

TReacLab: An object-oriented implementation of non-intrusive splitting methods to couple independent transport and geochemical software

Daniel Jara, Jean-Raynald de Dreuzy, Benoit Cochepin

▶ To cite this version:

Daniel Jara, Jean-Raynald de Dreuzy, Benoit Cochepin. TReacLab: An object-oriented implementation of non-intrusive splitting methods to couple independent transport and geochemical software. Computers & Geosciences, 2017, 109, pp.281-294. 10.1016/j.cageo.2017.09.005. insu-01588325

HAL Id: insu-01588325 https://insu.hal.science/insu-01588325

Submitted on 15 Sep 2017

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

Accepted Manuscript

TReacLab: An object-oriented implementation of non-intrusive splitting methods to couple independent transport and geochemical software

Daniel Jara, Jean-Raynald de Dreuzy, Benoit Cochepin

PII: S0098-3004(17)30251-0

DOI: 10.1016/j.cageo.2017.09.005

Reference: CAGEO 4019

To appear in: Computers and Geosciences

Received Date: 6 March 2017

Revised Date: 1 September 2017 Accepted Date: 7 September 2017

Please cite this article as: Jara, D., Dreuzy, J.-R.d., Cochepin, B., TReacLab: An object-oriented implementation of non-intrusive splitting methods to couple independent transport and geochemical software, *Computers and Geosciences* (2017), doi: 10.1016/j.cageo.2017.09.005.

This is a PDF file of an unedited manuscript that has been accepted for publication. As a service to our customers we are providing this early version of the manuscript. The manuscript will undergo copyediting, typesetting, and review of the resulting proof before it is published in its final form. Please note that during the production process errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.



- 1 TReacLab: an object-oriented implementation of non-intrusive
- 2 splitting methods to couple independent transport and geochemical
- 3 software
- 4 Daniel Jara¹, Jean-Raynald de Dreuzy¹, Benoit Cochepin²
- ¹Géosciences Rennes, UMR CNRS 6118, Campus de Beaulieu, University of Rennes 1,
- 6 Rennes, France
- ²ANDRA, 1/7 Rue Jean Monnet, 92298 Châtenay-Malabry, France

8 Abstract

Reactive transport modeling contributes to understand geophysical and geochemical processes in subsurface environments. Operator splitting methods have been proposed as non-intrusive coupling techniques that optimize the use of existing chemistry and transport codes. In this spirit, we propose a coupler relying on external geochemical and transport codes with appropriate operator segmentation that enables possible developments of additional splitting methods. We provide an object-oriented implementation in TReacLab developed in the MATLAB environment in a free open source frame with an accessible repository. TReacLab contains classical coupling methods, template interfaces and calling functions for two classical transport and reactive software (PHREEQC and COMSOL). It is tested on four classical benchmarks with homogeneous and heterogeneous reactions at equilibrium or kinetically-controlled. We show that full decoupling to the implementation level has a cost in terms of accuracy compared to more integrated and optimized codes. Use of non-intrusive implementations like TReacLab are still justified for coupling independent transport and chemical software at a minimal development effort but should be systematically and carefully assessed.

- 24 **Keywords:** Porous splitting; Object-oriented media; Reactive transport; Operator
- 25 programming.
- 26 Corresponding author: daniel.jara.heredia@gmail.com

27 1. Introduction

31

41

47

The fate of chemical species in geological media results from the interaction of physical 28 transport and chemical reactivity (Steefel et al., 2005). Understanding how they interact 29 requires field and laboratory studies as well as numerical models. Numerical models are 30 important for building predictive scenarios where experiments are limited spatially and temporally, as in long-term nuclear waste disposal assessment (Marty et al., 2014; Thouvenot 32 et al., 2013; Trotignon et al., 2007). On the physical transport side, extensive work in applied 33 34 mathematics and computational science has provided widely-used software for single and 35 multi-phase flows as well as transport of chemical species such as MODFLOW (McDonald and Harbaugh, 1988), MT3DMS (Zheng and Wang, 1999), HYDRUS (Kool and Van 36 Genuchten, 1991), COMSOL (COMSOL, 2010), FEFLOW (Diersch, 1996), MRST (Lie, 37 2014), and TOUGH2 (Pruess et al., 1999). On the chemistry side, geochemical software have 38 39 implemented a wide range of chemical functions and reactions, including equilibrium aqueous 40 speciation, equilibrium mineral dissolution/precipitation, gas phase exchange, ion exchange, redox reactions, and kinetic reactions. Some of these software are PHREEQC (Parkhurst and Appelo, 1999), GEMS (Kulik et al., 2013), CHEPROO (Bea et al., 2009), MINTEQ 42 (Peterson et al., 1987), CHESS (Van der Lee, 2002), and Geochemist's Workbench (Bethke, 43 44 2007). To combine physical and chemical reactivity, couplers have been developed between 45 transport and geochemical codes such as PHAST for coupling HST3D and PHREEQC 46 (Parkhurst et al., 2004), HP1 for HYDRUS and PHREEQC (Šimůnek et al., 2006), PHT3D

| 48 | for MT3DMS and PHREEQC (Prommer et al., 1999), HYTEC for RT1D/R2D2/METIS and |
|----|---|
| 49 | CHESS (van der Lee et al., 2003), OpenGeoSys-GEMS (Kulik et al., 2013) and iCP for |
| 50 | COMSOL and PHREEQC (Nardi et al., 2014), UTCHEM-IPhreeqc and |
| 51 | UTCHEM-EQBATCH (Kazemi Nia Korrani et al., 2015, 2016), multicomponent transport |
| 52 | software-IPhreeqc (Muniruzzaman and Rolle, 2016), FEFLOW-IPhreeqc (MIKE(DHI), |
| 53 | 2016), Lattice Boltzmann transport software-IPhreeqc (Patel et al., 2013). Most of the |
| 54 | previously cited codes have embedded the coupling method with the geochemical and |
| 55 | transport methods to enhance global performance and reliability. Here, in order to gain |
| 56 | flexibility, we propose in our code TReacLab a complementary development in the form of an |
| 57 | ensemble of Operator Splitting methods (OS) with a generic set of interfaces to transport and |
| 58 | reaction operators. In this context, OS decouples chemistry from transport as opposed to |
| 59 | global implicit solvers, which have been proven to be more accurate but less flexible |
| 60 | (Hammond et al., 2012; Hammond et al., 2014; Mayer, 2000; Steefel, 2009; Zhang, 2012). |
| 61 | TReacLab is designed as an open toolbox where additional OS techniques can be |
| 62 | implemented and benchmarked. Other transport and geochemical codes may also be used at |
| 63 | the minimal cost of developing the necessary interfaces. TReacLab is written in MATLAB |
| 64 | based on a series of abstract classes using object-oriented programming (Commend and |
| 65 | Zimmermann, 2001; Register, 2007; Rouson et al., 2011). |
| 66 | After recalling in section 2 the reactive transport and OS formalism used, we present in |
| 67 | section 3 our OS implementation. We especially show how to implement alternative OS |
| 68 | methods and how to connect other transport and geochemical codes. Methods are assessed |
| 69 | and discussed on the basis of 3 benchmarks in section 4. |

70 **2. Numerical model**

71 **2.1. Reactive transport equation**

72 The reactive transport equation can be written in a general way as (Saaltink *et al.*, 1998):

$$\frac{\partial \theta c}{\partial t} = ML(c) + \theta S {}_{\theta}^{t} r_{\theta} + \theta S {}_{k}^{t} r_{k} + Q, \tag{1}$$

where c is the vector of concentrations for N_s chemical species in the system. θ is a diagonal 73 matrix containing the porosity or volumetric content of the phase. M is a diagonal matrix that 74 75 specifies whether a species is mobile or immobile. Its diagonal elements are 1 or 0 accordingly. S_k^t and S_e^t are the transposed stoichiometric matrix for kinetic and equilibrium 76 reactions, respectively. r_e and r_k (ML⁻³T⁻¹) are the reaction rates of the N_e equilibrium and N_k 77 kinetic reactions, respectively. Q is the external sink/source term (ML⁻³T⁻¹). L is the transport 78 operator (ML⁻³T⁻¹), which includes advection and diffusion. In the following, we consider 79 80 only single-phase flow:

$$L(c) = \nabla \cdot [\mathbf{D}\nabla c - \theta vc]. \tag{2}$$

D (L²T⁻¹) is the effective dispersion-diffusion tensor (Bear, 1972). The velocity \boldsymbol{v} (LT⁻¹) is computed in a pre-processing phase, which can be decoupled from the reactive transport problem as long as hydraulic properties are not modified by the chemical reactivity. The chemical system can be generically written as the combination of the N_e equilibrium reactions:

$$\emptyset_{\mathfrak{g}}(c) = 0,$$
 (3)

85 and of the N_k kinetically-controlled reactions:

$$r_k = \emptyset_k(c). \tag{4}$$

- 86 The reactive transport problem is thus made up of the N_s mass balance equation (1) and of the $N_e + N_k$ equilibrium and kinetic equations (3) and (4). Its unknowns are the concentrations c87 and the reaction rates r_e and r_k . The chemical equilibrium system (3) is composed of the 88 89 conservation equation and of the mass action law, relating reactants and products (Apoung-Kamga et al., 2009; Molins et al., 2004):
- 90

$$S_{\sigma}\log(c) = \log(K),$$
 (5)

- where K is the vector of equilibrium constants. 91
- 92 Components u are generally introduced when considering equilibrium reactions (Saaltink et 93 al., 2011):

$$\mathbf{u} = \mathbf{U}\mathbf{c},\tag{6}$$

94 where U is the component matrix (Fang et al., 2003; Friedly and Rubin, 1992; Hoffmann et al., 2012; Kräutle and Knabner, 2005; Steefel et al., 2005). They are N_s - N_e linear 95 96 combinations of chemical species that are not modified by equilibrium reactions (Molins et al., 2004; Morel and Hering, 1993): 97

$$US_{\varepsilon}^{t}r_{\varepsilon}=0. \tag{7}$$

- The component matrix is not unique. However, its application to equation (1) always leads to
- 99 a reduced system without the equilibrium rates but with the components u (Molins $et \ al.$)
- 100 2004; Saaltink et al., 1998):

107

108

$$\frac{\partial u}{\partial t} = UML(c) + U\theta S {}_{k}^{t} r_{k} + UQ. \tag{8}$$

- The reactive transport problem is then made up of the $2N_s$ N_e + N_k equations (3-6) and (8)
- for the same number of unknowns u, c and r_k .
- 103 Under the assumption that solid species are not transported and all species have the same
- diffusion coefficient (i.e. UML(c) = UL(u)). Equation (8) classically gives the two following
- 105 formulations TC and CC (Amir and Kern, 2010):

TC:
$$\frac{\partial u}{\partial t} = L(u_a) + U\theta S_k^t r_k + UQ. \tag{9}$$

CC:
$$\frac{\partial u_a}{\partial t} + \frac{\partial u_f}{\partial t} = L(u_a) + U\theta S \frac{t}{k} r_k + UQ. \tag{10}$$

where $u_a = UMc$ and $u_f = U(I - M)c$ are the aqueous and fixed components. In the TC

formulation, the fixed species concentration are deducted from the solution in the total

component concentration (T) and the solute concentration (C). In the CC formulation, the

total component concentration is divided in aqueous and fixed components.

2.2. Usual first-order sequential non-iterative and iterative approaches

In this section, we show how the reactive transport problem can be solved using independent transport and chemical solvers. We distinguish the sequential non-iterative and iterative approaches respectively based on TC and CC formulations. For the sequential non-iterative approach, we extract from the TC formulation, the transport operator in which we keep the sink/source term:-

$$\frac{\partial u}{\partial t} = L(u_a) + UQ. \tag{11}$$

The chemical operator derives from equations (3-6), and (8). Note that it does not contain any source/sink term, as it has been included in the transport equation:

$$\frac{\partial u}{\partial t} = U S_k^{t} r_k$$

$$r_k = \emptyset_k(c) \tag{12}$$

$$\emptyset_{\varepsilon}(c) = 0$$

This is still a system of $2N_s - N_e + N_k$ equations for the same number of unknowns. This decoupled system can be solved with the classical sequential non-iterative approach using an explicit integration of temporal derivatives (herein, we assume forward Euler). The solution at time step n+1 can be obtained from the solution at time step n, with the following successive application of the transport and chemical operators in a sequential approach:

$$u^* = u_n + \Delta t (L(u_n) + UQ)$$

$$\begin{cases} u^* = Uc_{n+1} \\ \emptyset_e(c_{n+1}) = 0 \end{cases}$$

$$r_{k_{n+1}} = \emptyset_k(c_{n+1})$$
(13)

 $u_{n+1} = u^* + \Delta t U S_k^{\tau} r_{k_{n+1}}$

The transport operator (11) is applied to the components. Then the chemical operator is applied with the updated mobile components for speciation between fixed and solute concentrations. In the specific case where chemical reactions are all at equilibrium and no kinetics is involved, a TC formulation is used to fully decouple (de Dieuleveult *et al.*, 2009). In such case the decoupling does not then rely on operator splitting, but on a block Gauss-Seidel method. When the stability conditions of the explicit integration are too much constraining, implicit schemes should be used instead within a sequential iterative approach (Carrayrou *et al.*, 2004; de Dieuleveult and Erhel, 2010; Yeh and Tripathi, 1989):

$$u_{n+1} = u_n + \Delta t \left[L(u_{\alpha_{n+1}}) + \theta S_k^t r_{k_{n+1}} + UQ \right]$$

$$\begin{cases} u_{n+1} = U c_{n+1} \\ \emptyset_s(c_{n+1}) = 0 \end{cases}$$

$$r_{k_{n+1}} = \emptyset_k(c_{n+1}).$$
(14)

131 Classical Picard's method have been extensively used to solve such kind of problems:

$$u_{n+1}^{k+1} = u_n + \Delta t \left[L \left(u_{a_{n+1}}^{k+1} \right) + \theta S_k^t r_{k_{n+1}}^k + UQ \right]$$

$$\begin{cases} u_{n+1}^{k+1} = Uc_{n+1}^{k+1} \\ \emptyset_s(c_{n+1}^{k+1}) = 0 \end{cases}$$
 (15)

$$r_{k_{n+1}}^{k+1} = \emptyset_k(c_{n+1}^{k+1}),$$

where k is the index of the Picard iteration method instantiated by:

$$u_{n+1}^{k=1} = u_n$$

$$r_{k_{n+1}}^{k=1} = r_{k_n}.$$
(16)

We recall the necessity to check the consistency of the temporal integration scheme with the Operator Splitting method chosen. With this decomposition, explicit first-order scheme naturally leads to sequential non-iterative approach. The implicit first-order scheme requires a sequential iterative approach. Other choices are possible and might reduce errors depending on the chemical system (Barry *et al.*, 1996). As it should be possible to test and benchmark them at a reduced development cost, we use a generic decoupling formalism that can be used to implement a broad range of schemes.

2.3. Generic operator splitting implementation

133

134

135

136

137

138

139

140

- 141 The reactive transport system can be generically split in two operators. Using the formalism
- of Gasda *et al.* (2011), equation (1) can be written as:

$$\frac{\partial Z}{\partial z} = \mathcal{L}_1 Z + \mathcal{L}_2 Z, \qquad Z(t=0) = Z_0 \qquad 0 \le t \le T, \tag{17}$$

where Z is the unknown, L_1 and L_2 can be equation (11) and (12), respectively. Other 144 decomposition are possible, e.g. the transport operator can be subdivided into an advection

143

145

146

147

148

149

150

151

152

153

154

155

156

157

158

159

160

161

162

and a diffusion-dispersion operator (Clement et al., 1998), or one operator might contain

advection-reaction and the other diffusion (Liu and Ewing, 2005). Each operator will be

solved separately for a splitting time step $\Delta t = t^{n+1} - t^n$ using adapted numerical methods.

The generic operator splitting methods implemented into the Toolbox are the sequential splitting, additive splitting, Strang splitting, symetrically weighted splitting, and alternating method (Appendix A). Assuming exact integration of the operators and homogeneous boundary conditions in equation (18), the first two have a first-order temporal truncation error, and the following three a second-order one (Hundsdorfer and Verwer, 2013). Since the operators are usually solved using numerical methods, the global order of such approaches might be modified because of the order of the numerical methods used for each operator (Barry et al., 1996; Csomós and Faragó, 2008). The alternating splitting increases the order of the sequential splitting if the time steps are small enough (Simpson and Landman, 2008; Valocchi and Malmstead, 1992).

3. Operator splitting implementation and software organization

We provide in TReacLab an object-oriented toolbox for the non-intrusive operator splitting methods of the previous section. TReacLab is organized along three main components for coupling transport and reactivity, and proceeds in three pre-processing, processing and post-processing phases (Figure 1). These three components correspond to the three

well-identified coupler, transport and chemistry classes. The three classes are fully segmented and exchange information through interfaces. Segmentation ensures that any of the three coupler, transport and chemistry classes can be replaced without modifications of any of the two other ones. The solution of the reactive transport problem after spatial discretization eventually consists in the temporal integration with the chosen OS technique, which iteratively calls transport and geochemical solvers through interfaces (Figure 1, middle row). This is the core of the simulation that we identify as the processing phase. It is generic and does not require at run time any further specification of transport, reactivity and coupler methods. Standard error management techniques are used to stop the algorithm when any of the integration method of the three classes fails, stopping the running process and returning adapted error messages.

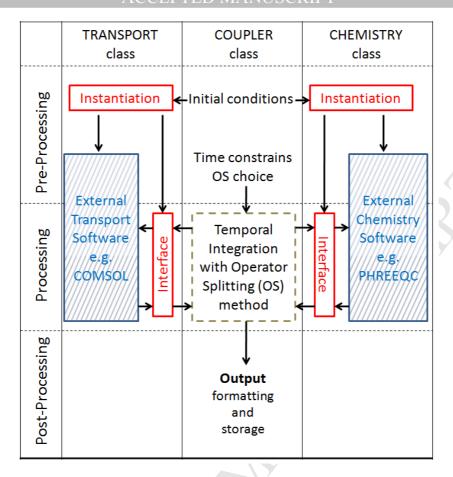


Figure 1: General software organization of TReacLab with the three coupler, transport and chemistry classes in columns, and the three pre-processing, processing, and post-processing phases in rows. Generic components represented in black are the organization and the coupler class. External software for transport and chemistry are represented in blue with hatched line (cannot be modified). Red boxes highlight the instantiation and interface methods that must be developed when connecting new transport or chemistry software.

The processing phase can be generic because all specifications of the coupler, transport, and chemistry classes are performed in a pre-processing phase (Figure 1, first row). The pre-processing phase consists in the instantiation of the coupler, transport and chemical classes, in the preparation of the interfaces that will transfer information and in the specifications of the initial conditions. As detailed in Appendix B, instantiations are code dependent. Instantiation can be done externally for example with the definition of a transport

187 or chemical problem through the graphical user interface of software like COMSOL or 188 PHREEQC. It can also be done internally by a method within TReacLab specifying the inputs 189 and parameters to existing interfaces like IPhreeqc (Charlton and Parkhurst, 2011), 190 PhreegcRM (Parkhurst and Wissmeier, 2015), or COMSOL livelink (COMSOL, 2010). Even 191 when instantiation is complex, it remains independent for each of the three classes. 192 Cross-dependencies and feedback between transport and reactivity like density-driven flows 193 with reacting species are not supported at this stage, although they may be important in some 194 applications like CO₂ sequestration (Abarca et al., 2013). 195 Pre-processing phase specifies the initial conditions and transfers them to the coupler in charge of starting the numerical integration. Post-processing is generic and only consists in 196 formatting and storing output concentrations and solver performances (Figure 1, bottom row). 197 198 Specifications are all restricted to the instantiation of the software and interface in the pre-processing phase while processing and post-processing remain fully generic. Connections 199 between specific algorithms and generic structures are done by interfaces. Appendix B 200 provides a detailed description of the transport and chemistry classes, defining the interfaces 201 202 to the external codes.

4. Examples and benchmarks

203

204

205

206

207

208

209

The three following examples validate the methods and illustrate the implementation presented in sections 2 and 3. The three of them are based on a 1D hydraulically homogeneous system with steady-state flow and uniform dispersion (equation (2)). The examples are compared visually against analytical solution or well-know numerical software. Moreover, we show a convergence study for the first case being the reference solution the numerical solution with finest time resolution.

210

211

212

213

214

215

216

217

218

219

220

221

222

223

224

225

226

227

The four examples display evolving degrees of complexity both in terms of chemical systems and in terms of software called for transport and reactivity, software versions are given in Table 1. The first example is a single-species transport with first-order decay. The transport solver is COMSOL and the chemical solver is a simple analytical solution. This example is used to assess the different coupling algorithms implemented and to check the implementation the interface with COMSOL. The second example equilibrium precipitation/dissolution chemical system in a 1D hydraulically homogeneous system. Chemical solver is IPhreegc. Several solvers have been compared for the transport solver, both to check IPhreegc interface implementation and to evaluate the effect of the transport solver. The third example is the most advanced in terms of chemistry. Chemical reactions are partly in equilibrium and partly kinetically controlled. They involve precipitation and dissolution reactions. The chemical code is PhreegcRM. It is used in combination with COMSOL as transport solver. The last problem face a 2D unsaturated system where transport is modeled by Richards equation and solved by COMSOL. Chemistry is solved by PhreeqcRM. These four test cases have been chosen to check the implementation and assess the coupling methods developed. They are also simple enough from the development point of view to be taken as starting points to model more advanced chemical systems and transport conditions.

| Software | Version |
|-----------|---------|
| MATLAB | R2013b |
| COMSOL | 4.3b |
| PHREEQC | 3.3.7 |
| IPhreeqc | 3.3.7 |
| PhreeqcRM | 3.3.9 |

228 Table 1: Software versions.

229

235

236

237

238

239

240

241

242

243

4.1. Single-species transport with first-order decay

A single-species transport with first-order decay using different OS methods is compared to an analytical solution (Van Genuchten and Alves, 1982). The reactive transport system contains a single solute species of concentration c:

$$\frac{\partial c}{\partial c} = L(c) - kc, \tag{18}$$

where *L* is given by equation (2). Equation (18) can straightforwardly be separated into transport and chemistry operators corresponding to the two right-hand side terms.

At time 0, the solute concentration is 0 in the domain (c(x, t=0) = 0). The concentration at the left boundary is constant and equal to 1 mol/m³ $(c(x = 0, t) = 1 \text{ mol/m}^3)$. The boundary condition on the right side of the domain is a perfectly absorbing condition $(c(x = x_{\text{max}}, t) = 0)$.

Parameters are derived from Steefel and MacQuarrie (1996) and given in Table 2. The solver for transport is COMSOL and an analytical solution is used for the first-order decay. Solute concentration progressively invades the domain from the left boundary with a smooth profile resulting from the combination of dispersion and decay (Figure 2). Second-order methods perform much better than first-order methods as expected. Errors are more pronounced at the inlet boundary condition on the left side of the domain where the concentration is higher

(Steefel and MacQuarrie, 1996; Valocchi and Malmstead, 1992). The sequential splitting method with the transport operator performed first overestimates the amount of reaction for the whole domain since it considers that all incoming solute is getting in without decay for the full first time step. If the sequence of operators is exchanged, namely first chemistry is solved, and then transport is solved, the amount of reaction is underestimated. The second-order alternating splitting, which alternates between transport-chemistry and chemistry-transport steps, shows strong improvement with compensations between overestimation in the first application of the chemical operator and underestimation in the second application of the chemical operator (Simpson and Landman, 2008; Valocchi and Malmstead, 1992).

| Parameter | Value |
|----------------------|--------------------|
| v [m/y] | 100 |
| $D [m^2/y]$ | 20 |
| $k [y^{-1}]$ | 100 |
| x_{max} [m] | 6 |
| Δx [m] | 0.4 |
| Δt [y] | 4 10 ⁻³ |

Table 2: Parameters for the single-species transport with first-order decay benchmark. v is

the velocity, D is the dispersion coefficient, k is the decay rate, x_{max} is the length of the 1D column, Δx is the grid size, and Δt is the time step.

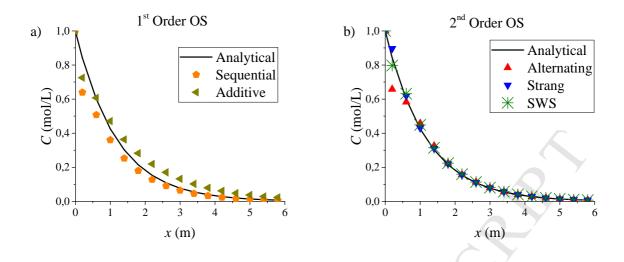


Figure 2: Comparison of first- and second-order OS for the single-species transport with

first-order decay at t = 0.5 y. Parameters are given in Table 2. Analytical solution is derived

from Van Genuchten and Alves (1982).

The error at time t = 0.5 y is taken as the quadratic relative difference over the domain of the finest time step of the numerical solution and the numerical solutions for the corresponding time step, c_{NF} and c_{N} respectively:

$$||e||_{2} = \sqrt{\sum_{i=1}^{N_{x}} \left(\frac{c_{NF}^{i}(t) - c_{N}^{i}(t)}{c_{NF}^{i}(t)}\right)^{2}}.$$
 (19)

Table 3 displays the values for evolving time steps and shows that all methods converge with the time. The reference finest time step for each method has been $\Delta t = 2 \ 10^{-4} \ \mathrm{s}$ (i.e. $c_{NF}^{t}(t)$ value). While all methods perform well, the sequential method is more accurate than the additive one and second-order methods are overall more accurate than first-order methods. The performance on convergence arranged on descending order is given by Strang, symmetrically weighted splitting, alternating, sequential and additive.

| | | 1 st 0 | order | | 2 nd order | |
|--------------|--------------------|-------------------|------------|----------------|-----------------------|-------|
| m | Δt (y) | Additive | Sequential | Alternating OS | Strang | SWS |
| = 0.4 m | 4 10 ⁻³ | 1,107 | 0,1667 | 0,075 | 0,032 | 0,049 |
| $\Delta x =$ | 2 10 ⁻³ | 0,514 | 0,079 | 0,032 | 0,026 | 0,028 |
| | 4 10 ⁻⁴ | 0,114 | 0,019 | 0,029 | 0,031 | 0,029 |

Table 3: Error $\|\mathbf{e}\|_2$ of equation (19) for the single-species transport with first-order decay

with different OS methods and splitting time steps.

4.2. Calcite dissolution

Calcite dissolution and dolomite formation has become a classical benchmark for reactive transport problems with sharp precipitation/dissolution fronts (Beyer *et al.*, 2012; Engesgaard and Kipp, 1992; Prommer *et al.*, 1999). Progressive introduction of magnesium calcium in a domain at equilibrium between calcium carbonate in solution and calcite (CaCO₃) dissolves the calcite and precipitates dolomite (CaMg(CO₃)₂). This chemical system has been modeled with the physical and chemical parameters given by Table 4, Table 5, and Table 6. Chemical concentrations are initially homogeneous. At the initial time (t = 0), the chemical system is destabilized with the introduction of magnesium instead of calcium at the upper boundary condition (x = 0), inducing the dissolution/precipitation process. The boundary condition at the downstream limit (x_{max}) is a simple outflow of the solutes.

Here, we show how transport solvers can be applied and validate our interface to IPhreeqc. IPhreeqc performs the computation of components, aqueous speciation, precipitation and dissolution reactions (Charlton and Parkhurst, 2011). The database used is 'NAPSI_290502(260802).dat'. Transport is solved either with COMSOL Multiphysics (COMSOL, 2012), with a finite difference spatial discretization and forward Euler time

integration, derived from built-in pdepe function of MATLAB (Skeel and Berzins, 1990).

Transport and chemistry are coupled through the simple sequential approach of equations

(A.1)-(A.3). PHREEQC is independently run as 1D reactive transport solver for general comparison.

| Parameter | Value |
|---------------------|----------------------|
| v [m/s] | 10 ⁻⁵ |
| $D [m^2/s]$ | 6.7 10 ⁻⁸ |
| $x_{\text{max}}[m]$ | 0.25 |
| Δx [m] | 0.01 |
| Δt [s] | 50 |

290 Table 4: Physical parameters for the calcite dissolution benchmark. ¥ is the average velocity,

291 *D* is the dispersion coefficient, x_{max} is the maximum length of the column, Δx is the grid size, 292 and Δt is the time step.

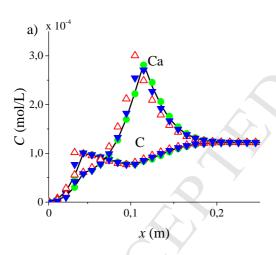
| Chemical Component and Species | Initial value | Boundary value at x=0 |
|--------------------------------------|-----------------------|-----------------------------|
| Ca [mol/L] | 1.23 10 ⁻⁴ | 0 |
| C [mol/L] | $1.23 \ 10^{-4}$ | 0 |
| Cl [mol/L] | 0 | 2 10 ⁻³ |
| Mg [mol/L] | 0 | 10^{-3} |
| pH [-] | 9.91 | 7 |
| Calcite [mol/L] | 2 10 ⁻⁴ | - |
| Dolomite [mol/L] | 0 | - |

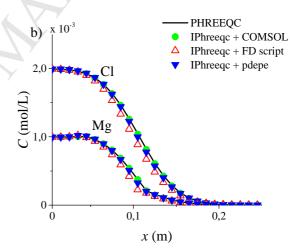
Table 5: Calcite dissolution benchmark initial and boundary values for aqueous components and mineral species. In PHREEQC, components are called elements.

| Homogeneous reactions | log (K) |
|--|----------|
| $2H^+ + 2e^- \leftrightarrow H_2$ | -3.1055 |
| $2H_2O - 4H^+ - 4e^- \leftrightarrow O_2$ | -85.9862 |
| $HCO_3^- + 9H^+ + 8e^ 3H_2O \leftrightarrow CH_4$ | 27.8493 |
| $H_2O - H^+ \leftrightarrow OH^-$ | -13.9995 |
| $H^+-H_2O+HCO_3^- \leftrightarrow CO_2$ | 6.3519 |
| $HCO_3^ H^+ \leftrightarrow CO_3^{2-}$ | -10.3289 |
| $C\alpha^{2+} - H^{+} + HCO_{3}^{-} \leftrightarrow C\alpha CO_{3}$ | -7.1048 |
| $Ca^{2+} + HCO_3^- \leftrightarrow CaHCO_3^+$ | 1.1057 |
| $Ca^{2+} + H_2O - H^+ \leftrightarrow CaOH^+$ | -12.78 |
| $Mg^{2+} - H^+ + HCO_3^- \leftrightarrow MgCO_3$ | -7.3492 |
| $Mg^{2+} + HCO_3^- \leftrightarrow MgHCO_3^+$ | 1.0682 |
| $Mg^{2+} + H_2O - H^+ \leftrightarrow MgOH^+$ | -11.44 |
| Homogeneous reactions | |
| Calcite | |
| $C\alpha CO_{3} \leftrightarrow C\alpha^{2+} - H^{+} + HCO_{3}^{-}$ | 1.849 |
| Dolomite | 4.118 |
| $CaMg(CO_3)_2 \leftrightarrow Ca^{2+} + Mg^{2+} - 2H^{+} + 2HCO_3^{-}$ | 4.110 |

Table 6: Chemical system of the calcite dissolution benchmark. The upper part comprises the homogeneous equations and the lower part the heterogeneous reactions. The first column shows the equilibrium reactions and the second one the logarithms of equilibrium constants.

Figure 3 and Figure 4 display aqueous and mineral equivalent concentrations at time $t = 10^4$ s. As magnesium and chloride get in the domain (Figure 3b and Figure 3d), calcite progressively dissolves and is replaced by dolomite as expected (Figure 4). Some of the calcium remains in solution and is flushed out (Figure 3a and Figure 3c). Because of the subsequent absence of calcium in solution, dolomite dissolves again with some increase of calcium in solution (Figure 3a and Figure 3c). The three different transport solvers give the same tendency as the PHREEQC solution.





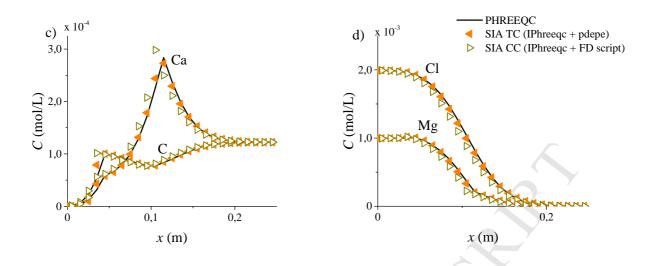


Figure 3: Aqueous concentration profiles at time $t = 10^4$ s.

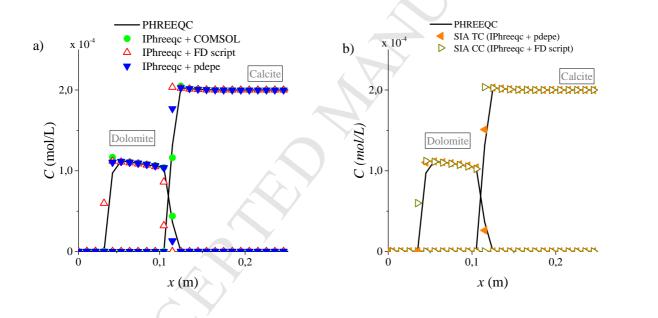


Figure 4: Dolomite and calcite equivalent concentration profiles with open and filled symbols respectively at time $t = 10^4$ s.

Although COMSOL leads to good results, it is more than one order of magnitude slower than the two other transport methods (Table 7). We checked that this large difference in performances does not come from the numerical method but from the large time required for COMSOL to start and stop when called numerous times externally. While this might not be

an issue for large transport problems for which limitations will rather come from transport operator, it is a constrain for smaller tests and benchmarks.

| Software Coupling | Time |
|----------------------|-------|
| IPhreeqc + COMSOL | 668 s |
| IPhreeqc + FD script | 24 s |
| IPhreeqc + pdepe | 40 s |

315 Table 7: Time performance for the calcite dissolution benchmark using a sequential operator splitting.

Whatever the coupling method, the consistency with PHREEQC is overall good. Although COMSOL uses, as default, implicit time integration schemes for solving the transport equation instead of the required explicit method, it still compares well with PHREEQC and the other software couplings. Indeed, the sequential non iterative method requires an explicit time integration for transport (equation (13)). It is not the case for COMSOL which uses (as default) a backward differentiation formula temporal integration scheme, which order varies with the internal time step adaptation (COMSOL, 2012). It thus introduces an additional error in the coupling scheme (de Dieuleveult *et al.*, 2009). However, by using such stable and accurate temporal integrations, it enhances the robustness of the transport scheme.

4.3. Mixed equilibrium-kinetic system

We simulate the progressive increase of dissolved species in an atmospheric water infiltrating a granitic bedrock. This test case is derived from Nardi *et al.* (2014). The hydraulic properties of the system are found in Table 8. The infiltrating water has much lower concentrations of dissolved species than the resident water. It interacts with five minerals (Table 9). It is in equilibrium with calcite. The four other minerals k-feldspar, illite, albite and pyrite are subject to kinetically controlled dissolution with rates ranging from 10^{-13} to 10^{-11} mol/s. All

parameters and rate laws of the simulation are provided in the PHREEQC file of iCP (Nardi *et al.*, 2014). The infiltrating water dissolves calcite to maintain equilibrium, increasing both the concentration of calcium and the pH of the solution. Other minerals also dissolve and increase the concentrations of Al and K in solution, however at a much slower rate because of the kinetic control of the reactions. pH is eventually buffered by the dissolution of illite and pyrite.

| Parameter | Value |
|---------------------|--|
| ₽ [m/s] | 2.78 10 ⁻⁶ m/s |
| D [m/s] | $5.55 \ 10^{-9} \ \text{m}^2/\text{s}$ |
| $x_{\text{max}}[m]$ | 0.08 m |
| Δx [m] | 10 ⁻³ m |
| Δt [s] | 720 s |

Table 8: Parameters for mixed equilibrium-kinetic benchmark. v is the average velocity, D is

the dispersion coefficient, x_{max} is the maximum length of the column, Δx is the grid size, and

 Δt is the splitting time.

| Chemical Component and Species | Initial value | Boundary value |
|--------------------------------------|----------------------|----------------------|
| Ca [mol/L] | $1.4\ 10^{-3}$ | 3 10 ⁻⁴ |
| C [mol/L] | $4.9\ 10^{-3}$ | 1.9 10 ⁻⁴ |
| Cl [mol/L] | 1.1 10 ⁻² | 9 10 ⁻⁴ |
| Mg [mol/L] | $7.4 10^{-4}$ | $2\ 10^{-4}$ |
| Mn [mol/L] | $3.4\ 10^{-6}$ | 0 |
| S [mol/L] | 9.6 10 ⁻⁴ | 4.8 10 ⁻⁴ |
| Na [mol/L] | $1.3 \ 10^{-2}$ | 3 10 ⁻⁴ |
| K [mol/L] | $2.5 \ 10^{-4}$ | 7.1 10 ⁻⁴ |
| Fe [mol/L] | $7.2 10^{-6}$ | 5.4 10 ⁻⁵ |
| Sr [mol/L] | 0 | 6.8 10 ⁻⁷ |
| Si [mol/L] | 2 10 ⁻⁴ | $2.5 \ 10^{-6}$ |
| Al [mol/L] | 5.1 10 ⁻⁹ | 10 ⁻⁸ |
| P [mol/L] | $3.8 \ 10^{-6}$ | 0 |
| Br [mol/L] | 1.7 10 ⁻⁵ | 0 |
| F [mol/L] | 3.1 10 ⁻⁵ | 1.6 10 ⁻⁵ |
| pH [-] | 7.5144 | 7.3 |
| pe [-] | -3.0836 | 13.6 |
| Calcite [mol/L] | 6.065 | - |
| K-feldspar [mol/L] | 0.239 | - |
| Illite [mol/L] | 0.144 | - |
| Albite [mol/L] | 0.289 | - |
| Pyrite [mol/L] | 1.17 | - |

343 Table 9: Aqueous components and mineral species for mixed equilibrium-kinetic benchmark.

To simulate this set of reactions, we have chosen PhreeqcRM to assess the flexibility of TReaCLab. Transport is simulated with COMSOL to benefit from the accurate transport solver, it uses a variable order (between 1 and 5) backward differentiation formula. In the

| presence of both kinetically controlled and equilibrium reactions, both the quality of the |
|--|
| transport and reactive integrations and coupling issues may be critical. We choose a simple |
| sequential OS method with the successive integration of transport and reactivity. The results |
| obtained by the coupling of COMSOL and PhreeqcRM are close to the solution given by |
| PHREEQC alone for the dissolved species and kinetically dissolving minerals (Figure 5). The |
| time step of the coupled PhreeqcRM and COMSOL integration has been taken smaller than |
| the characteristic mesh scale transport time and reactive time at least for the kinetical reaction |
| to ensure accurate integrations. The most difficult quantity to get accurately is the calcium |
| concentration because calcite is at equilibrium. The time step must be reduced to recover a |
| steeper reactive front (Figure 6). |
| This more advanced test shows that the computational load should be well balanced between |
| the coupler, transport and chemistry methods. While coupling is the critical component in |
| cases of equilibrium reactions and may even require highly integrated coupling strategies like |
| global implicit methods (Hoffmann et al., 2010; Saaltink et al., 2001), it is not the case for |
| kinetically controlled reactions. In this case of mixed equilibrium kinetic reaction, elementary |
| coupling and accurate transport and reactive solvers can be efficient with small enough time |
| stens where sharp reaction fronts are involved |

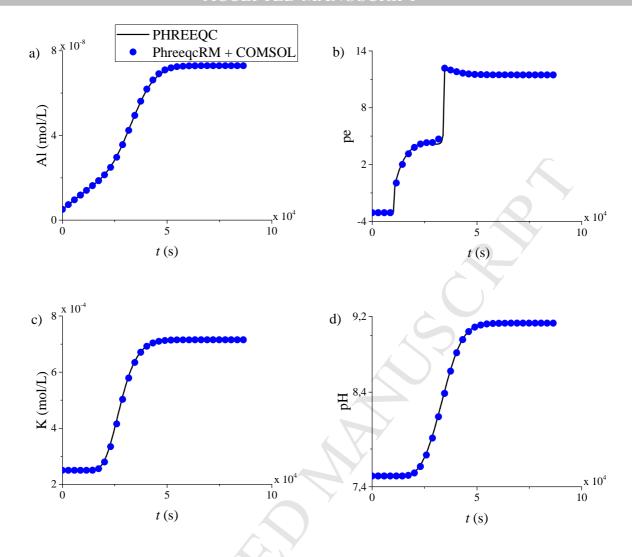


Figure 5: Comparison of results between the coupling of PhreeqcRM and COMSOL and PHREEQC observed for the mixed equilibrium-kinetic benchmark at the output of the column.

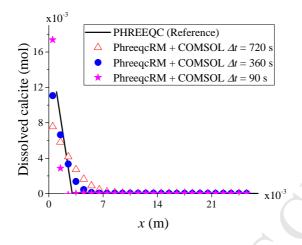
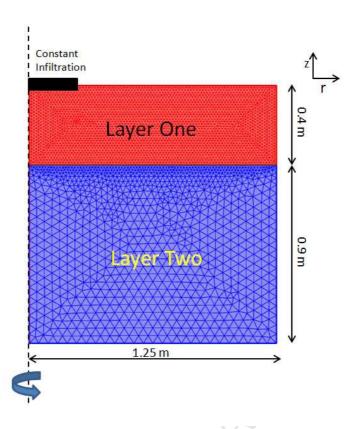


Figure 6: Quantity of dissolved calcite with PhreeqcRM and COMSOL for two different splitting time steps $\Delta t = 720$ s, 360 s and 90 s. PHREEQC independently is used as reference.

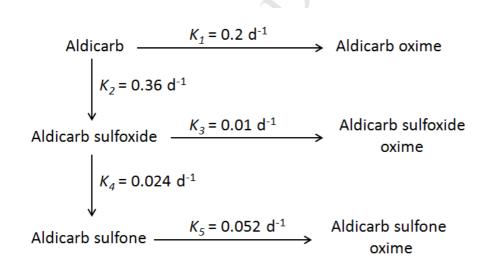
4.4. Pesticide infiltration

The following benchmark concerns the infiltration in an unsaturated soil column of a carbamate insecticide (Aldicarb) (MIKE(DHI), 2016; Multiphysics, 2008; Šimůnek $et\ al.$, 1994; Wissmeier and Barry, 2011). The soil column is a 2D axisymmetric cylinder made up of two layers with a smaller hydraulic conductivity in the upper layer but higher saturation. Transport is modeled by Richards' equation and solved by COMSOL (Figure 7). Aldicarb is transported downwards and sideways from the infiltration (top of the column from r=0 m to r=0.25 m). Chemistry is described by first-order decay chain reactions (Figure 8), being only mobile Aldicarb, Aldicarb sulfoxide and Aldicarb sulfone (i.e. the other species are fix species). These system of ordinary differential equations is solved by PhreeqcRM.



380

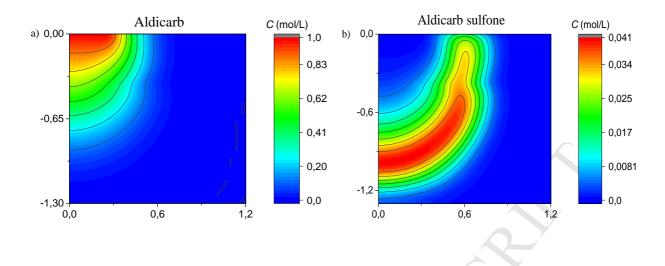
381 Figure 7: Soil column geometry and mesh.

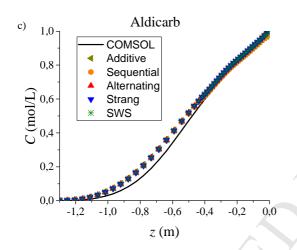


382

383 Figure 8:Aldicarb reaction chain.

The simulation time is 8 days with a splitting time step of 0.05 days. The number of nodes is 3936 nodes. Figure 9a and Figure 9b show the concentration in the soil column of Aldicarb and Aldicarb sulfone, respectively. Aldicarb disappears fast from the domain since its kinetic constant are fast in comparison to the kinetic constants of the daughter species. Therefore, Aldicarb (and also Aldicarb oxime) are presented close to the infiltration condition. On the contrary, the other daughter species (Aldicarb sulfoxide, aldicarb sulfone, aldicarb sulfoxide oxime, aldicarb sulfone oxime) have a similar distribution in the domain. Figure 9c and 9d show the concentration of Aldicarb and Aldicarb sulfone when r=0 m for the different OS methods and COMSOL alone. It is possible to see a good agreement between all the methods, although a discrepancy between the methods and COMSOL is observable. The discrepancy is related to the OS error and the chosen integration time scheme for the chemistry step.





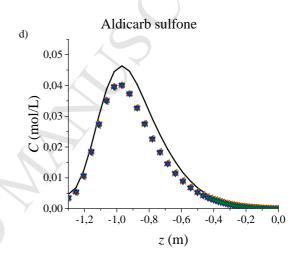


Figure 9: a) Aldicarb contour plot after 8 days, b) Aldicarb oxime contour plot after 8 days, c) Concentration aldicarb at r = 0 m for all the methods and Comsol, d) Concentration aldicarb oxime at r = 0 m for all the methods and Comsol.

5. Discussion

As shown by many previous studies and by the four examples of the previous section, reactive transport problems can be solved by a wide diversity of transport, chemistry, and operator splitting methods. No method is currently accepted as systematically more accurate and efficient than any other. Integration of the transport and chemistry operators in PHREEQC using more appropriate splitting with advection-reaction on one side and diffusion-reaction on

| the other side leads to better resolution of chemical fronts as shown in the second and third |
|---|
| cases (Parkhurst and Appelo, 1999), but such front can be obtained by smaller time steps. It is |
| not only the integration but also the successive improvements of the methods that lead to |
| significantly more accurate schemes. TReacLab results remain however close, displaying the |
| same overall behavior both on solute and mineral concentrations. The interest of fully |
| segmented reactive transport implementations like in TReacLab is not motivated by the |
| accuracy and should not be used when other more integrated and optimized software are |
| appropriate and freely available. |
| Despite their lower accuracy, fully segmented implementations may be useful in situations |
| where flexibility is essential. It is the case when extensive modeling work has been performed |
| in independent software environments for transport or chemistry, and extensions to reactive |
| transport problems are required. Transport and chemistry solvers are then imposed and should |
| be coupled with as few specific developments as possible. For example, COMSOL and |
| PHREEQC have been interfaced here and in several other works because of their |
| complementarity (Nardi et al., 2014; Nasir et al., 2014; Wissmeier and Barry, 2011). It is |
| possible to specify advanced geometrical configurations in COMSOL through a convenient |
| graphical user interface (Azad et al., 2016). PHREEQC provides advanced capacities for |
| modeling complex geochemical systems with extensive database of reactions (Charlton and |
| Parkhurst, 2011; Parkhurst and Wissmeier, 2015). In such cases, building the structure of the |
| model may be the first and dominant issue in developing simulation capacities. That is when |
| codes like TReacLab can provide practical bridges for reactive transport systems. The |
| examples of section 4 however shows that they must be used with great care. Especially, the |
| called software may have different temporal integration schemes than the explicit and implicit |
| methods required by the SNIA and SIA coupling methods as discussed for the higher-order |
| schemes of COMSOL in section 4. Using codes like COMSOL may enhance robustness at a |

| 429 | certain cost of accuracy. Thus, implementation capacity does not guarantee validity. Validity |
|-----|--|
| 430 | must be carefully checked and argued with other comparable cases or with appropriate |
| 431 | convergence analysis. |
| 432 | Another targeted use of TReacLab concerns the development and test of new coupling |
| 433 | methods or strategies. Operator splitting can be performed with various methods including for |
| 434 | example adaptative time stepping (Belfort et al., 2007; Gasda et al., 2011). Global implicit |
| 435 | approaches that separate geochemical and transport software might also be more widely tested |
| 436 | providing the Jacobian of the chemical operator and taking into account current limitations |
| 437 | such as the difficulties to model precipitation/dissolution reactions (Amir and Kern, 2010). |
| 438 | TReacLab may then be used as a platform where interfaces to chemical and transport |
| 439 | operators are available and have been tested and documented for other coupling methods. |
| 440 | These applications are possible because TReacLab is a fully free and open software that can |
| 441 | be directly accessed and downloaded (https://github.com/TReacLab/TReacLab). The free and |
| 442 | open use of TReacLab has been dominant in its development and in the choices made for its |
| 443 | organization. The repository thus provides two main directories with sources and examples |
| 444 | respectively. Sources are organized in four main categories for chemistry, transport, coupler |
| 445 | and utilitaries. At the root of the chemistry, transport and coupler directories are the virtual |
| 446 | classes as main entries. Examples of instantiations are provided in the subdirectories. |
| 447 | Additional developments may take advantage of the documented examples provided at the |
| 448 | different levels of the software. |
| | |

6. Conclusion

449

450

451

We provide in the TReacLab code a fully segmented implementation of the coupling of independent geochemical and transport software. Coupling is based on a general expression of

the split-operator strategy with a set of classical methods. TReacLab should facilitate the development of reactive transport simulation capacities for independent reactive and transport software. Systematic comparison to the well-established PHREEQC model for uniform 1D reactive transport cases shows a good agreement with TReacLab. Systematic comparison against COMSOL for the 2D problem shows that full decoupling at the implementation level has a cost in accuracy. Sharp dissolution fronts of thermodynamically controlled reactions especially are generally smoothed in split-operator strategies. Steeper fronts might be recovered with smaller splitting time steps at larger computational costs. Beyond the implementation and the simulation capacity, consistency and validity of the numerical models should be systematically assessed. TReacLab can be freely accessed and used to promote the development of coupling methods and to provide additional modeling capacity for reactive transport coupling in geological media.

Appendix A: Implemented operator splitting methods

452

453

454

455

456

457

458

459

460

461

462

463

464

We detail the mathematical formulation for the sequential splitting (Geiser, 2009): 465

$$\frac{\partial Z^{1}}{\partial t} = \mathcal{L}_{1}Z^{1}, \qquad Z^{1}(x, t^{n}) = Z(x, t^{n}), \qquad t^{n} \leq t \leq t^{n+1}, \tag{A.1}$$

$$\frac{\partial Z^{2}}{\partial t} = \mathcal{L}_{2}Z^{2}, \qquad Z^{2}(x, t^{n}) = Z^{1}(x, t^{n+1}), \qquad t^{n} \leq t \leq t^{n+1}, \tag{A.2}$$

$$\frac{\partial Z^2}{\partial t} = \mathcal{L}_2 Z^2, \qquad Z^2(x, t^n) = Z^1(x, t^{n+1}), \qquad t^n \le t \le t^{n+1}, \tag{A.2}$$

$$Z(x,t^{n+1}) = Z^{2}(x,t^{n+1}),$$
 (A.3)

466 the additive splitting (Faragó et al., 2008a; Faragó et al., 2008b):

$$\frac{\partial Z^{1}}{\partial t} = \mathcal{L}_{1} Z^{1}, \qquad Z^{1}(x, t^{n}) = Z(x, t^{n}), \qquad t^{n} \leq t \leq t^{n+1}, \tag{A.4}$$

$$\frac{\partial Z^2}{\partial t} = \mathcal{L}_2 Z^2, \qquad Z^2(x, t^n) = Z(x, t^n), \qquad t^n \le t \le t^{n+1}, \tag{A.5}$$

$$Z(x,t^{n+1}) = Z^{1}(x,t^{n+1}) + Z^{2}(x,t^{n+1}) - Z(x,t^{n}), \tag{A.6}$$

the Strang splitting (Strang, 1968):

$$\frac{\partial Z^1}{\partial t} = \mathcal{L}_1 Z^1, \qquad Z^1(x, t^n) = Z(x, t^n), \qquad t^n \le t \le t^{n+1/2}, \tag{A.7}$$

$$\frac{\partial Z^2}{\partial t} = \mathcal{L}_2 Z^2, \qquad Z^2(x, t^n) = Z^1(x, t^{n+1/2}), \qquad t^n \le t \le t^{n+1}, \tag{A.8}$$

$$\frac{\partial z^{1}}{\partial t} = \mathcal{L}_{1} Z^{1}, \qquad Z^{1}(x, t^{n+1/2}) = Z^{2}(x, t^{n+1}), \quad t^{n+1/2} \le t \le t^{n+1}, \quad (A.9)$$

$$Z(x,t^{n+1}) = Z^{1}(x,t^{n+1}),$$
 (A.10)

and the symmetrically weighted splitting (SWS) (Csomós et al., 2005):

$$\frac{\partial Z^1}{\partial t} = \mathcal{L}_1 Z^1, \qquad Z^1(x, t^n) = Z(x, t^n), \qquad t^n \le t \le t^{n+1}, \qquad (A.11)$$

$$\frac{\partial Z^2}{\partial t} = \mathcal{L}_2 Z^2,$$
 $Z^2(x, t^n) = Z^1(x, t^{n+1}),$ $t^n \le t \le t^{n+1},$ (A.12)

$$\frac{\partial Z^{2*}}{\partial t} = \mathcal{L}_2 Z^{2*}, \qquad Z^{2*}(x, t^n) = Z(x, t^n), \qquad t^n \le t \le t^{n+1}, \qquad (A.13)$$

$$\frac{\partial Z^{2^*}}{\partial t} = \mathcal{L}_1 Z^{1*}, \qquad Z^{1*}(x, t^n) = Z^{2*}(x, t^{n+1}), \qquad t^n \le t \le t^{n+1}, \qquad (A.14)$$

$$Z(x,t^{n+1}) = \frac{z^2(x,t^{n+1}) + z^{1*}(x,t^{n+1})}{2}.$$
 (A.15)

- The alternating splitting algorithm (Valocchi and Malmstead, 1992) is based on a sequential
- splitting. It is defined by two successive splitting time steps with a permutation of the operator
- sequence between the splitting time steps.

472 Appendix B: Complementary notes on software organization

- We successively describe the general toolbox organization, the coupler, transport and
- 474 chemistry classes. We concretely show how operator splitting methods can be introduced and
- how other transport and geochemical codes can be connected.

B.1 Coupling methods

- The coupler is at the center of TReacLab as it performs the temporal integration and calls the
- 478 transport and chemistry solvers through the OS algorithm. In the pre-processing phase, it gets
- 479 the initial conditions and the temporal constrains of the integration. It is also in charge of

480

481

482

483

484

485

486

487

488

489

490

491

492

493

494

495

496

497

498

499

500

501

502

503

504

storing the required results before formatting and outputting them in the post-processing phase. Because the coupler is at the core of the toolbox, its methods remain generic. Interactions with the transport and chemistry solvers are also fully generic thanks to template interfaces calling external software and managing the exchange of information. Calling external software relies on the so-called Solve Engine method for both transport and chemistry software. Solve Engine takes as inputs the concentration data and the time step over which the integration must be performed. It returns the updated concentrations, a flag to check the success of the integration and an error message in case of failure to activate and inform the error management procedure mentioned in the former section. The coupler is based on a fixed structure of concentration data. Whatever the structure of concentrations in the transport and chemical codes, the structure of concentrations within the coupler is always the same. It consists in a matrix with in columns chemical species and in rows the position within the domain (Figure B.1). The size of the matrix is equal to the number of cells times the number of chemical species and components passed through the coupler. Chemical species include solutes and fixed species. As this is the sole link between the chemical code and the coupler also in charge of temporary results storage for the post-processing, it must transfer all quantities necessary for the algorithm and for the later extraction. The format of the matrix is set in the pre-processing phase and it is fixed for the whole simulation. TReacLab does not support yet any modification of species number to transfer between codes. Even if some solute species are absent over some time of the simulation, they will be transferred. This choice does not limit the capacity of the software as long as the chemical system is known from the beginning but might have some consequences on its performance in cases where solute composition strongly evolves. The choice of generality and flexibility, here like in other places, has a cost in efficiency. All modifications of concentration format are eventually performed in the interfaces between the coupler and the transport and chemistry solvers

505 (Figure 1).

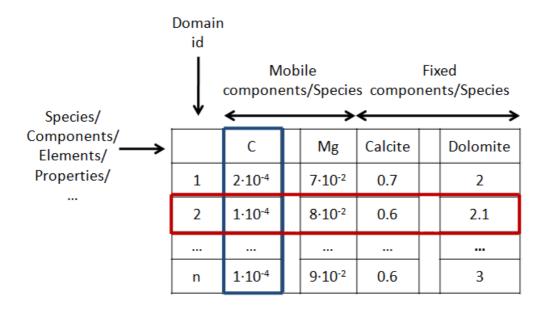


Figure B.1: Concentration format internal to the coupler class. To ensure generality, this structure of concentration is always the same and does not depend on the external transport and chemistry software. Species concentration are given in columns and are passed to the transport software as such. Concentrations at given locations are stored in rows with both mobile and fixed species. They are transferred either line per line or globally to the chemistry software. Fixed species are transferred from the chemistry code to the coupler to enable their possible use in the post-processing phase for results and outputs.

Thanks to the template methods calling the transport and chemical solvers and to the generic concentration format, operator splitting methods can be simply implemented. These are not more than a combination of simple calls of solvers passing and updating concentration information. Several sequential non-iterative techniques have thus been implemented, as detailed in section 2.3.

Specifications of the coupler are thus the name of the coupling method necessary to switch to the corresponding method in the coupler class, the temporal constrains of the integration and a vector of additional parameters. Temporal constrains of the integration are not only the initial and final times of the integration but also the times at which the solution must be stored. All time related parameters are stored into a time class. Additional parameters may be tolerances for example when using sequential iterative approaches. Instantiation of the coupler class thus consists in providing the identifier of the chosen coupling technique, the time constrains in the time class (initial time, final time, time to save the results, OS time step) and the additional parameters possibly needed by the algorithm.

B.2 Geochemical solver

Geochemical codes widely differ by their principles, the type of reactivity they consider and their input/output formats and parameters. We propose to normalize some of their interface to simplify exchanges with the coupler. In any case of equilibrium or kinetic reactions or of a mixed combination of them, geochemical codes steadily take concentrations, reaction constants, rate parameters, reaction times, as inputs and return output concentrations. All specifications linked to the choice of components, primary and secondary species should be set in the geochemical code or in the interface so that the geochemical solver does not have to be modified and the coupler remains generic. Whether components are used or not, the definition of the chemical system is not unique. Even when components are used, several alternative and reliable definitions can be chosen (Fang *et al.*, 2003; Hoffmann *et al.*, 2012; Molins *et al.*, 2004). Numerical and conceptual consistencies between the transport and chemical systems should thus be ensured externally before any implementation.

While solute concentrations are instantiated by the coupler and systematically passed to the geochemical solver, equilibrium and kinetic constants are considered as constant. They are

| defined once for all in the pre-processing phase. For example in PHREEQC, chemical |
|--|
| reactions and constants are already defined in databases like 'Phreeqc.dat' or 'Ilnl.dat'. |
| Initialization of mineral quantities is done at the beginning of the simulation when setting the |
| initial conditions through the coupler. The interface between the coupler and the geochemical |
| solver is made up of the Solve_Engine that calls the geochemical solver and the methods that |
| modify the concentration format. By default, the geochemical solver is instantiated and stored |
| for each of the nodes of the computational grid for the whole domain of the simulation. Any |
| data that are not passed to the coupler is, in general, kept in the instances of the geochemical |
| code. Another option is provided by software that allow simultaneous computations for |
| several independent batches like it is for example the case of PHREEQC. In such cases only |
| one instance of the geochemical solver is necessary. Exchanges of data between the coupler |
| and the geochemical solver are defined in the pre-processing phase and remain fixed for the |
| whole duration of the simulation. It is precisely at this stage that components are derived |
| through the algebraic operations of equation (6) and passed to the coupler. The coupler does |
| not manage the transformation of concentration and species but just their transfer between the |
| transport and geochemical solvers. The use of components does not fundamentally change the |
| calling sequence of the geochemical operator but modifies its interface to the coupler. |
| Components may be specified by the geochemical code like in PHREEQC or by the user in |
| the pre-processing phase by loading the matrix of U (equation (6)). In this latter case, |
| components are defined by the user in the pre-processing phase and are computed by the |
| interface that adapts the information to be passed through the coupler to the transport solver. |
| Connection of a new geochemical code requires essentially four operations. First, a new |
| daughter class of the template chemistry class must be defined. It can be built up using, as |
| template, one of the examples provided and described in the section 4. Second, an interface |
| must be created to filter the required information given from the coupler to the Solve Engine |

method. Third, an instantiation procedure should be provided whether it is internal or external to TReacLab. Fourth, the template Solve_Engine calling function of the geochemical solver must be written and optionally tested before being effectively used in reactive transport problems.

B.3 Transport solver

Despite the diversity of the transport mechanisms and numerical schemes to solve them, we provide here a basic interface designed mostly to address transport in a generic way. As previously stated, this approach assumes that transport parameters are not modified by the species concentration. This absence of feedback currently precludes density driven flows as well as permeability and porosity modifications due to precipitation or dissolution. TReacLab might be extended in this direction on the basis of slow evolutions of porosity or density. The transport operator relies on concentration independent parameters. We detail in the following the interaction between the coupler and the transport classes with the exchange of data and the instantiation of the transport solver. We will conclude this section with the development required to connect other transport codes.

While geochemical codes operate on species concentration on a given computational node, transport codes operate on a given species concentration over all the domain. In terms of data structure, each of the columns of the concentration array are successively transferred to the geochemical code and each of the rows (or linear combinations of rows) are given to the transport code (Figure B.1). The transport operator is thus iteratively called for each of the species or components explicitly specified in the interface between the coupler and the transport solver (Figure 1). The time range over which temporal integration should be performed and the identifiers of the transported species are also transferred to the transport solver. Species identification is essential when considering species sensitive diffusion

| 592 | coefficient. The transport solver returns the updated concentration field at the final time of the |
|-----|--|
| 593 | time range, an indicator of success or failure of the integration and a message to document |
| 594 | algorithm failures. The basic exchange of concentrations with the imposed integration times |
| 595 | are the sole requirements for the coupler to proceed. |
| 596 | All other parameters of the transport code should be set in the pre-processing phase, which |
| 597 | may become an important part of the eventual reactive transport code. In fact it does not cover |
| 598 | only the flow and transport parameters but more broadly the full structure of the domain, of |
| 599 | the computational grid, and of the boundary conditions. As for the geochemical code, the |
| 600 | transport code can be instantiated internally or externally. In case of internal definition, it |
| 601 | should contain at least the flow and transport properties, the morphology of the domain and |
| 602 | the structure of the computational grid (coordinates of the computational nodes). A default set |
| 603 | of classes is provided for 1D problems as templates for the morphology (domain definitions), |
| 604 | the computational grid (identification and coordinates of nodes and edges), the boundary |
| 605 | conditions (nature and values for boundary conditions) and the hydraulic and transport |
| 606 | properties. We recall as also said in section 2.2 that some operator splitting techniques might |
| 607 | impose limitations on the transport solver in terms of integration scheme or in terms of time |
| 608 | step (de Dieuleveult et al., 2009). Both the OS technique and the transport integration should |
| 609 | be chosen consistent. |
| 610 | Operations on the transport class are thus decomposed between the pre-processing and the |
| 611 | processing phases. Specifications of the operator with all necessary parameters is performed |
| 612 | in the pre-processing phase. Only generic exchanges of concentrations are needed in the |
| 613 | processing phase. Additional information would generally be needed externally to identify the |
| 614 | location of the computational nodes. More advanced information from the definition of the |
| 615 | domain, parameters and boundary conditions will be generally defined in the transport code |

| rather than in TReacLab. For example, Comsol or Modflow have their own grid definitions. |
|--|
| They are complete and efficient. It may be straightforwardly extracted and cross-referenced |
| with the results of TReacLab as long as the cell numbers correspond, a basic but necessary |
| requirement. This choice is motivated by both the generality and the simplicity of TReacLab. |
| It also highlights that TReacLab remains a coupler that transfers information and does not |
| process in any way the relation of concentrations between cells like a transport operator does. |
| The methodological choice of handling the spatial dimension of the problem within the |
| transport operator is not only operational. It is also ensuring the capacity to connect a wide |
| range of transport codes with their own logic and structure. For example, the multi-physics |
| software COMSOL has its own mesh generator methods and internal structures that should |
| not be duplicated in TReacLab but interfaced. Connecting other codes would thus require |
| reduced work as long as they can already be called from the same environment of |
| development (here MATLAB) on a discretized time basis. More in details, any new transport |
| code would require: 1) the development of the main calling function Solve_Engine to call it |
| from the coupler 2) the adaptation of the concentration format in the interface methods that |
| match the concentrations to the internal data structure of the external code, 3) the instantiation |
| of the transport class and 4) the access to the coordinates of the computational nodes for |
| outputs purposes. As for the geochemical code, implementation of the interface should be |
| checked before any full reactive transport coupling. This can be completed within TReacLab |
| by using an idle process instead of the geochemical code. |
| A descended and service We as less seeds does ANIDD A seed the ANID was been HOMNIOA and does the |
| Acknowledgements: We acknowledge ANDRA and the ANR project H2MNO4 under the |
| number ANR-12-MONU0012-01 for their financial support, Jocelyne Erhel and David L. |
| Parkhurst for constant and fruitful discussions and Javier Molinero for initial exchanges |

| _ | _ | _ |
|---|---|---|
| 6 | 3 | a |
| | | |

| 640 REFERENCI |
|----------------------|
|----------------------|

- Abarca, E., Nardi, A., Grandia, F., Molinero, J., 2013. Feedback between reactive transport
- and convective flow during CO2 migration in a saline aquifer, EGU General
- Assembly Conference Abstracts, p. 7707.
- 644 Amir, L., Kern, M., 2010. A global method for coupling transport with chemistry in
- heterogeneous porous media. Computational Geosciences 14, 465-481.
- Apoung-Kamga, J.-B., Have, P., Houot, J., Kern, M., Semin, A., 2009. Reactive Transport in
- Porous Media. ESAIM: Proceedings 28, 227 245.
- 648 Azad, V.J., Li, C., Verba, C., Ideker, J.H., Isgor, O.B., 2016. A COMSOL-GEMS interface
- for modeling coupled reactive-transport geochemical processes. Computers &
- 650 Geosciences 92, 79-89.
- Barry, D.A., Miller, C.T., Culligan-Hensley, P.J., 1996. Temporal discretisation errors in non-
- iterative split-operator approaches to solving chemical reaction/groundwater transport
- models. Journal of Contaminant Hydrology 22, 1-17.
- Bea, S.A., Carrera, J., Ayora, C., Batlle, F., Saaltink, M.W., 2009. CHEPROO: A Fortran 90
- object-oriented module to solve chemical processes in Earth Science models.
- 656 Computers & Geosciences 35, 1098-1112.
- Bear, J., 1972. Dynamics of fluids in porous media. Elsevier, New York.
- Belfort, B., Carrayrou, J., Lehmann, F., 2007. Implementation of Richardson extrapolation in
- an efficient adaptive time stepping method: applications to reactive transport and
- unsaturated flow in porous media. Transport in Porous Media 69, 123-138.

661 Bethke, C.M., 2007. Geochemical and biogeochemical reaction modeling. Cambridge 662 University Press. Beyer, C., Li, D., De Lucia, M., Kühn, M., Bauer, S., 2012. Modelling CO2-induced fluid-663 664 rock interactions in the Altensalzwedel gas reservoir. Part II: coupled reactive transport simulation. Environmental Earth Sciences 67, 573-588. 665 Carrayrou, J., Mosé, R., Behra, P., 2004. Operator-splitting procedures for reactive transport 666 667 and comparison of mass balance errors. Journal of Contaminant Hydrology 68, 239-268. 668 669 Charlton, S.R., Parkhurst, D.L., 2011. Modules based on the geochemical model PHREEOC 670 for use in scripting and programming languages. Computers & Geosciences 37, 1653-671 1663. 672 Clement, T.P., Sun, Y., Hooker, B.S., Petersen, J.N., 1998. Modeling Multispecies Reactive 673 Transport in Ground Water. Ground Water Monitoring & Remediation 18, 79-92. Commend, S., Zimmermann, T., 2001. Object-oriented nonlinear finite element 674 675 programming: a primer. Advances in Engineering Software 32, 611-628. COMSOL, A., 2010. COMSOL Multiphysics-LiveLink for Matlab User's Guide, comsol 4.1 676 677 edition. 678 COMSOL, A., 2012. 4.3 User's Guide. Comsol. 679 Csomós, P., Faragó, I., 2008. Error analysis of the numerical solution of split differential 680 equations. Mathematical and Computer Modelling 48, 1090-1106. 681 Csomós, P., Faragó, I., Havasi, Á., 2005. Weighted sequential splittings and their analysis. Computers & Mathematics with Applications 50, 1017-1031. 682

683 de Dieuleveult, C., Erhel, J., 2010. A global approach to reactive transport: application to the MoMas benchmark. Computational Geosciences 14, 451-464. 684 685 de Dieuleveult, C., Erhel, J., Kern, M., 2009. A global strategy for solving reactive transport 686 equations. Journal of Computational Physics 228, 6395-6410. 687 Diersch, H.-J.G., 1996. Interactive, graphics-based finite-element simulation system FEFLOW for modeling groundwater flow, contaminant mass and heat transport 688 689 processes. WASY Institute for Water Resource Planning and System Research Ltd., 690 Berlin, Germany. 691 Engesgaard, P., Kipp, K.L., 1992. A geochemical transport model for redox-controlled 692 movement of mineral fronts in groundwater flow systems: A case of nitrate removal 693 by oxidation of pyrite. Water Resources Research 28, 2829-2843. 694 Fang, Y., Yeh, G.-T., Burgos, W.D., 2003. A general paradigm to model reaction-based 695 biogeochemical processes in batch systems. Water Resources Research 39, 1083. Faragó, I., Gnandt, B., Havasi, Á., 2008a. Additive and iterative operator splitting methods 696 and their numerical investigation. Computers & Mathematics with Applications 55, 697 698 2266-2279. 699 Faragó, I., Thomsen, P.G., Zlatev, Z., 2008b. On the additive splitting procedures and their 700 computer realization. Applied Mathematical Modelling 32, 1552-1569. 701 Friedly, J.C., Rubin, J., 1992. Solute transport with multiple equilibrium-controlled or 702 kinetically controlled chemical reactions. Water Resources Research 28, 1935-1953. 703 Gasda, S.E., Farthing, M.W., Kees, C.E., Miller, C.T., 2011. Adaptive split-operator methods 704 for modeling transport phenomena in porous medium systems. Advances in Water 705 Resources 34, 1268-1282.

| 706 | Geiser, J., 2009. Decomposition methods for differential equations: theory and applications. |
|-----|--|
| 707 | CRC Press. |
| 708 | Hammond, G., Lichtner, P., Lu, C., Mills, R., 2012. Pflotran: reactive flow & transport code |
| 709 | for use on laptops to leadership-class supercomputers. Groundwater reactive transport |
| 710 | models, 141-159. |
| 711 | Hammond, G.E., Lichtner, P.C., Mills, R.T., 2014. Evaluating the performance of parallel |
| 712 | subsurface simulators: An illustrative example with PFLOTRAN. Water Resources |
| 713 | Research 50, 208-228. |
| 714 | Hoffmann, J., Kräutle, S., Knabner, P., 2010. A parallel global-implicit 2-D solver for reactive |
| 715 | transport problems in porous media based on a reduction scheme and its application to |
| 716 | the MoMaS benchmark problem. Computational Geosciences 14, 421-433. |
| 717 | Hoffmann, J., Kräutle, S., Knabner, P., 2012. A general reduction scheme for reactive |
| 718 | transport in porous media. Computational Geosciences 16, 1081-1099. |
| 719 | Hundsdorfer, W., Verwer, J.G., 2013. Numerical solution of time-dependent advection- |
| 720 | diffusion-reaction equations. Springer Science & Business Media. |
| 721 | Kazemi Nia Korrani, A., Sepehrnoori, K., Delshad, M., 2015. Coupling IPhreeqc with |
| 722 | UTCHEM to model reactive flow and transport. Computers & Geosciences 82, 152- |
| 723 | 169. |
| 724 | Kazemi Nia Korrani, A., Sepehrnoori, K., Delshad, M., 2016. A Mechanistic Integrated |
| 725 | Geochemical and Chemical-Flooding Tool for Alkaline/Surfactant/Polymer Floods. |
| 726 | SPE Journal 21, 32-54. |

| 727 | Kool, J., Van Genuchten, M.T., 1991. Hydrus: One-dimensional Variably Saturated Flow and |
|-----|--|
| 728 | Transport Model, Including Hysteresis and Root Water Uptake; Version 3.3. US |
| 729 | Salinity Laboratory. |
| 730 | Kräutle, S., Knabner, P., 2005. A new numerical reduction scheme for fully coupled |
| 731 | multicomponent transport-reaction problems in porous media. Water Resources |
| 732 | Research 41, W09414. |
| 733 | Kulik, D.A., Wagner, T., Dmytrieva, S.V., Kosakowski, G., Hingerl, F.F., Chudnenko, K.V., |
| 734 | Berner, U.R., 2013. GEM-Selektor geochemical modeling package: revised algorithm |
| 735 | and GEMS3K numerical kernel for coupled simulation codes. Computational |
| 736 | Geosciences 17, 1-24. |
| 737 | Lie, K., 2014. An introduction to reservoir simulation using MATLAB: User guide for the |
| 738 | Matlab reservoir simulation toolbox (MRST), SINTEF ICT, Department of Applied |
| 739 | Mathematics, Oslo, Norway. |
| 740 | Liu, J., Ewing, R.E., 2005. An Operator Splitting Method for Nonlinear Reactive Transport |
| 741 | Equations and Its Implementation Based on DLL and COM, In: Zhang, W., Tong, W., |
| 742 | Chen, Z., Glowinski, R. (Eds.), Current Trends in High Performance Computing and |
| 743 | Its Applications: Proceedings of the International Conference on High Performance |
| 744 | Computing and Applications, August 8-10, 2004, Shanghai, P.R. China. Springer |
| 745 | Berlin Heidelberg, Berlin, Heidelberg, pp. 93-102. |
| 746 | Marty, N.C.M., Munier, I., Gaucher, E.C., Tournassat, C., Gaboreau, S., Vong, C.Q., Giffaut, |
| 747 | E., Cochepin, B., Claret, F., 2014. Simulation of Cement/Clay Interactions: Feedback |
| 748 | on the Increasing Complexity of Modelling Strategies. Transport in Porous Media 104, |
| 749 | 385-405. |

750 Mayer, K., 2000. MIN3P V1. 0 User Guide. University of Waterloo, Department of Earth 751 Sciences 26. 752 McDonald, M.G., Harbaugh, A.W., 1988. A modular three-dimensional finite-difference 753 ground-water flow model, In: Survey, U.S.G. (Ed.), Denver, Colorado. 754 MIKE(DHI), 2016. piChem: A FEFLOW Plugin for Advanced Geochemical Reactions, User 755 Guide. 756 Molins, S., Carrera, J., Ayora, C., Saaltink, M.W., 2004. A formulation for decoupling 757 components in reactive transport problems. Water Resources Research 40, W10301. 758 Morel, F.M., Hering, J.G., 1993. Principles and applications of aquatic chemistry. John Wiley 759 & Sons. 760 Multiphysics, C., 2008. Pesticide transport and reaction in soil. Earth Science Module Model 761 Library. Muniruzzaman, M., Rolle, M., 2016. Modeling multicomponent ionic transport in 762 763 groundwater with IPhreeqc coupling: Electrostatic interactions and geochemical 764 reactions in homogeneous and heterogeneous domains. Advances in Water Resources 98, 1-15. 765 Nardi, A., Idiart, A., Trinchero, P., de Vries, L.M., Molinero, J., 2014. Interface COMSOL-766 PHREEQC (iCP), an efficient numerical framework for the solution of coupled 767 multiphysics and geochemistry. Computers & Geosciences 69, 10-21. 768 769 Nasir, O., Fall, M., Evgin, E., 2014. A simulator for modeling of porosity and permeability 770 changes in near field sedimentary host rocks for nuclear waste under climate change 771 influences. Tunnelling and Underground Space Technology 42, 122-135.

| 772 | Parkhurst, D.L., Appelo, C., 1999. User's guide to PHREEQC (Version 2): A computer |
|-----|---|
| 773 | program for speciation, batch-reaction, one-dimensional transport, and inverse |
| 774 | geochemical calculations. |
| 775 | Parkhurst, D.L., Kipp, K.L., Engesgaard, P., Charlton, S.R., 2004. PHAST, a program for |
| 776 | simulating ground-waterflow, solute transport, and multicomponent geochemica |
| 777 | reactions. USGS Techniques and Methods 6, A8. |
| 778 | Parkhurst, D.L., Wissmeier, L., 2015. PhreeqcRM: A reaction module for transport simulators |
| 779 | based on the geochemical model PHREEQC. Advances in Water Resources 83, 176 |
| 780 | 189. |
| 781 | Patel, R., Perko, J., Jacques, D., De Schutter, G., Ye, G., Van Breugel, K., 2013. Lattice |
| 782 | Boltzmann based multicomponent reactive transport model coupled with geochemica |
| 783 | solver for scale simulations, 5th International Conference on Computational Methods |
| 784 | for Coupled Problems in Science and Engineering. International Center for Numerica |
| 785 | Methods in Engineering (CIMNE), pp. 806-817. |
| 786 | Peterson, S., Hostetler, C., Deutsch, W., Cowan, C., 1987. MINTEQ user's manual. Pacific |
| 787 | Northwest Lab., Richland, WA (USA); Nuclear Regulatory Commission, Washington |
| 788 | DC (USA). Div. of Waste Management. |
| 789 | Prommer, H., Davis, G., Barry, D., 1999. PHT3D—A three-dimensional biogeochemica |
| 790 | transport model for modelling natural and enhanced remediation. Contaminated Site |
| 791 | Remediation: Challenges Posed by Urban and Industrial Contaminants. Centre for |
| 792 | Groundwater Studies, Fremantle, Western Australia, 351-358. |
| 793 | Pruess, K., Oldenburg, C., Moridis, G., 1999. TOUGH2 user's guide version 2. Lawrence |
| | |

Register, A.H., 2007. A guide to MATLAB object-oriented programming. CRC Press.

796 Rouson, D., Xia, J., Xu, X., 2011. Scientific software design: the object-oriented way. 797 Cambridge University Press. Saaltink, M.W., Ayora, C., Carrera, J., 1998. A mathematical formulation for reactive 798 799 transport that eliminates mineral concentrations. Water Resources Research 34, 1649-1656. 800 801 Saaltink, M.W., Carrera, J., Ayora, C., 2001. On the behavior of approaches to simulate 802 reactive transport. Journal of Contaminant Hydrology 48, 213-235. 803 Saaltink, M.W., Yakirevich, A., Carrera, J., Ayora, C., 2011. Fluid flow, solute and heat 804 transport equations. Geochemical Modeling of Groundwater, Vadose and Geothermal 805 Systems, 83. 806 Simpson, M.J., Landman, K.A., 2008. Theoretical analysis and physical interpretation of 807 temporal truncation errors in operator split algorithms. Mathematics and Computers in 808 Simulation 77, 9-21. 809 Šimůnek, J., Jacques, D., Van Genuchten, M.T., Mallants, D., 2006. Multicomponent 810 geochemical transport modeling using HYDRUS-1D and HP1. J. Am. Water Resour. Assoc 42, 1537-1547. 811 Šimůnek, J., Vogel, T., Van Genuchten, M.T., 1994. The SWMS-2D code for simulating 812 813 water and solute transport in two dimensional variably saturated media–Version 1.2. 814 Research Report 132, US Salinity Lab., Agric. Res. Serv. USDA, Riverside, 815 California, USA. 816 Skeel, R.D., Berzins, M., 1990. A method for the spatial discretization of parabolic equations in one space variable. SIAM journal on scientific and statistical computing 11, 1-32. 817

| 818 | Steefel, C., 2009. CrunchFlow software for modeling multicomponent reactive flow and |
|-----|---|
| 819 | transport. User's manual. Earth Sciences Division. Lawrence Berkeley, National |
| 820 | Laboratory, Berkeley, CA. October, 12-91. |
| 821 | Steefel, C.I., DePaolo, D.J., Lichtner, P.C., 2005. Reactive transport modeling: An essential |
| 822 | tool and a new research approach for the Earth sciences. Earth and Planetary Science |
| 823 | Letters 240, 539-558. |
| 824 | Steefel, C.I., MacQuarrie, K.T., 1996. Approaches to modeling of reactive transport in porous |
| 825 | media. Reviews in Mineralogy and Geochemistry 34, 85-129. |
| 826 | Strang, G., 1968. On the construction and comparison of difference schemes. SIAM Journal |
| 827 | on Numerical Analysis 5, 506-517. |
| 828 | Thouvenot, P., Bildstein, O., Munier, I., Cochepin, B., Poyet, S., Bourbon, X., Treille, E., |
| 829 | 2013. Modeling of concrete carbonation in deep geological disposal of intermediate |
| 830 | level waste, EPJ Web of Conferences. EDP Sciences, p. 05004. |
| 831 | Trotignon, L., Devallois, V., Peycelon, H., Tiffreau, C., Bourbon, X., 2007. Predicting the |
| 832 | long term durability of concrete engineered barriers in a geological repository for |
| 833 | radioactive waste. Physics and Chemistry of the Earth, Parts A/B/C 32, 259-274. |
| 834 | Valocchi, A.J., Malmstead, M., 1992. Accuracy of operator splitting for advection-dispersion- |
| 835 | reaction problems. Water Resources Research 28, 1471-1476. |
| 836 | Van der Lee, J., 2002. CHESS Software for Geochemistry. Hydrology and Environmental |
| 837 | Science, École des Mines de Paris, Fontainebleau, France. |
| 838 | van der Lee, J., De Windt, L., Lagneau, V., Goblet, P., 2003. Module-oriented modeling of |
| 839 | reactive transport with HYTEC. Computers & Geosciences 29, 265-275. |

| 840 | Van Genuchten, M.T., Alves, W., 1982. Analytical solutions of the one-dimensional |
|-----|---|
| 841 | convective-dispersive solute transport equation. United States Department of |
| 842 | Agriculture, Economic Research Service. |
| 843 | Wissmeier, L., Barry, D.A., 2011. Simulation tool for variably saturated flow with |
| 844 | comprehensive geochemical reactions in two- and three-dimensional domains. |
| 845 | Environmental Modelling & Software 26, 210-218. |
| 846 | Yeh, G.T., Tripathi, V.S., 1989. A critical evaluation of recent developments in |
| 847 | hydrogeochemical transport models of reactive multichemical components. Water |
| 848 | Resources Research 25, 93-108. |
| 849 | Zhang, F., 2012. Groundwater reactive transport models. Bentham Science Publishers. |
| 850 | Zheng, C., Wang, P.P., 1999. MT3DMS: a modular three-dimensional multispecies transport |
| 851 | model for simulation of advection, dispersion, and chemical reactions of contaminants |
| 852 | in groundwater systems; documentation and user's guide. DTIC Document. |
| 853 | |

Highlights

- Object-oriented implementation of non-intrusive couplings for reactive transport
- Validation for 3 equilibrium and kinetically controlled 1D cases with PhreeqC and one 2D case with COMSOL.
- Illustration of implementation flexibility with different transport and reaction codes