensation particle counter (CPC) provided a constant sample flow rate of 1.0 LPM in all applications. Extended exposure of BC-BrC or PSL to sulfuric acid vapor was carried out in the first residence chamber (RC 1), and excess sulfuric acid was eliminated by a KOH trap. Ammonia (NH₃) or triethylamine ((C_2H_5)₃N) was introduced downstream to interact with H₂SO₄-coated particles. While a 10-cm-long and 6-mm-i.d. tubing worked for NH₃ experiments, a 100 ml residence chamber (RC 2) was used for (C_2H_5)₃N experiments.



2. BC-BrC particle characterization

Supplementary Figure S1. Size distribution of a polydisperse fresh BC-BrC population. It was characterized by a count median diameter of 69 nm and a geometric standard deviation of 1.24.

The fraction of doubly charged particles was estimated based on the particle size distribution and the bipolar charge distribution (Wiedensohler, 1988). The charge efficiency for singly charged 150 nm and doubly charged 225 nm particles is predicted to be 23% and 8%, respectively (Wiedensohler, 1988). On the other hand, the particle size distribution showed a very narrow peak shape with a median diameter at 69 nm and a geometric standard deviation of 1.24. We benefit from the narrow size distribution so that the ratio of number concentration of 150 nm and 225 nm particles was 23:1. Hence, doubly charged particle with a diameter of 225 nm contributed ~1% to singly charged particle with a diameter of 150 nm classified by DMA. This would lead to an overestimation of 7% and 3% for scattering and absorption coefficients for singly charged BC-BrC particles according to $\sigma_{scat} \propto D_p^5$ and $\sigma_{abs} \propto D_p^3$, respectively, assuming an identical refractive index for 150 nm and 225 nm BC-BrC.

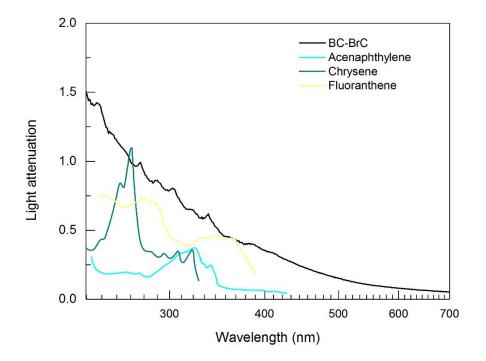


Supplementary Figure S2. UV-Vis spectra of BC-BrC in dichloromethane (CH_2Cl_2 , HPLC grade) emulsion. UV-Vis spectra of three polycyclic aromatic hydrocarbons (PAHs, acenaphthylene, chrysene and fluoranthene) detected by GC-MS are plotted as references (Linstrom and Mallard, 2014).

The BC-BrC particles were deposited on a glass slide placed above the chimney exhaust, and then were dissolved and/or dispersed by placing the glass slide into dichloromethane in an ultrasonic bath for 30 min. The light attenuation of the emulsion was measured by a photodiode array spectrophotometer (Scinco s-3150). An extinction Ångström exponent (EAE) of 2.9 was derived from the light attenuation in the wavelength range of 250-800 nm.

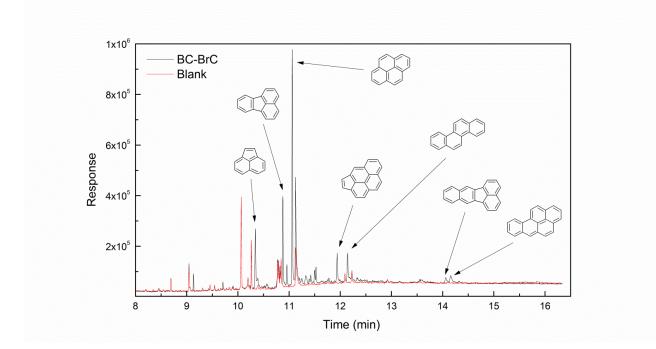
The obtained light attenuation should be assigned to light extinction because scattering of suspended BC in dichloromethane also contributed. Hence, the light attenuation should be considered as a sum of light absorption by dissolved OC (mostly PAHs up to 600 Da molecule weight) (Michela et al., 2008) and extinction by dispersed BC primary spheres.

A Mie theory calculation using a volume equivalent diameter of 100 nm and a complex refractive index of 1.22+0.31i yielded an extinction Ångström exponent of ~1.05. The complex refractive index was the refractive index of BC as 1.74+0.44i (Bond and Bergstrom, 2006) normalized by that of dichloromethane as 1.424+0.00i. If the BC particles break up into primary spheres with a smaller size (e.g., 20 nm), the EAE can be equal to unity. Therefore, we draw the conclusion that an EAE of ~2.9 suggests the presence of BrC, in line with the GC-MS analysis.



Supplementary Figure S3. GC-MS analysis of chemical composition of BC-BrC.

The emulsion of deposited BC-BrC was sonicated for 10 min and filtered, and CH_2Cl_2 -soluble organic components were analyzed using an Agilent 6890 series GC system with a HP-5 MS capillary column (30 m × 0.32 mm × 0.25 µm), coupled to a HP 5973 mass selective detector. The CH_2Cl_2 extracts indicates the presence of a series of PAHs including acenaphthylene, fluoranthene, pyrene, cyclopenta[cd]pyrene, chrysene, benzo[K]fluoranthene, and benzo[a]pyrene.



3. Scattering Ångström exponents

Scattering Ångström exponents of fresh and processed BC-BrC with an initial diameter of 150 nm ranged from 3.43 to 3.56. Since scattering Ångström exponents are largely determined by the particle size, no significant alteration was observed for fresh and processed BC-BrC upon surface reaction and hydration. In addition, inter-comparison between our measurement and theoretical calculation has been performed. For 150 nm PSL particles with a complex radiative index of 1.598 + 0.00i, the measured Ångstrom scattering coefficient in the 450-700 nm wavelength range by a nephelometer was 3.76, in good agreement with the theoretical prediction of 3.75

Supplementary Table S1. Scattering Ångström exponents of fresh and processed BC-BrC and authentic PSL.

5% RH	85% RH
3.56	
3.53	3.52
3.45	3.53
3.44	3.43
3.76	
	3.53 3.45 3.44

4. Uncertainty analysis

The CRDS system measured the extinction coefficient of aerosol particles with an uncertainty of 3% under laboratory experimental conditions (Li et al., 2011). Following the procedures recommended by Anderson and Ogren (1998), the precision uncertainty for scattering coefficient measurement using nephelometer (TSI, 3563) during our experiment was within 1%, while the accuracy was within 7%. Therefore, the uncertainty of absorption coefficient and single scattering albedo (SSA) was estimated to be 7%-9% and 8%, respectively, according to the following equations:

$$\frac{\Delta \sigma_{abs}}{\sigma_{abs}} = \sqrt{\left(\Delta \sigma_{ext}\right)^2 + \left(\Delta \sigma_{scat}\right)^2}$$
$$\frac{\Delta SSA}{SSA} = \sqrt{\left(\frac{\Delta \sigma_{ext}}{\sigma_{ext}}\right)^2 + \left(\frac{\Delta \sigma_{scat}}{\sigma_{scat}}\right)^2}$$

where $\Delta \sigma_{abs}$, $\Delta \sigma_{ext}$, $\Delta \sigma_{scat}$ and ΔSSA are one standard deviation of absorption coefficient, extinction coefficient, scattering coefficient and single scattering albedo, respectively. Note that the uncertainty of absorption/scattering/extinction cross sections partially originated from the fluctuation of particle concentration monitored by CPC.

$$\frac{\Delta C_{abs/scat/ext}}{C_{abs/scat/ext}} = \sqrt{\left(\frac{\Delta \sigma_{abs/scat/ext}}{\sigma_{abs/scat/ext}}\right)^2 + \left(\frac{\Delta N}{N}\right)^2}$$

where $\Delta C_{abs/scat/ext}$ is the uncertainty of absorption/scattering/extinction cross section and ΔN represents one standard deviation of the particle concentration.

5. R_{BC-BrC}

We define R_{BC-BrC} as the ratio between the mass of nonrefractory inorganic coatings (m_{NR-I}) and the mass of BC-BrC particle (m_{BC-BrC}) :

$$R_{BC-BrC} = \frac{m_{NR-I}}{m_{BC-BrC}} = \frac{\pi \rho_{eff} \times D_0^3 \times (G_f^3 - 1)}{6m_{BC-BrC}}$$

where ρ_{eff} is the effective density of the nonrefractory coating, D_0 is the initial particle diameter, G_f is the measured growth factor using TDMA.

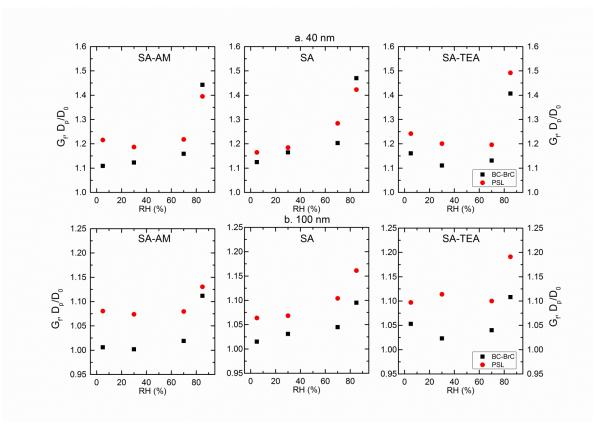
	RH	$ ho_{e\!f\!f}~({ m g/cm}^3)*$	m_{NR-I} (fg)	R_{BC-BrC}
BC-BrC-SA	5%	1.60	0.09	0.04
BC-BrC-SA-AM	5%	0.77	0.22	0.11
BC-BrC-SA-TEA	5%	1.23	0.42	0.22
BC-BrC-SA	85%	1.16	0.77	0.39
BC-BrC-SA-AM	85%	1.21	0.60	0.31
BC-BrC-SA-TEA	85%	1.12	0.84	0.43

Supplementary Table S2. Mass and densities of the nonrefractory coating, and R_{BC-BrC} .

* ρ_{eff} is determined by the physicochemical properties of nonrefractory coating dependent on their chemical composition and morphology (see Supplementary Tables 3a and 3b).

6. Hygroscopicity of processed particles

Supplementary Figure S4. Hygroscopic growth factors (G_f) for BC-BrC (square) and PSL (circle) with initial diameter of (a) 40 nm and (b) 100 nm for three different sulfate coatings: SA, SA-AM and SA-TEA.



7. Moles of sulfuric ions (SO_4^{2-}) and effective density ρ_{eff}

The moles of SO_4^{2-} at variable relative humidity have been derived from G_f for PSL-SA and PSL-SA-AM and physicochemical properties of corresponding compounds.

$$mole_{SO_4^{2-},RH} = \frac{\pi \rho_{RH} \times wt_{RH} \times D_0^3 \times (G_{f,RH}^3 - 1)}{6M_s}$$

where ρ_{RH} is the density of solution, wt_{RH} is the weight percentage of solute, and M_s is the molar mass of solute. ρ_{RH} and wt_{RH} are calculated from the method recommended by Biskos et al. (2009) and E-AIM (http://www.aim.env.uea.ac.uk, Wexler and Clegg (2002)).

The calculated results exhibited good agreement among PSL-SA at the RH range of 30-85% and PSL-SA-AM at 85% RH, which suggested no sulfuric acid loss or accumulation during the humidity elevation or neutralization processes (Table S3a). The derived moles of SO_4^{2-} for PSL-SA at 5% RH didn't agree well with the values at 30%-85% RH, which might be explained by the uncertainties in measuring G_f when coating thickness was close to the detection limit of TDMA. Calculated molecules of SO_4^{2-} in the PSL-SA-AM coating at 5%-70% RH varied in a narrow range, but were higher than those in the solution coating. It inferred an overestimated coating volume (thickness) derived from the G_f measurements, which can be rationalized by the formation of nonuniform solid crystalline of ammonium sulfate on the particle surface (Gibson et al., 2007). The effective densities (ρ_{eff}) of those AS coating below deliquescence relative humidity were calculated by the equation:

$$\rho_{eff,AS} = \frac{6M_{AS} \times \overline{mole_{SA}}}{\pi D_0^3 \left(G_{f,RH}^3 - 1\right)}$$

where $\overline{mole_{SA}}$ represents the average derived from data sets at 30%-85% RH with $\pm 1\sigma$ uncertainty.

The moles of SO_4^{2-} derived from PSL-SA-TEA as a triethylaminium sulfate (TEAS) solution at 85% RH was identical to $\overline{mole_{SA}}$. However, the calculated moles of SO_4^{2-} for PSL-SA-TEA at 5%, 30% and 70% RH were generally smaller when physicochemical properties of TEAS used, which indicated that the PSL-SA was partially neutralized leading to the formation of a mixture of triethylaminium bisulfate/sulfate (TEASBS/TEAS) (Supplementary Fig. S5). Because the ethyl group terminal restricted the hygroscopic growth until 85%, the physicochemical properties of TEABS or TEAS from solution (Clegg et al., 2013) are likely to be unsuitable at intermediate humidity environment in our study.

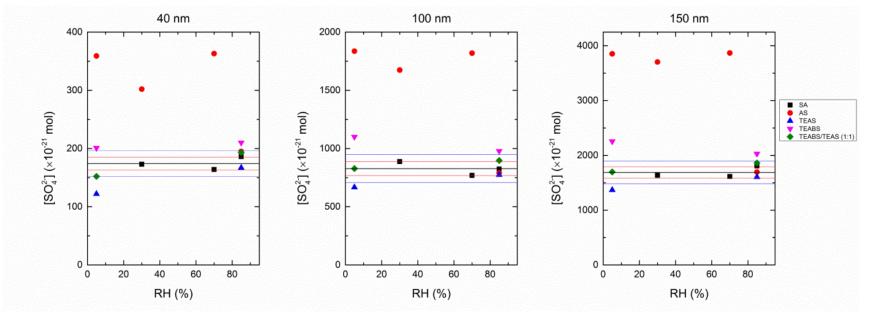
	RH	D_0	G_{f}	$ ho_{_{RH}}$	wt	$\left[SO_4^{2-}\right]$	$ ho_{_{e\!f\!f}}$
		nm		g/cm ³		$\times 10^{-21} \text{ mol}$	g/cm ³
		40	1.16	1.60	0.69	200	
	5%	100	1.06	1.60	0.69	1209	
		150	1.01	1.60	0.69	606	
		40	1.18	1.43	0.53	173	
	30%	100	1.07	1.42	0.53	888	
PSL-SA		150	1.04	1.42	0.53	1637	
PSL-SA		40	1.28	1.25	0.34	164	
	70%	100	1.10	1.24	0.34	769	
		150	1.07	1.24	0.33	1616	
		40	1.42	1.17	0.25	186	
	85%	100	1.16	1.16	0.23	824	
		150	1.11	1.16	0.23	1807	
	30%- 85%	40				174 ± 11	
$\overline{mole_{SA}}$		100				827 ± 60	
011		150				1687 ± 104	
		40	1.22	1.77	1.00	359	0.86
	5%	100	1.08	1.77	1.00	1836	0.80
		150	1.05	1.77	1.00	3853	0.77
		40	1.19	1.77	1.00	302	1.02
	30%	100	1.07	1.77	1.00	1674	0.87
PSL-SA-		150	1.05	1.77	1.00	3704	0.81
AM		40	1.22	1.77	1.00	363	0.85
	70%	100	1.08	1.77	1.00	1819	0.80
		150	1.05	1.77	1.00	3869	0.77
	85%	40	1.40	1.21	0.37	195	
		100	1.13	1.21	0.37	790	
		150	1.09	1.21	0.37	1696	

Supplementary Table S3a. Moles of SO_4^{2-} calculated for PSL-SA and PSL-SA-AM.

	RH	D_0	G_{f}	$ ho_{\scriptscriptstyle RH}$	wt	$\left[SO_4^{2-}\right]$	SA:TEA
		nm		g/cm ³		$\times 10^{-21} \text{ mol}$	assumed
		40	1.24	1.19	1.00	122]
	5%	100	1.10	1.19	1.00	668	1:2
		150	1.06	1.19	1.00	1368	J
		40	1.24	1.23	0.99	152	J
	5%	100	1.10	1.23	0.99	828	1:1.5
		150	1.06	1.23	0.99	1697	J
		40	1.20	1.18	0.92	90]
PSL-SA- TEA	30%	100	1.11	1.18	0.92	733	1:2
IEA		150	1.08	1.18	0.92	1725	J
		40	1.20	1.15	0.72	68]
	70%	100	1.10	1.15	0.72	556	1:2
		150	1.07	1.15	0.72	1308	J
		40	1.49	1.12	0.57	167]
	85%	100	1.19	1.12	0.57	774	1:2
		150	1.13	1.12	0.57	1605	J
	85%	40	1.49	1.12	0.54	193]
		100	1.19	1.12	0.54	897	1:1.5
		150	1.13	1.12	0.54	1861]

Supplementary Table S3b. Moles of SO_4^{2-} calculated for PSL-SA and PSL-SA-TEA.

Supplementary Figure S5. Moles of $SO_4^{2^-}$ calculated for PSL-SA, PSL-SA-AM, and PSL-SA-TEA, as listed in Supplementary Tables S3a and S3b. The black line represents mean values for moles of $SO_4^{2^-}$ derived from hygroscopic growth of PSL-SA between 30%-85% RH with ±1 σ (red line) and ±2 σ (blue line) uncertainty. For PSL-SA and PSL-SA-AM particles, the $[SO_4^{2^-}]$ for SA (square) and AS (circle) coating, respectively, is plotted. For PSL-SA-TEA particles, the $[SO_4^{2^-}]$ with assumptions of various coating material, including TEAS (up triangle), TEABS (down triangle), and TEABS/TEAS (1:1, diamond), is plotted. It was indicated that TEA can only partially neutralize coated H₂SO₄, *i.e.*, TEABS/TEAS coating at 5% RH and TEAS solution at 85% RH.



8. Charring of OC in the TDMA-TD system

There is a concern that heating of BC-BrC can lead to charring of OC. BrC (mainly PAHs) can be the main OC for carbonaceous particles produced in fuel-rich (oxygen-deficient) flames (Slowik et al., 2004). Hence, charring would have occurred, if ever, during the particle generation where temperature was higher than 250 °C, which is not evidenced in our morphology analysis.

Charring of OCs from particulate matters collected on the quartz filter has been well studied, and it mainly happens when the temperature is raised to >250 °C (Chow et al., 2004). Kondo et al. (2011) reported that interferences in absorption measurements due to charring after heating particles to 300 °C or 400 °C were 1.7% and 4.6% in maximum and 0.4% and 1.8% on average, respectively. Also, constant relative intensities of C1⁺ to C3⁺ ions (which dominant the BC signal) were measured by AMS in CalNex and CARES studies using a thermal denuder at 245 °C, indicating that charring did not occur (Cappa et al., 2012).

On the other hand, refractory sulfur-containing species was formed during the condensation of H_2SO_4 as evidenced in our EDX results. However, the amount of "refractory sulfur-containing species" was minor.

Hence, given the heating temperature of 250 °C, we don't expect any significant charring of our BC-BrC particles.

9. Mie theory calculation with a core-shell configuration

The estimation of absorption and scattering cross section were performed by a MATLAB functions based on Mie theory (Mätzler, 2002).

The volume equivalent diameter was derived according to the equation:

$$\frac{\pi}{6}d_{ve}{}^{3}\rho_{p} = \frac{\pi}{6}d_{m}{}^{3}\rho_{eff}$$

where d_{ve} is the volume equivalent diameter, d_m is the mobility diameter, ρ_p is the particle density, and ρ_{eff} is the effective particle density (DeCarlo et al., 2004). The particle density was derived from the mass fraction (34% and 66%) and density (1.3 g/cm³ and 1.8 g/cm³) of OC and EC, respectively. The volume equivalent diameter was then determined to be 132.7 nm for BC-BrC particles with a mobility diameter of 150 nm. The shell thickness was defined as the difference between the volume equivalent diameters of fresh and processed BC-BrC particles.

The complex refractive indices of fresh BC-BrC particles were retrieved to be 2.00+0.20i from Mie "closure" calculations based on particle size, scattering, and absorption measurements. The shell refractive indices of pure sulfuric acid, ammonium sulfate and triethylaminium sulfate are 1.42, 1.50 and 1.42, respectively (Lavi et al., 2013). Under high humidity, the shell refractive indices of the corresponding solution are calculated according to the volume weighted mixing rule (Michel Flores et al., 2012). The water refractive index is 1.335, and the volume of solvent and water are derived using the E-AIM model. The absorption cross section under variable relative humidity conditions for different forms of coating was ready to be obtained using a coreshell configuration. Note that uniform coating of sulfate ammonium was assumed here, which would lead to overestimation of absorption below deliquescent relative humidity. E_{abs} was derived from the measured absorption cross section of fresh BC-BrC and the calculated one of processed BC-BrC according to:

$$E_{abs} = \frac{C_{abs, \text{processed BC-BrC}}}{C_{abs, \text{fresh BC-BrC}}}$$

To investigate the effect of R_{BC-BrC} over a large span, i.e. variable mass of coating, the model was performed in the range of $\left\lceil SO_4^{2-} \right\rceil$ 10-fold lower or higher than that in this study.

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