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RESEARCH ARTICLE

10.1029/2019JD030419

Unprecedented Atmospheric Ammonia Concentrations
Detected in the High Arctic From the 2017
Canadian Wildfires

Key Points:

- The 2017 Canadian wildfires resulted in the largest observed NH_3 total column enhancements in the high Arctic from 1999–2017
- GEOS-Chem surface layer NH_3 was enhanced by 0.01–0.11 ppbv (14–550%) in the Canadian high Arctic from 15–23 August 2017 due to wildfires
- Wildfires may be an important episodic source of NH_3 in the summertime high Arctic in addition to the persistent seabird colony source

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Abstract From 17–22 August 2017 simultaneous enhancements of ammonia (NH_3), carbon monoxide (CO), hydrogen cyanide (HCN), and ethane (C_2H_6) were detected from ground-based solar absorption Fourier transform infrared (FTIR) spectroscopic measurements at two high-Arctic sites: Eureka (80.05°N, 86.42°W) Nunavut, Canada, and Thule (76.53°N, 68.74°W), Greenland. These enhancements were attributed to wildfires in British Columbia and the Northwest Territories of Canada using FLEXPART back-trajectories and fire locations from Moderate Resolution Imaging Spectroradiometer (MODIS) and found to be the greatest observed enhancements in more than a decade of measurements at Eureka (2006–2017) and Thule (1999–2017). Observations of gas-phase NH_3 from these wildfires illustrate that boreal wildfires may be a considerable episodic source of NH_3 in the summertime high Arctic. Comparisons of GEOS-Chem model simulations using the Global Fire Assimilation System (GFASv1.2) biomass burning emissions to FTIR measurements and Infrared Atmospheric Sounding Interferometer (IASI) measurements showed that the transport of wildfire emissions to the Arctic was underestimated in GEOS-Chem. However, GEOS-Chem simulations showed that these wildfires contributed to surface layer NH_3 and NH_4^+ enhancements of 0.01–0.11 ppbv and 0.05–1.07 ppbv, respectively, over the Canadian Archipelago from 15–23 August 2017.

1. Introduction

Ammonia (NH_3) is an important form of reactive nitrogen, it and its particulate form ammonium (NH_4^+) are the main forms of reactive nitrogen in the atmosphere. NH_3 is a precursor to the formation of particulate matter, which has a negative impact on air quality (Erisman et al., 2011; Pope et al., 2009). In high concentrations, NH_3 is toxic to biodiversity (Aneja et al., 2001; Asman et al., 1998; Galloway et al., 2003) and contributes to the acidification of soil and water bodies, which can lead to the eutrophication of ecosystems (Bobbink et al., 1998, 2010). The Arctic is free of the common anthropogenic sources of NH_3 , such as agriculture practices (Bouwman et al., 1997; Dentener & Crutzen, 1994; Schlesinger & Hartley, 1992), and the short lifetime of NH_3 , typically less than 1 day (Lefer et al., 1999) was thought to limit the long-range transport of gas-phase NH_3 from the lower latitudes. Biomass burning has been identified as a major source of NH_3 (Bouwman et al., 2002), with boreal wildfires, in particular, found to contribute to NH_4^+ in the Arctic (Fisher et al., 2011). Detection of boreal wildfire NH_3 emissions transported to the high Arctic was identified using ground-based Fourier transform infrared (FTIR) measurements in Eureka (80.05°N, 86.42°W), Nunavut, by Lutsch et al. (2016). Solar absorption FTIR measurements were used to retrieve total column abundances of NH_3 and were the first long-term measurements of NH_3 in the high Arctic (2006–2017). In

July and August 2014, simultaneous enhancements of NH_3 and the long-lived biomass burning tracers carbon monoxide (CO), hydrogen cyanide (HCN), and ethane (C_2H_6) were detected and attributed to wildfires in the Great Slave Lake (61.67°N, 114.00°W) region of the Northwest Territories. Enhancements of these species in 2014 were also detected using ground-based FTIR measurements in Toronto (43.66°N, 79.40°W), Ontario (Lutsch et al., 2016). Accounting for loss of NH_3 during transport to each site, Lutsch et al. (2016) estimated a gas-phase NH_3 lifetime in the smoke plume of ≈ 2 days, which would allow for long-range transport of gas-phase NH_3 . However, since FTIR measurements of this kind sample the smoke plume at a distance from the fire source, the lifetime may be considerably shorter near the fire source due to the rapid gas-aerosol partitioning (Paulot et al., 2017). It is also possible that the lifetime of NH_3 may be prolonged by injection of wildfire emissions into the free troposphere, mitigating depositional and chemical losses; however, the physical and chemical mechanisms leading to the long-range transport of gas-phase NH_3 within a smoke plume remain uncertain.

In the high Arctic, in situ shipborne measurements of NH_3 were performed by Wentworth et al. (2016) as part of the 2014 NETCARE (Network on Climate and Aerosols; Abbatt et al., 2019) Campaign in the Canadian Archipelago during July and August 2014. These measurements indicated enhancements of surface layer NH_3 in the Lancaster Sound region (74°N, 84°W), which were attributed in part to the 2014 Northwest Territories wildfires. At higher latitudes near Alert (82.50°N, 62.35°W), Nunavut, surface layer NH_3 was attributed to local seabird colony and tundra sources. Surface layer NH_3 concentrations were found to range from 40 to 870 pptv during the 2014 NETCARE Campaign, with greatest concentrations occurring near Lancaster Sound. It has been concluded from several studies that seabird guano is a persistent source of NH_3 in the high Arctic summer (Blackall et al., 2007; Croft et al., 2016; Riddick et al., 2012; Wentworth et al., 2016). Emissions of NH_3 from seabird colonies have been associated with neutralization of aerosols (Wentworth et al., 2016) and the production of cloud condensation nuclei, which may contribute to a negative radiative forcing of 1–2 W/m^2 in regions of high seabird colony NH_3 emissions (Croft et al., 2016). As was suggested by Lutsch et al. (2016), it is likely that boreal wildfires may be a considerable episodic source of NH_3 to the high Arctic, in addition to the persistent seabird colony source during the Arctic warm season. Currently, the influence of wildfire NH_3 emissions on the Arctic remains unknown due to the scarcity of observations and episodic nature of wildfire events. Furthermore, the relative influence of wildfire and seabird colony NH_3 has not been evaluated on a long-term time scale.

Global satellite observations of NH_3 have been routinely made by several thermal infrared instruments since 2002: Atmospheric Infrared Sounder (AIRS; Warner et al., 2016; starting in 2002), the Tropospheric Emission Spectrometer (TES; Shephard et al., 2011, 2015; starting in 2004), Infrared Atmospheric Sounding Interferometer (IASI; Van Damme et al., 2015; starting in 2006), and Cross-track Infrared Sounder (CrIS; Shephard & Cady-Pereira, 2015; starting in 2011). These instruments have been used in a number of studies to quantify NH_3 emissions from boreal wildfire sources. Emission ratios of NH_3 were derived from TES observations for the 2007 Alaskan wildfires (Alvarado et al., 2011) and the 2010 Russian wildfires (R'Honi et al., 2013). CrIS has been utilized for studies of wildfires in Northern Canada (Adams et al., 2019; Kharol et al., 2018; Whaley et al., 2018); and the IASI instrument has been used to investigate emissions from the 2008 Siberian wildfires (Coheur et al., 2009), the 2016 Fort McMurray wildfires in Alberta, Canada (Adams et al., 2019), and Alaskan wildfires from 2009–2013 (Paulot et al., 2017). IASI observations have also been utilized in several studies of tropical biomass burning (Whitburn et al., 2015, 2016, 2017) in addition to global emission studies (Van Damme et al., 2018). Observations of NH_3 in the Arctic from these satellite instruments remain sparse due to the low concentrations of NH_3 , in addition to the low thermal contrast and high cloud cover of the Arctic region.

The use of ground-based FTIR instruments in the Arctic is advantageous as these instruments provide sufficient sensitivity to the free troposphere where transported wildfire emissions are most abundant (Petetin et al., 2018), in addition to having a lower detection limit than satellite observations (Dammers et al., 2016). The study of Lutsch et al. (2016) utilized the FTIR NH_3 retrieval method of Dammers et al. (2015) to retrieve NH_3 total columns at Eureka, Nunavut. The use of FTIR measurements also allows for simultaneous measurements of the long-lived biomass burning tracers carbon monoxide (CO), hydrogen cyanide (HCN), and ethane (C_2H_6), along with a multitude of other biomass burning species (Paton-Walsh et al., 2005; Paton-Walsh et al., 2010; Vigouroux et al., 2012; Viatte et al., 2013, 2015; Zhao, 2002). Measurements of CO, HCN, and C_2H_6 in addition to NH_3 provides insight into the origin, transport, and aging of the smoke plume. The utilization of ground-based FTIR measurement sites in the Arctic provides a long-term time

series of NH_3 and other trace gases where measurements from other platforms are lacking sensitivity or are nonexistent.

The transport of wildfire emissions of NH_3 to the Arctic has not been thoroughly investigated in global chemical transport models, while in general, accurate modeling of the Arctic atmosphere remains difficult. Both Wentworth et al. (2016) and Croft et al. (2016) used the GEOS-Chem chemical transport model (CTM) to investigate the influence of seabird colony NH_3 in the Canadian Arctic. Wildfire emissions of NH_3 were included in these studies but were not explicitly examined, although Wentworth et al. (2016) noted that in the GEOS-Chem model, NH_3 from the 2014 Northwest Territories wildfires contributed significantly to surface layer NH_3 in the low Arctic. Kharol et al. (2018) examined the deposition of NH_3 at northern latitudes with CrIS observations and the GEM-MACH (Global Environmental Multi-scale-Modelling Air quality and Chemistry) model and found NH_3 deposition fluxes 2 to 3 times greater in the presence of wildfires. Deposition of NH_3 in the Arctic was not examined in this study. Due to the scarcity of NH_3 measurements in high-latitude regions and the episodic nature of wildfire events, more effort is needed to evaluate model performance in order to infer the influence of wildfire NH_3 in the Arctic from model results.

In this paper, we examine two separate wildfire events in British Columbia (BC) and the Northwest Territories (NWT) of Canada in August 2017. We aim to quantify the influence of transported NH_3 from these events in the Arctic with the use of FTIR measurements at two high-Arctic sites: Eureka, Nunavut, and Thule, Greenland. The simultaneous influence of the BC and NWT wildfires resulted in the greatest observed enhancements of NH_3 in the decade-long time series at both Eureka (2006–2017) and Thule (1999–2017). The exceptional magnitude of these enhancements was due to the intensity of both the BC and NWT wildfires, which resulted in volcano-like injection of smoke throughout the troposphere and lower stratosphere (Khaykin et al., 2018; Peterson et al., 2018). The record-setting magnitude of the 2017 BC wildfires, with over 1.2×10^6 ha burned (BC Wildfire Service, 2017), was the result of persistent warm temperatures and increased wildfire risk made substantially greater by anthropogenic climate change (Kirchmeier-Young et al., 2018). Given the extreme magnitude of this event, these wildfires present an opportunity to investigate the influence of wildfire NH_3 emissions on the Canadian high Arctic. The exceptional magnitude of transported NH_3 emissions from the 2017 BC and NWT wildfires may serve as a harbinger of future events as wildfire frequency and intensity are expected to increase in the future (Boulanger et al., 2014; Flannigan et al., 2009; Wotton et al., 2010, 2017).

The following sections in this paper are structured as follows. The ground-based FTIR instruments, GEOS-Chem model, and IASI observations used in this study are first described in section 2. The FTIR time series of NH_3 , CO, HCN, and C_2H_6 are presented and discussed in section 3.1. Source attribution of the detected enhancements at Eureka and Thule are then examined with the use of the FLEXPART Lagrangian transport model in section 3.2, while in section 3.3, enhancement ratios of NH_3 , HCN, and C_2H_6 are calculated for fire-affected FTIR measurements at Eureka and Thule. These enhancement ratios give further insight into the wildfire source influence at each site. To interpret the FTIR measurements, the GEOS-Chem CTM is compared to FTIR and IASI measurements in sections 3.4 and 3.5, respectively, for NH_3 and CO. Lastly, the contribution of the transported wildfire emissions to surface and total column NH_3 and NH_4^+ in the Canadian high Arctic is evaluated in the GEOS-Chem model for the fire-affected period of 15–23 August 2017 and compared to the contribution from local seabird colony NH_3 sources.

2. Methods

2.1. FTIR Sites and Instruments

Trace gas profiles and integrated column amounts are retrieved from solar-absorption spectra obtained by two FTIR spectrometers: a Bruker IFS 125HR at the Polar Environment Atmospheric Research Laboratory (PEARL) in Eureka (80.05°N, 86.42°W), Nunavut, and a Bruker IFS 125HR at the Thule Air Base (TAB) in Thule (76.53°N, 68.74°W), Greenland. Detailed descriptions of the Eureka and Thule FTIR instruments are given by Batchelor et al. (2009) and Hannigan et al. (2009), respectively. Measurements from both instruments contribute to the Network for Detection of Atmospheric Composition Change (NDACC; www.ndacc.org). Measurements are made in the mid-infrared at a resolution of 0.0035 cm^{-1} using a KBr beam splitter with a spectral coverage of $700\text{--}5,000 \text{ cm}^{-1}$, with both InSb ($1,850\text{--}10,000 \text{ cm}^{-1}$) and HgCdTe ($600\text{--}6,000 \text{ cm}^{-1}$) detectors and a series of optical filters. For both instruments, spectra are recorded continuously during clear-sky daylight hours by alternating between optical filters while a heliostat system tracks the Sun.

2.2. FTIR Retrievals

Solar-absorption spectra were processed using the SFIT4 retrieval algorithm (<https://wiki.ucar.edu/display/sfit4/>) based on the optimal estimation method (Rodgers & Connor, 2003). Retrievals of CO, HCN, and C₂H₆ at Eureka and Thule are performed following NDACC Infrared Working Group (IRWG) recommendations as described in Viatte et al. (2013, 2014, 2015) and Lutsch et al. (2016) and references therein. For both sites, retrievals of NH₃ were performed following the procedure of Lutsch et al. (2016) based on the method of Dammers et al. (2015). The SFIT4 forward model is a line-by-line radiative transfer model which uses temperature and pressure profiles from the National Centers for Environmental Prediction (NCEP) and spectroscopic parameters from HITRAN 2008 (Rothman et al., 2009). For each site, 40-year average (1980–2020) profiles from the Whole-Atmosphere Community Climate Model (WACCM V4; Eyring et al., 2007) were used as the a priori profiles for CO, HCN, and C₂H₆, while a priori profiles for NH₃ were derived from balloon-based measurements (Toon et al., 1999) taken near Fairbanks (64.84°N, 147.72°W), Alaska following Dammers et al. (2015).

For all retrievals, a full error analysis was performed following Rodgers and Connor (2003). The retrieval uncertainties include forward model parameter error and measurement noise error. Adding these in quadrature, average uncertainties for the retrieved total columns of CO, HCN, C₂H₆, and NH₃ are 5.4%, 4.6%, 4.2%, and 30.2%, respectively, for Eureka and 2.6%, 6.9%, 1.6%, and 12.5%, respectively, for Thule. The average degrees of freedom for signal (DOFS) for CO, HCN, C₂H₆, and NH₃ are 2.0, 2.4, 1.7, and 1.0, respectively, for Eureka and 3.0, 3.5, 1.7, and 1.1, respectively, for Thule. For CO, HCN, and C₂H₆, the total column-averaging kernels show peak values in the upper troposphere at both sites, whereas the total column-averaging kernels of NH₃ peak near the surface and decrease upward to 5 km. The shape of the total column-averaging kernel profiles is due to the vertical distribution of the species, which follows a similar structure to the a priori profile. The minimum retrieved total columns of NH₃ are on the order of $\sim 10^{14}$ molec/cm² in clear-sky conditions, which corresponds to NH₃ surface concentration of ~ 0.1 – 0.3 ppbv. The surface concentration is estimated as the retrieved volume mixing ratio of the lowest level in the FTIR retrieval grid. The ambient concentrations of CO, HCN, and C₂H₆ exceed the detection limits of the FTIR instrument under normal clear-sky conditions.

2.3. IASI Observations

We use CO and NH₃ total column abundances retrieved from observations by the IASI instrument on board the Metop-A satellite. The satellite has a polar Sun-synchronous orbit providing twice daily global coverage at 9:30 local time and 21:30 local time overpasses. The instrument has a scan swath width of approximately 2,200 km and observations have a circular footprint of ≈ 12 km in diameter at nadir (Clerbaux et al., 2009). We used the most recent CO and NH₃ data products (<http://iasi.aeris-data.fr/>). For CO, the FORLI-CO product (Hurtmans et al., 2012) was used, which has been validated against ground-based FTIR measurements and found to show good agreement with no significant bias (Kerzenmacher et al., 2012). For NH₃, we use the latest ANNI-NH₃-v2.1 product (Van Damme et al., 2017). The current ANNI-NH₃-v2.1 version has not yet been compared to FTIR measurements, although the previous version, v1 (Whitburn et al., 2016), was validated using ground-based FTIR observations by Dammers et al. (2016) and Dammers et al. (2017). IASI has been found to underestimate NH₃ total columns by an average of $\sim 32\%$ in comparison to FTIR observations. Greater underestimation ($>50\%$) was observed for FTIR sites with low NH₃ total columns ($< 5 \times 10^{15}$ molec/cm²), and IASI showed better agreement ($\sim 22\%$) for FTIR sites with greater local NH₃ total columns. The underestimation of IASI is the result of the dependence of the retrieval on the NH₃ concentration and the thermal contrast at the surface. For instance, Van Damme et al. (2015) and Van Damme et al. (2014) estimated a detection limit of 9.68×10^{15} molec/cm² for a thermal contrast of 20 K, while a thermal contrast of 10 K corresponds to a detection limit of 1.69×10^{16} molec/cm². An approximate 3 ppbv surface layer NH₃ concentration was estimated as an average detection limit of IASI, although in an NH₃ plume with well-mixed concentrations in a thick boundary layer, the detection limit could be as low as 1 ppbv at the surface with sufficient thermal contrast (Clarisse et al., 2010).

2.4. GEOS-Chem Chemical Transport Model

The GEOS-Chem CTM (www.geos-chem.org), version v11.01, was used at a horizontal resolution of $2^\circ \times 2.5^\circ$ with 47 vertical hybrid levels from the surface to 0.01 hPa in altitude. The model is driven with assimilated meteorological fields from the Goddard Earth Observing System version 5.11.0 (GEOS-FP) from the NASA Global Model and Assimilation Office (GMAO). For each simulation, an 8-month model spin-up was performed from 1 January to 1 August 2017. Transport and chemical operator time steps of 10 and

20 min, respectively, were chosen to minimize simulation errors (Philip et al., 2016). Global anthropogenic emissions are provided by the Emissions for Database for Global Atmospheric Research (EDGARv4.3.1; Janssens-Maenhout et al., 2013) emissions inventory. Global anthropogenic and natural NH_3 emissions are provided by Bouwman et al. (1997). Biomass burning emissions from the Global Fire Assimilation System (GFASv1.2 Kaiser et al., 2012) are used. GFAS provides global emissions for open fires on a $0.1^\circ \times 0.1^\circ$ grid at 3-hourly resolution, which are derived from assimilation of fire radiative power (FRP) observations of the Moderate Resolution Imaging Spectroradiometer (MODIS) on the Aqua and Terra satellites. The use of the GFAS emission inventory was chosen based on the availability of emissions for the 2017 period compared to the Global Fire Emission Database (GFEDv4), and the better representation of boreal North American emissions in comparison to the Fire Inventory from NCAR (FINNv1.0; Shi et al., 2015). Seabird colony NH_3 emissions based on Riddick et al. (2012) were also included to investigate the contribution of seabird colony NH_3 emissions in the Arctic. Emissions were considered for seabird colonies located north of 50°N following Wentworth et al. (2016) and Croft et al. (2016).

Surface emissions in GEOS-Chem v11.01, including biomass burning emissions, are released within the boundary layer, and boundary layer mixing is implemented using the nonlocal mixing scheme of Holtslag and Boville (1993). Simulated aerosol species include sulfate-nitrate-ammonium (Park, 2004; Park et al., 2006), with the sulfate-nitrate-ammonium chemistry based on the ISORROPIA II thermodynamic model (Fountoukis & Nenes, 2007), which partitions ammonia and nitric acid between the gas and aerosol phases. Dry deposition in GEOS-Chem is implemented following a standard resistance in series scheme (Wesely, 1989), with surface resistances for sulfate, nitrate, and ammonium aerosols as described by Zhang et al. (2001). Wet removal takes place in large-scale clouds and convective updrafts (Liu et al., 2001).

3. Results & Discussion

3.1. FTIR Time Series

The time series of CO, HCN, C_2H_6 , and NH_3 total columns for Eureka and Thule are shown in Figure 1. For CO and C_2H_6 , similar seasonal cycles, with a maximum in February and March and a minimum in July and August are observed as a result of the common anthropogenic source of these species from fossil fuel combustion and their main sink due to reaction with the hydroxyl radical (OH; Stein & Rudolph, 2007; Xiao et al., 2008). During the polar night, OH formation ceases and transport of pollutants from midlatitude regions (Klonecki, 2003; Stohl, 2006; Shindell et al., 2008) allows CO and C_2H_6 to accumulate in the Arctic atmosphere. Decreasing total columns are observed throughout the spring and summer as a result of increased loss due to the production of OH during the sunlit season. Average lifetimes of CO and C_2H_6 are ≈ 2 months (Singh & Zimmerman, 1992) and 3 months (Xiao et al., 2008), respectively, making them suitable tracers of long-range transport.

The seasonal cycles of HCN and NH_3 are opposite in phase to those of CO and C_2H_6 as result of their different sources. The primary source of HCN is biomass burning, but plant and fungal emissions also represent considerable sources (Li et al., 2003). Due to its long atmospheric lifetime of ~ 2 – 4 months (Li et al., 2000), HCN tends to accumulate in the atmosphere. Its main loss is a result of dry deposition to the ocean (Li et al., 2003) and reaction with OH and $\text{O}^1(\text{D})$ (Cicerone & Zellner, 1983). For NH_3 , the main source in the Arctic is from the hydrolysis of migratory seabird colony guano (Blackall et al., 2007; Riddick et al., 2012; Wentworth et al., 2016), while wildfires may present a considerable episodic source (Lutsch et al., 2016). As a result, the seasonal cycle of NH_3 shows a summer maximum at both sites, due to the presence of migratory seabirds and the influence of tundra and wildfire NH_3 emissions. However, the relative contributions of these sources have not yet been quantified on a multiyear time scale.

For the long-lived species CO, HCN, and C_2H_6 , the seasonal amplitudes are similar between Eureka and Thule as these species are well mixed in the troposphere. For NH_3 , the mean total columns in the peak summer months are over a factor of 2 greater at Thule than Eureka. Due to the short lifetime of gas-phase NH_3 , on the order of a few hours to a day, the transport of NH_3 surface emissions to the free troposphere is limited (Adams et al., 2019). The Eureka FTIR is located at an altitude of 610 m a.s.l. (above sea level), whereas the Thule instrument is at 210 m a.s.l., and consequently the Thule measurements are more sensitive to the influence of surface emissions. Additionally, large seabird colonies have been identified on the western coast of Greenland (Circumpolar seabird data portal, 2018) and would likely contribute to the greater NH_3 total columns at Thule.

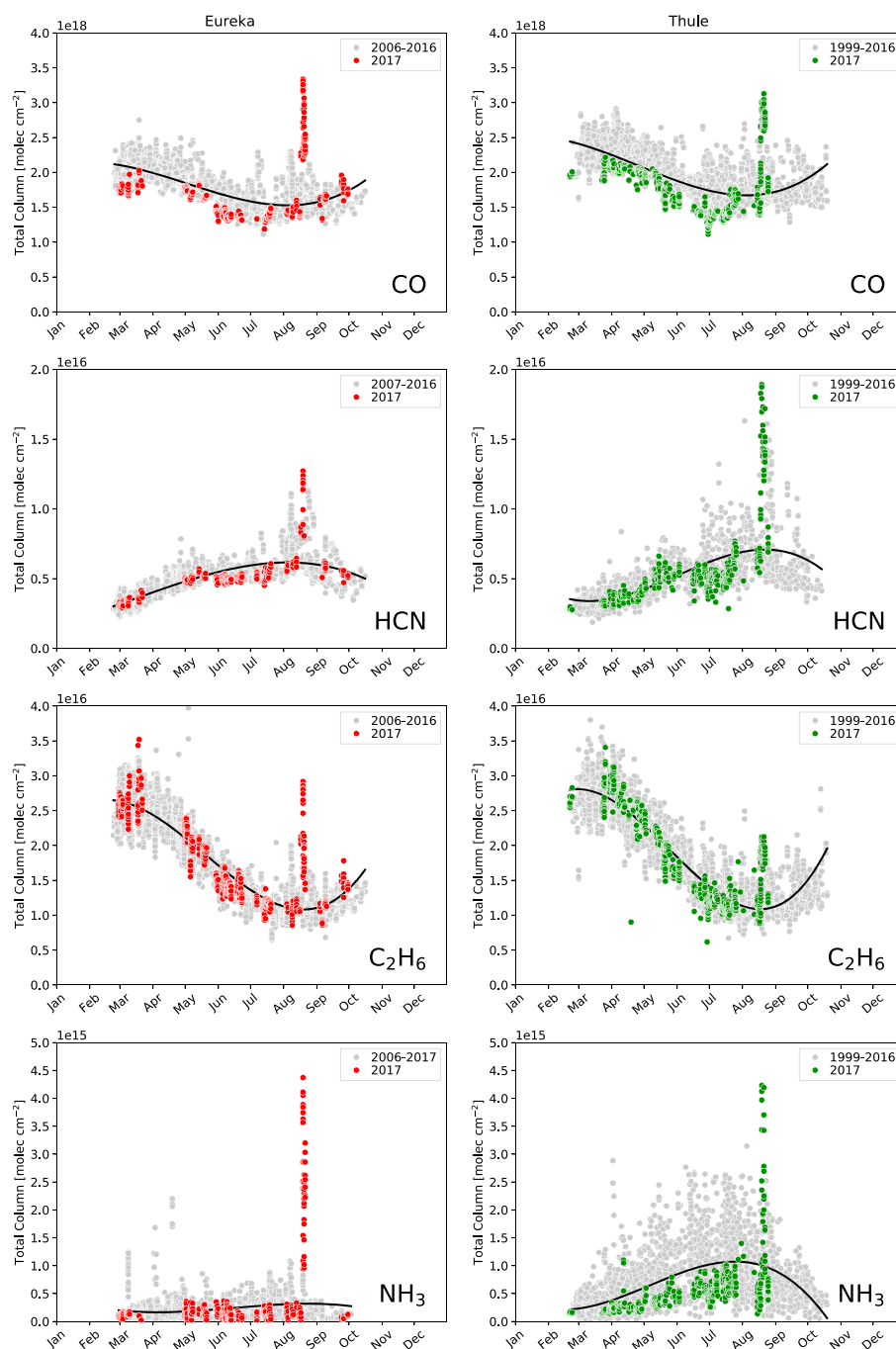


Figure 1. Total column time series of CO, HCN, C₂H₆, and NH₃ for Eureka (left, 2006–2017) and Thule (right, 1999–2017). The colored points indicate measurements taken in 2017, while all other years are shown in gray. The black line indicates a third-order polynomial fit to all measurements.

For CO, HCN, and C₂H₆, enhanced total columns are periodically observed in May to September at Eureka and Thule, corresponding to the boreal wildfire season (Macias Fauria & Johnson, 2008). While enhancements of CO, HCN, and C₂H₆ due to wildfire emissions are observed annually (Viatte et al., 2015), the only clear case of NH₃ wildfire emissions in the past was observed in July to August 2014 from the NWT wildfires (Lutsch et al., 2016). However, this does not take into account possible missed events due to instrument downtime or nonclear sky conditions preventing measurements. For all species, the 2017 Canadian wildfires in BC and NWT resulted in the greatest observed enhancements in the entire time series at both sites,

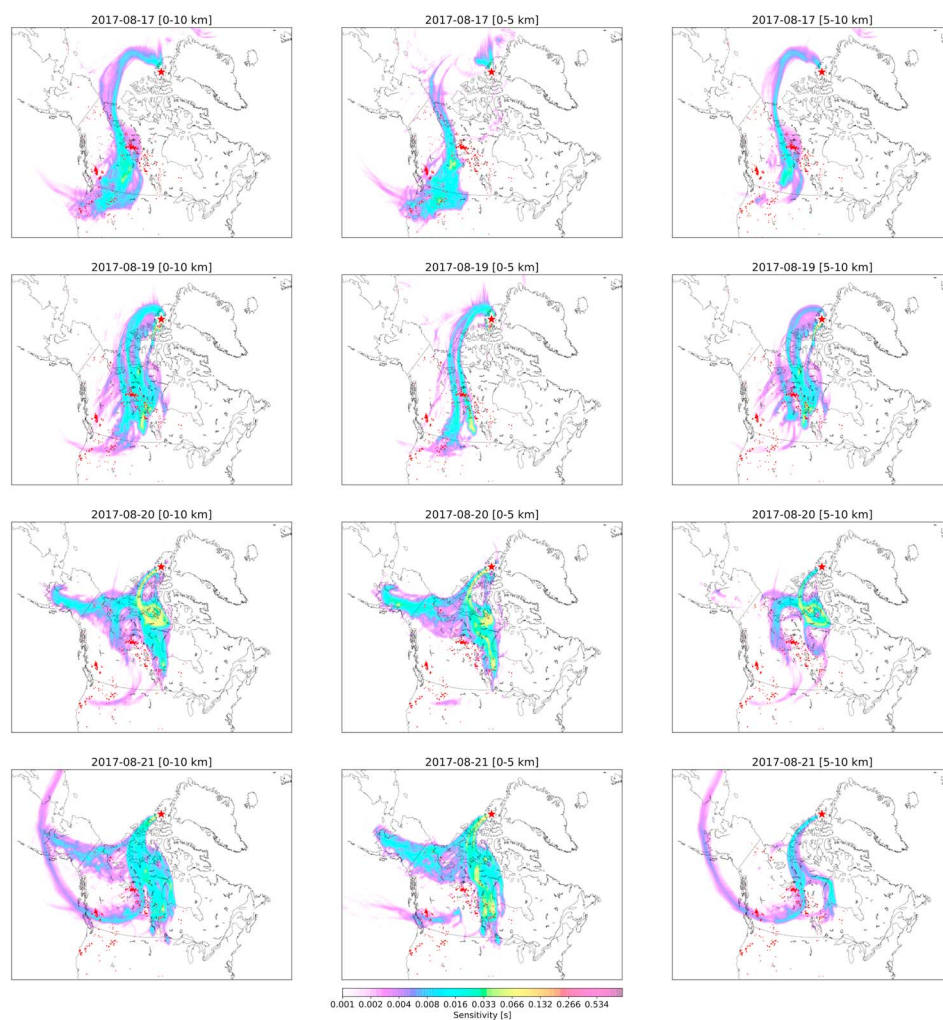


Figure 2. Typical FLEXPART sensitivity for Eureka on measurement days 17 and 19–21 August 2017. The red areas indicate MODIS burned areas for 7 days prior to the time of measurement, and the location of the Eureka FTIR site is indicated by the red star. The vertical region (left column: 0–10 km, middle column: 0–5 km, and right column: 5–10 km) indicates the vertical range of released particles. FLEXPART = FLEXible PARTICle; MODIS = Moderate Resolution Imaging Spectroradiometer; FTIR = Fourier transform infrared.

which were on the order of 2–5 times their ambient concentrations. The source attributions for the observed enhancements due to wildfire emissions for the 2017 events are discussed in the following section.

3.2. Source Attributions

The FLEXPART (FLEXible PARTICle; Stohl et al., 2005) Lagrangian transport model is used to diagnose the sensitivity of FTIR measurements to source regions. Enhancements in the measured FTIR total columns are first identified by CO total columns that are a standard deviation of 1σ greater than the monthly mean taken over all years of measurements at each site. The identified CO enhancements then define the period of fire-affected measurements. The days of fire-affected measurements are 17 and 19–21 August for Eureka and 19–22 August for Thule. It should be noted that nonclear sky conditions prevented measurements on 18 August at Eureka and for several days after 21 and 22 August for Eureka and Thule, respectively. For each day of fire-affected measurements, FLEXPART simulations are initialized for every hour of measurements. The FLEXPART model was run backward in time for 7 days, driven by meteorological data from the NCEP Climate Forecast System (CFS V2) 6-hr product (Saha et al., 2011). For each FLEXPART simulation, an ensemble of 1,000 air tracer particles was released over a 1-hr period from the surface to an altitude of 10 km. This vertical region was chosen in order to capture the likely altitude of the plume and the range where the CO total column measurement sensitivity is the greatest (total column averaging kernels at both sites show values near 1).

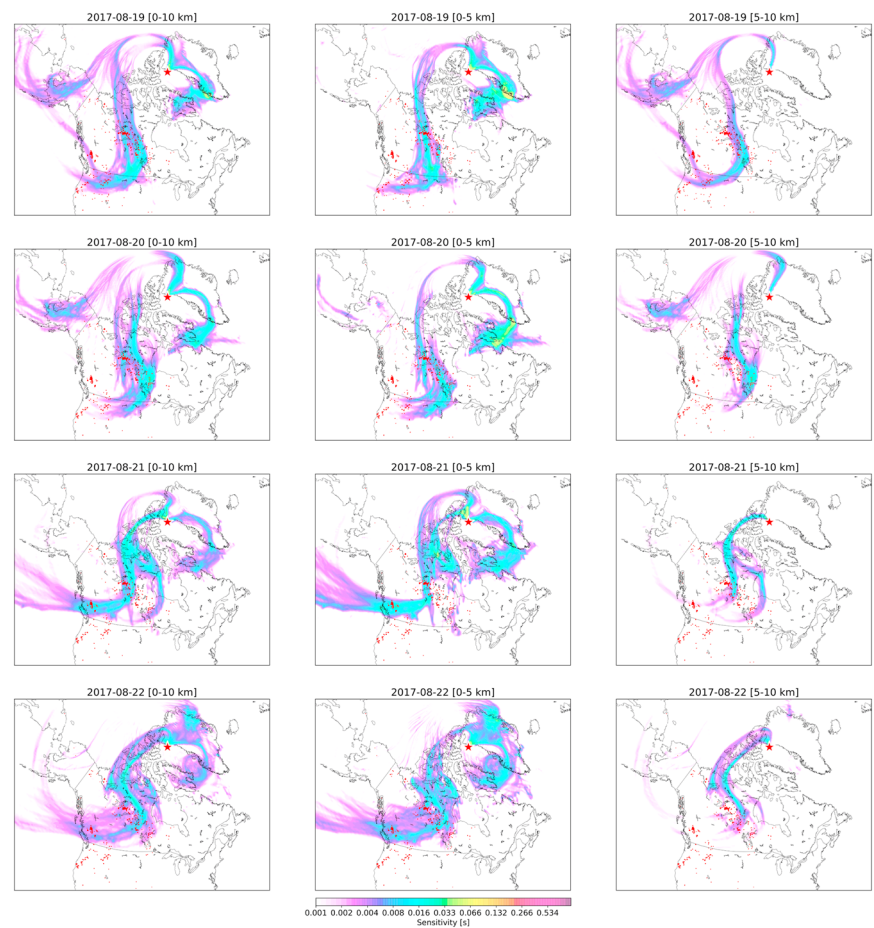


Figure 3. Same as Figure 2 but for Thule.

The sensitivity of the measurements to various source regions, or the so-called potential emission sensitivity (PES, in units of seconds), is proportional to the residence time of the air tracer particles. For each FLEXPART simulation, the sensitivity is considered for the following three cases:

- (a) Residence time at the surface for all particles released in the tropospheric column (0–10 km);
- (b) Residence time at the surface for particles released in the lower-tropospheric column (0–5 km);
- (c) Residence time at the 5-km altitude surface for particles released in the upper-tropospheric column (5–10 km).

For (a), all particles released from the 0- to 10-km altitude range are considered and the residence time at the surface is given, therefore, representing the tropospheric column sensitivity to surface emissions. For (b), the residence time at the surface of particles released in the lower troposphere (0–5 km) is considered and represents the lower-tropospheric column sensitivity to near-surface emissions. For (c), only particles released in the upper troposphere are considered (5–10 km) and the residence time at the 5-km altitude surface is therefore indicative of upper-tropospheric sensitivity to emissions injected into the upper troposphere. In all three cases, the residence time is reported at the lowest level of the altitude range considered.

The FLEXPART sensitivities are shown in Figures 2 and 3 for Eureka and Thule, respectively. The MODIS Burned Area Product Collection 6 (Giglio et al., 2018) provides the locations of burned areas during the 7-day period of the FLEXPART simulation. It is seen in Figures 2 and 3 that sensitivity to fire-affected regions is observed for all days of fire-affected FTIR measurements at both sites; however, the lower and upper tropospheric sensitivity differ between measurement days as discussed below.

For Eureka on 17 August, the first day of fire-affected measurements, the FLEXPART PES for 0–10 km shows greatest sensitivity to the wildfires in the NWT, a number of smaller fires in the northwestern United States, and little sensitivity to the BC wildfires. A similar spatial distribution of the sensitivity is observed for the

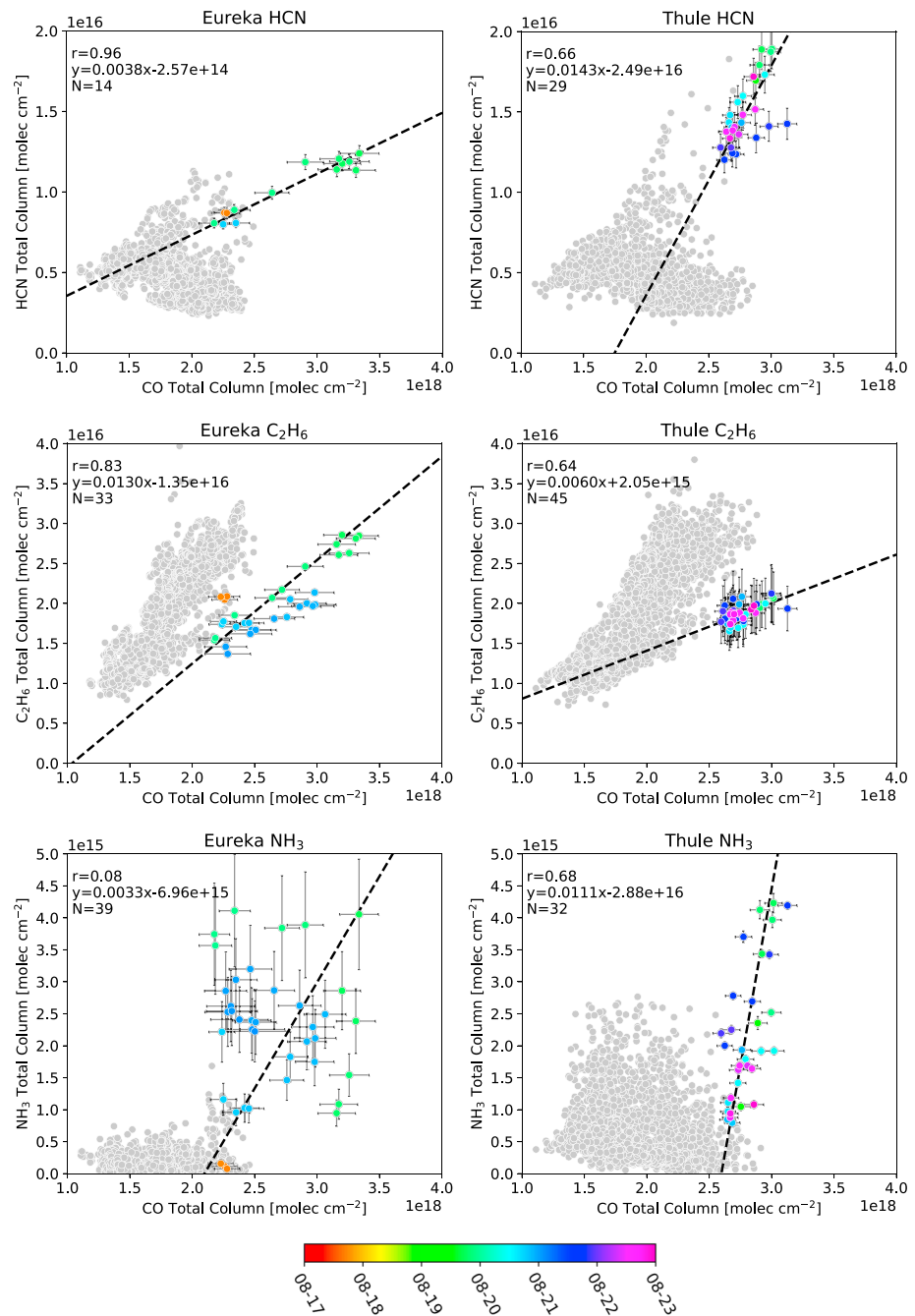


Figure 4. The enhancement ratios of HCN, C₂H₆, and NH₃ at (left column) Eureka and (right column) Thule. The gray points indicate all years of measurements. The colored points are measurements taken during the 2017 wildfire event where the colorbar represents the time of the measurement. The dashed black line is the linear regression to the colored points. The number of measurements (*N*), linear equation and correlation coefficient (*r*) are also shown.

0- to 5-km PES, although with no sensitivity to the NWT wildfires. For the 5- to 10-km PES, the sensitivity is primarily to the NWT wildfires, with no sensitivity to the other fire source regions. The difference in sensitivity between the 0- to 5- and 5- to 10-km PES implies that enhancements originating from the BC wildfires are due to near-surface emissions, and those from the NWT wildfires are due to upper tropospheric injection of emissions. On 19 August a similar spatial distribution is seen for the FLEXPART PES with some sensitivity observed to the BC wildfires and an eastward shift of the 0- to 5-km PES to the NWT fires. The FLEXPART PES on 17 and 19 August suggests that the main contribution to the enhanced total columns of

all species measured at Eureka is emissions from the NWT wildfires, with upper tropospheric injection of emissions being predominant on 17 August and near-surface emissions on 19 August. For 20 and 21 August at Eureka, the FLEXPART PES is similar for the 0- to 10-km range, indicating sensitivity to both the BC and NWT wildfires on those two days. The main sensitivity to the BC wildfires is observed for the 5- to 10-km PES with some sensitivity observed for the 0- to 5-km PES. For the NWT wildfires, the sensitivities for the 0- to 5- and 5- to 10-km PES are similar. On 20 and 21 August, the greater sensitivity to BC wildfires for the 5- to 10-km PES suggests that the enhancements observed at Eureka on these days are the likely result of the upper tropospheric injection of emissions from the BC wildfires.

For Thule, the FLEXPART PES of Figure 3 shows similar spatial patterns between the four fire-affected measurement days from 19–22 August. On 19 August, sensitivity to both the BC and NWT wildfires is observed, with the main contribution from the NWT fires occurring in the 0- to 5-km range and the main contribution from the BC wildfires occurring in the 5- to 10-km range. For 20 August, the NWT wildfires are the primary contributors to the measured total column enhancements for both the 0- to 5- and 5- to 10-km-altitude ranges. For 21 and 22 August, the spatial distribution of the FLEXPART PES is dominated by the 0- to 5-km range, indicating sensitivity to surface emissions from both the BC and NWT wildfires.

3.3. Trace Gas Correlations

Since the FTIR measurements are made at a distance from the fire source, the composition of the smoke plume is subject to physical and chemical processing. Therefore, we present the enhancement ratio of each species with respect to CO (Lefer et al., 1994) rather than the emission ratio (Akagi et al., 2011; Andreae & Merlet, 2001). The enhancement ratio is related to the emission ratio, which quantifies emissions near the source without the influence of plume aging. Since measurements are made downwind of the fire source, we did not account for background concentrations in the calculation of the enhancement ratio, or the so-called excess mixing ratio, since ambient concentrations of all species are likely to vary greatly over large spatial scales. The varying background conditions and influence of mixing may therefore cause uncertainty in interpreting the excess enhancement ratio (Yokelson et al., 2013).

To calculate the enhancement ratio for each species, enhanced CO measurements are identified by a standard deviation of 1σ from the monthly mean of all measurements, including enhancements, taken over all years, as described in section 3.2. The identified CO enhancements are then paired with the nearest measurement of the target species, HCN, C_2H_6 , and NH_3 recorded within 1 hr, with each CO measurement only used once for each species. A 1-hr window was chosen in order to maximize the number of pairs for each species while being sufficiently short to minimize the effects of plume aging. In most cases, the differences in measurement times between each species and the paired CO measurement were within 20 min. The enhancement ratio (in units of $\text{molec cm}^{-2}/\text{molec cm}^{-2}$) is then defined as the slope of the linear correlation of the target species relative to CO. The unified least squares procedure of York et al. (2004), which accounts for errors in both the ordinal and abscissa coordinates was used to determine a linear regression for the fire-affected measurements. The results of the linear regression are shown in Figure 4 and tabulated in Table 1.

For the case of a smoke plume originating from a single source, it is expected that a strong linear correlation would be found for the enhancement ratio, even over several days of measurements. This is particularly true for the long-lived species HCN and C_2H_6 , as the lifetimes of these species are much greater than the plume travel times, which are generally on the order of several days. A strong linear correlation was observed for the enhancement ratios of HCN and C_2H_6 at Eureka for the 2014 NWT fires (Lutsch et al., 2016) and similarly for the 2010 Russian fires (Viatte et al., 2013), with correlation coefficients (r) generally greater than 0.7 and in many cases greater than 0.85 for the enhancement ratios of HCN and C_2H_6 taken over one week or more of measurements. For the fire-affected measurements detected at Thule in August 2017, the linear correlations are generally weaker with correlation coefficients of 0.66 and 0.64 for HCN and C_2H_6 , respectively, taken over four days of fire-affected measurements. For Eureka, stronger correlations are observed, with correlation coefficients of 0.96 and 0.83 for HCN and C_2H_6 , respectively. The strong correlations at Eureka are the result of measurements predominately being recorded over one day for HCN and 2 days for C_2H_6 , therefore minimizing the influence of the variability of emissions and transport between days. For the enhancement ratios of NH_3 , the Eureka data show a much weaker linear correlation than Thule, with correlation coefficients of 0.08 and 0.68, respectively.

Table 1
Enhancement Ratios ($\text{molec cm}^{-2}/\text{molec cm}^{-2}$) for HCN, C_2H_6 , and NH_3 with Respect to CO for Each Day of Fire-Affected Measurements at Eureka and Thule

Site	Date	HCN			C_2H_6			NH_3		
		N	r	EnhR	N	r	EnhR	N	r	EnhR
Eureka	2017-8-17	2	—	—	3	—	—	3	—	—
	2017-8-19	10	0.95	0.0036 (0.0005)	12	0.99	0.0113 (0.0011)	11	-0.56	-0.0033 (0.0004)
	2017-8-20	2	—	—	15	0.85	0.0059 (0.0009)	16	0.46	0.0018 (0.0004)
	2017-8-21	0	—	—	3	—	—	9	-0.25	-0.0021 (0.0019)
	All	14	0.96	0.0038 (0.0004)	33	0.83	0.0130 (0.0011)	39	0.08	0.0033 (0.0003)
Thule	2017-8-19	5	0.80	0.0147 (0.0147)	6	0.95	0.0133 (0.0253)	7	0.75	0.0164 (0.0069)
	2017-8-20	8	0.89	0.0122 (0.0054)	14	0.54	0.0084 (0.0100)	9	0.81	0.0048 (0.0012)
	2017-8-21	8	0.92	0.0041 (0.0019)	12	0.46	0.0033 (0.0054)	8	0.86	0.0050 (0.0009)
	2017-8-22	8	0.82	0.0137 (0.0063)	13	0.64	0.0051 (0.0109)	8	0.49	0.0080 (0.0047)
	All	29	0.66	0.0143 (0.0022)	45	0.64	0.0060 (0.0035)	32	0.68	0.0111 (0.0014)

Note. Enhancement ratios were only calculated for days with five or more measurements for the species of interest. For each enhancement ratio calculated, the number of measurements (N) and the linear correlation coefficient (r) are also shown. The uncertainty on the calculated enhancement ratios are indicated by the value in parentheses, given by the standard error of the slope of the linear regression. Dates are formatted as year-month-day.

To examine the influence of the variability of the source sensitivities presented in section 3.2 on the measured concentration of each species, the enhancement ratio of each species was calculated for each day of fire-affected measurements. The results are summarized in Table 1. The enhancement ratio for each species was calculated only if the number of measurements paired with CO for that day was greater than or equal to 5. For Eureka, on 17 August, nonclear sky conditions limited the number of measurements, and therefore no enhancement ratios were calculated. On 19 August, the strongest linear correlations are found for HCN and C_2H_6 at Eureka, with correlation coefficients of 0.95 and 0.99, and a negative correlation for the NH_3 enhancement ratio with a correlation coefficient of -0.56 . A negative correlation has no physical interpretation but is due to the large variability of NH_3 as a result of transformation to its particulate phase or loss by deposition. Enhancement ratios of 0.0036 ± 0.0005 and 0.0113 ± 0.0011 for HCN and C_2H_6 , respectively, were found to be in agreement with the enhancement ratios for the 2014 NWT wildfires of 0.0037 ± 0.0005 for HCN and 0.0126 ± 0.0005 for C_2H_6 reported by Lutsch et al. (2016). The agreement is consistent with the FLEXPART PES of Figure 2 on 19 August, which shows sensitivity to wildfires in a nearby region to those identified by Lutsch et al. (2016), suggesting that a similar vegetation type was burned. The forest type for both the 2014 and 2017 NWT fires were classified as the Taiga Shield (Natural Resources Canada, 2013), consisting mainly of spruce trees, mosses, and lichens (Beaudoin et al., 2014). In contrast, the 2017 BC wildfires corresponded to the Montane Cordillera forest type, which primarily consists of Douglas fir and Ponderosa pine (Ireland & Petropoulos, 2015). On 20 August, measurements of HCN were limited at Eureka and no HCN enhancement ratio could be determined. A strong linear correlation was found for the C_2H_6 enhancement ratio ($r = 0.85$) and a moderate linear correlation was found for NH_3 ($r = 0.46$). For 21 August, no enhancement ratios were found for HCN or C_2H_6 , and the absence of a positive correlation was found for NH_3 with a poor correlation ($r = -0.25$).

The HCN enhancement ratios at Thule show strong linear correlations on individual measurement days, with correlation coefficients exceeding 0.80. The enhancement ratios of HCN at Thule are greater than at Eureka, likely due to the greater sensitivity to the BC wildfires (as shown in Figure 3) as compared to Eureka (shown in Figure 2). Additionally, for Thule, the sensitivities of both the BC and NWT wildfires tend to correspond to near surface emissions, commonly associated with smoldering combustion, which tends to have greater HCN emissions, while emissions of C_2H_6 tend to have less dependence on burning phase as it is produced in both smoldering and flaming combustion processes (Burling et al., 2010; Burling et al., 2011). For Thule, lower correlation coefficients of C_2H_6 enhancement ratios in comparison to those for HCN could also indicate that smoldering combustion was the dominant source of the smoke plume. However, the plume is likely a combination of both smoldering and flaming components that cannot be distinguished by ground-based measurements of this kind.

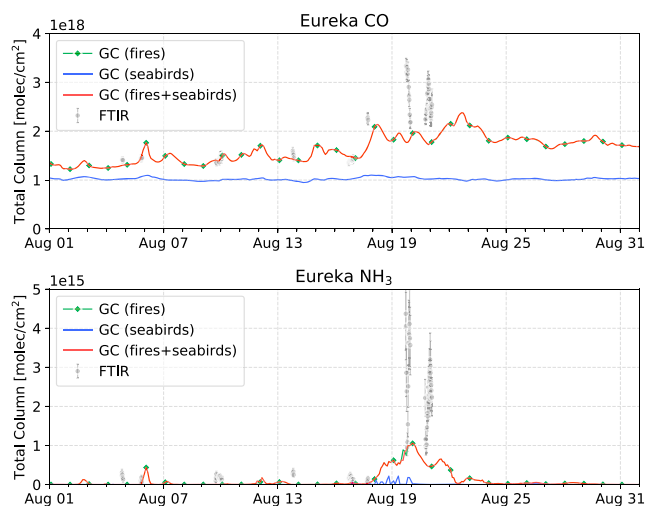


Figure 5. Smoothed GEOS-Chem CO (top) and NH₃ (bottom) total column time series for the month of August 2017. The three GOES-Chem simulation scenarios are shown: fire emission only (green), seabird colony NH₃ emissions only (blue), and fire and seabird emissions (red). FTIR measurements are shown in gray along with their associated measurement uncertainties represented by the error bars. GC = GEOS-Chem; FTIR = Fourier transform infrared.

For Thule, the enhancement ratios of NH₃ illustrate positive linear correlations on all days; however, the enhancement ratios are quite variable between measurement days, with the greatest values (0.0164 ± 0.0069) measured on 19 August and a minimum (0.0048 ± 0.0012) on 20 August. The variability of NH₃ enhancement ratios at Thule is partly due to the differences in emission sensitivity between days, as shown in Figure 3, as is the variability of the enhancement ratios for longer-lived HCN and C₂H₆. In addition, it was shown here that the enhancement ratios of HCN and NH₃ at Thule were on average greater than those at Eureka. For the days of fire-affected measurements at both sites, Eureka was more predominantly influenced by emissions from the NWT wildfires, whereas for Thule, greater sensitivity was observed to the BC wildfires. It has been found that for wildfires in the northern Rocky Mountains of the United States, a region of similar vegetation type to BC, the modified combustion efficiency (MCE) of these fires was substantially lower than for other coniferous dominated forest types (Urbanski, 2013). Lower MCE implies incomplete smoldering combustion (Ward & Hardy, 1991), resulting in greater emissions of reduced nitrogen compounds such as HCN and NH₃ (Burling et al., 2011). The MCE of a fire is dependent on the vegetation type, fuel load, moisture, and meteorology, and therefore can be highly variable. Pyrolysis temperature has also been found to have a strong influence on the emissions of HCN and NH₃, with greater emissions of HCN and NH₃ for high-temperature and low-temperature pyrolysis, respectively (Sekimoto et al., 2018). It should be noted that Sekimoto et al. (2018)

also found that low-temperature and high-temperature pyrolysis may not exactly correspond to smoldering and flaming combustion, respectively, while the pyrolysis-temperature emission profiles are similar between fuel types. The fuel types studied were representative of Western United States ecosystems consisting mainly of pine and fir species. However, the combination of larger HCN and NH₃ enhancements at Thule in comparison to Eureka and the greater surface sensitivity of Thule measurements to the BC wildfires suggests that smoldering phase emissions from these fires were the dominant contribution to the measured enhancements at Thule. For Eureka, the comparison of the enhancement ratios in Table 1 and the FLEXPART sensitivities shown in Figure 3 suggest that the NWT wildfires were the main contribution for enhanced total columns of all species at Eureka for 17 and 19 August, and a combination of both the BC and NWT wildfires for 20 and 21 August.

3.4. GEOS-Chem Comparison to FTIR Measurements

Due to the scarcity of NH₃ measurements at the low concentrations often observed in the high Arctic, it is difficult to investigate the large-scale influence of wildfire NH₃. The use of a chemical transport model allows for transported emissions of NH₃ to the Arctic to be simulated and therefore provides estimates of both surface and total column NH₃ concentrations in high Arctic regions. The FTIR measurements provide a means of evaluating model performance in the high Arctic. As a result of its long lifetime, atmospheric CO acts a tracer of wildfire emissions and is relatively unaffected by chemical aging during transport over several days. It therefore serves as a diagnostic of model transport and emissions. In contrast, NH₃ is short-lived and is strongly influenced by model chemistry and removal processes.

Comparisons of GEOS-Chem modeled CO and NH₃ total columns were performed as follows. The GEOS-Chem partial column profiles were smoothed by the FTIR total column averaging kernel following Rodgers and Connor (2003). Smoothing removes biases due to the a priori profile and limited vertical sensitivity of the FTIR measurement and therefore allows the model total column to be treated as if it were measured using the ground-based FTIR instrument. For each GEOS-Chem model output time step, the profile was interpolated onto the FTIR vertical grid and smoothed by the mean FTIR total column averaging kernel taken over all years for the respective species and site. The smoothed GEOS-Chem total columns for CO and NH₃ at Eureka and Thule are shown in Figures 5 and 6, respectively, for the following three simulation scenarios:

1. Fire emissions only;
2. Seabird colony NH₃ emissions only;
3. Fire and seabird colony NH₃ emissions.

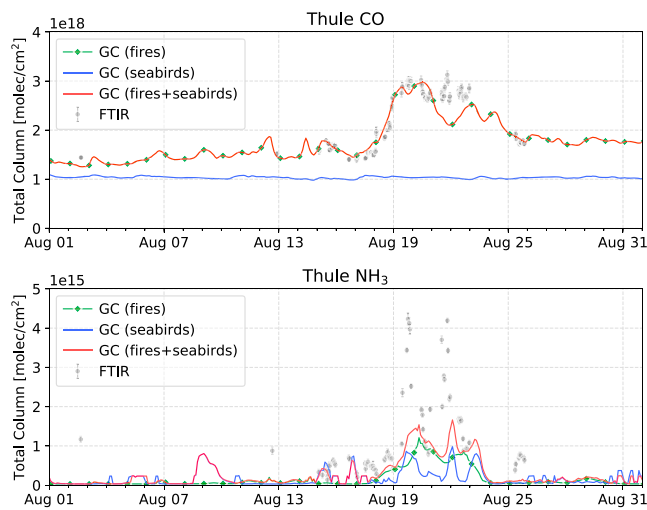


Figure 6. Same as Figure 5 but for Thule.

wildfire enhancements of CO observed in the FTIR measurements are well captured by the model. As discussed in section 3.2, Thule was predominantly influenced by the smoke plume originating from the BC wildfires, whereas these wildfires had a lesser influence at Eureka. As will be shown in the following sections, the GFAS emissions appear to underestimate the magnitude of the NWT plume in comparison to IASI measurements, resulting in the underestimation of modeled CO emissions at Eureka. In contrast, the plume originating from the BC wildfires shows better agreement between the model and IASI CO measurements.

The GEOS-Chem NH₃ simulation shows an underestimation in all three cases in comparison to the FTIR measurements at both sites. For Eureka, as shown in Figure 5, there is little influence of seabird colony NH₃ emissions and therefore, GEOS-Chem NH₃ at Eureka is attributed to transported wildfire emissions. For Thule, as shown in Figure 6, the inclusion of both seabird colony NH₃ and fire emissions most accurately represents the FTIR measurements. It is also seen in Figure 6 that the influence of seabird colony and wildfire NH₃ appear to occur simultaneously; this is a result of the wind direction, transporting both seabird colony and wildfire emissions from north of Thule.

It has been found that the bidirectional exchange of NH₃ may extend the spatial influence of NH₃ emission sources (Whaley et al., 2018; Zhu et al., 2015), but it was not included in the model simulations here. However, the locations of the seabird colonies are primarily in coastal regions, while the Arctic Ocean has been found to be a net sink of NH₃ (Wentworth et al., 2016). Therefore, transport of NH₃ from the larger seabird colonies of the Greenland coast to the Canadian Archipelago by bidirectional exchange processes would be limited. The influence of bidirectional exchange could likely be of greater importance for wildfire NH₃. The deposition of wildfire NH₃ would increase the nitrogen pool at the surface, which may be released at later times, thereby prolonging the influence of wildfire NH₃. Tundra emissions may also be a considerable source of NH₃ in the Arctic that was not considered here. Croft et al. (2019) reported an upper estimate of tundra NH₃ emissions, which was approximately 1.5 times the contribution from seabird colony sources in the Canadian Archipelago. Tundra NH₃ emissions are highly uncertain, but inclusion of these emissions in model simulations would likely increase background NH₃ concentrations as found by Croft et al. (2019).

The general underestimation of both CO and NH₃ at Eureka and Thule, and the inability of the model to capture the magnitude of the enhancements, is possibly a result of underestimation of modeled emissions and the assumption of boundary layer emissions in GEOS-Chem that may not be realistic in many cases as wildfire emissions are often injected into the free troposphere (Turquety et al., 2007; Val Martin et al., 2010). Once injected into the free troposphere, transport patterns are likely to differ from the surface. Additionally, the loss processes may differ in the free troposphere, particularly for the short-lived NH₃, which is subject to loss by wet and dry deposition and chemical loss. Injection of NH₃ into the free troposphere could favor the production of NH₄⁺ due to colder temperatures, and with sufficient quantities of nitrate (NO₃⁻) and nitric

Scenario (1) allows for the contributions of wildfire emissions of NH₃ to the Arctic to be examined, whereas scenario (2) is representative of background concentrations of NH₃ in the Arctic due to the persistent influence of local seabird colony NH₃ emissions during the warm season. The inclusion of both seabird colony and fire emissions of scenario (3) represents the real-world case. It should be noted that inclusion of seabird colony NH₃ emissions in the model has little influence on CO, as seen in Figures 5 and 6. Gas-phase NH₃ concentrations do not necessarily respond linearly to additional NH₃ emissions. This nonlinearity is because NH₃ partitioning to particulate NH₄⁺ is sensitive to the amount of available sulfuric and nitric acid. The overall increase in CO in scenarios (1) and (3) in comparison to the seabird-only case of scenario (2) illustrates the influence of accumulation of CO in the atmosphere from global biomass burning sources.

In Figure 5, there is some indication that GEOS-Chem captures the influence of the wildfire emissions during the period of enhanced FTIR measurements from 17–21 August, although the model is underestimating the magnitude of the enhancements. For Thule, as shown in Figure 6,

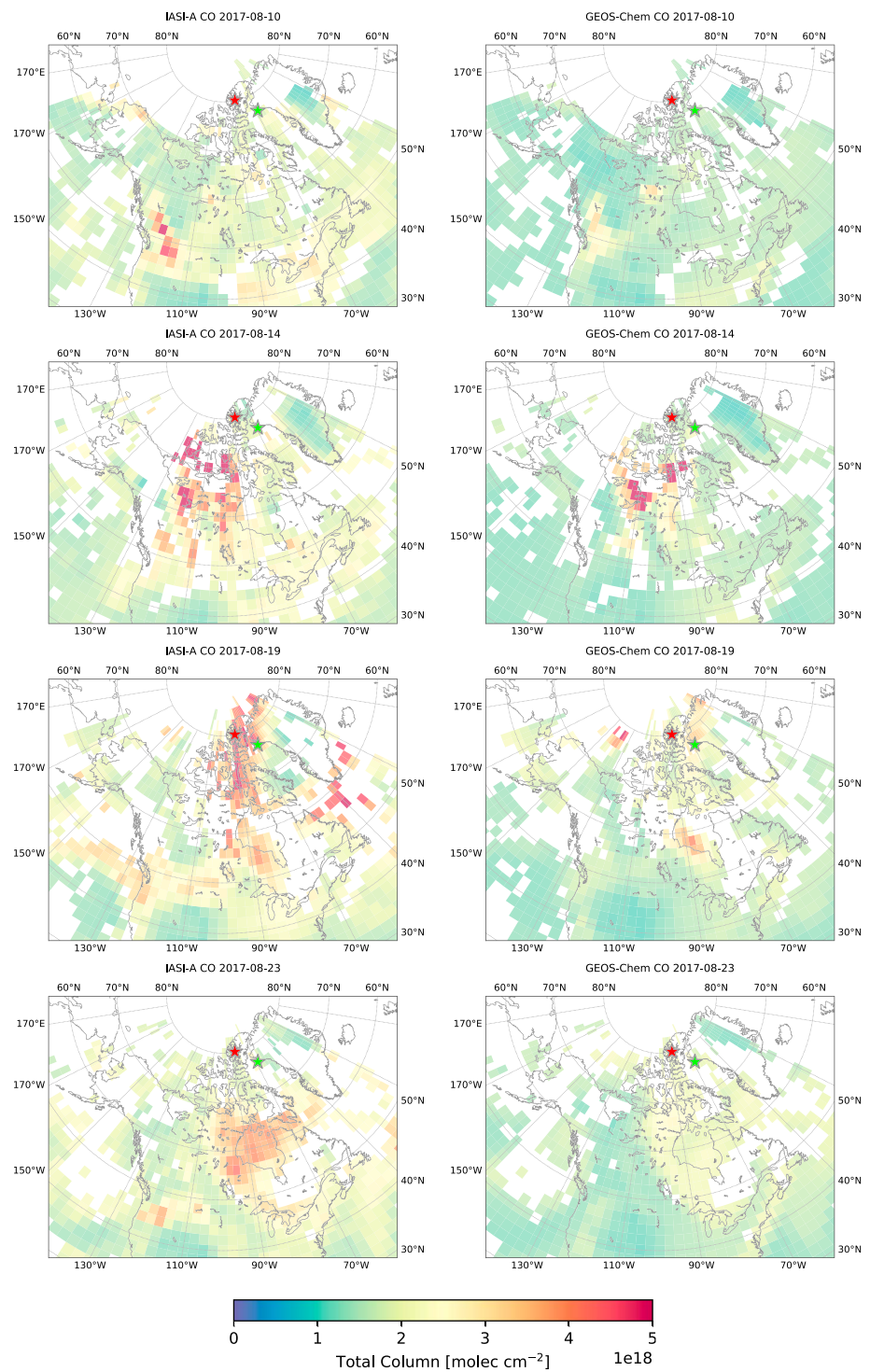


Figure 7. (left column) IASI CO measurements gridded onto the $2^\circ \times 2.5^\circ$ GEOS-Chem grid and daily averaged. (right column) GEOS-Chem CO total columns interpolated to match IASI overpass time and daily averaged. Locations of the FTIR sites Eureka and Thule are shown by the red and green stars, respectively. IASI = Infrared Atmospheric Sounding Interferometer; FTIR = Fourier transform infrared.

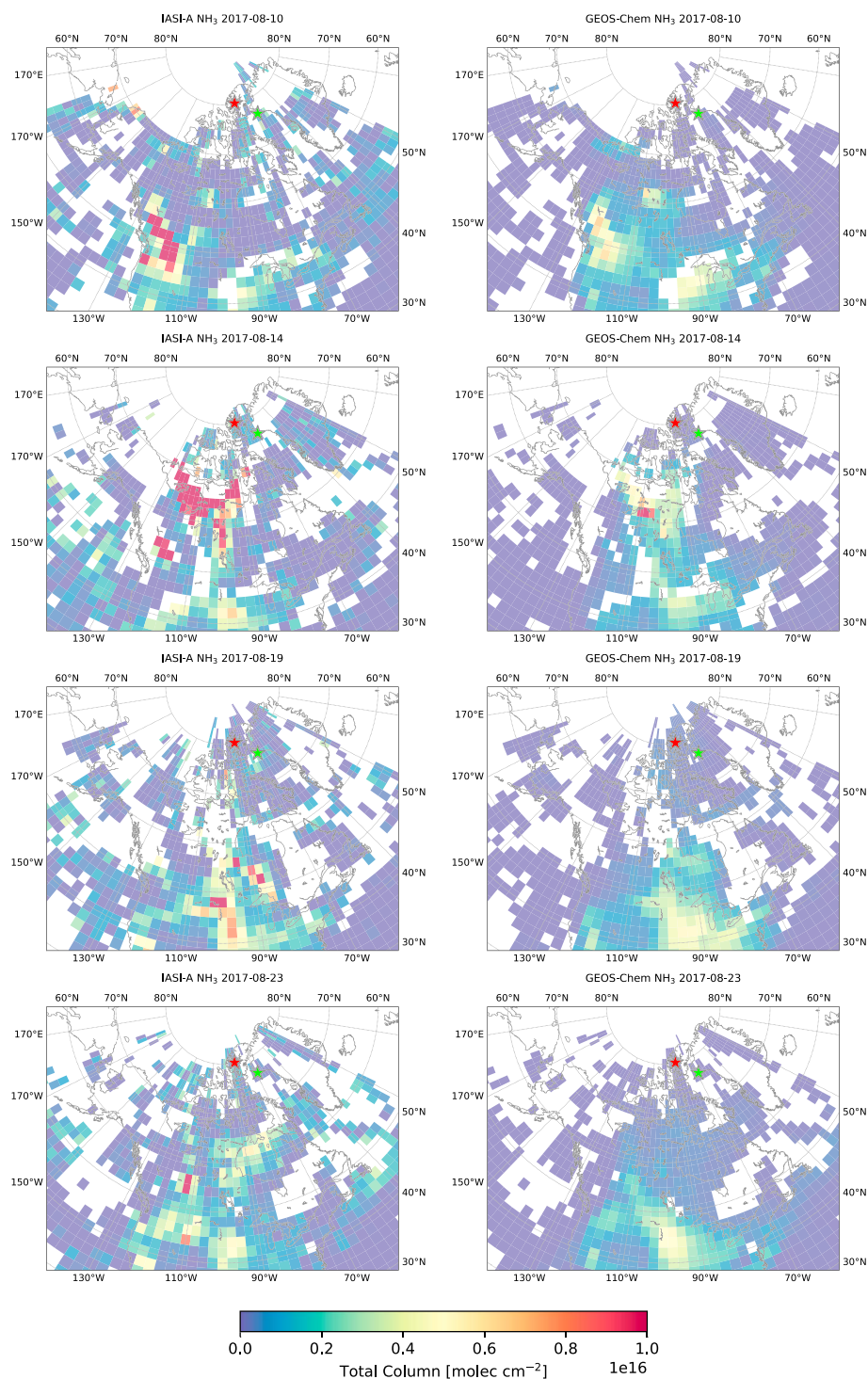


Figure 8. Same as Figure 7 but for NH_3 .

acid (HNO_3), could also promote the production of ammonium nitrate (NH_4HNO_3). Loss of NH_3 by dry and wet deposition may also be limited due to the reduced turbulent diffusion and drier conditions of the free troposphere. Global models also tend to suffer from numerical diffusion errors (Eastham & Jacob, 2017; Rastigejev et al., 2010) as a result of the coarse vertical and horizontal resolution. Underestimation of plume transport due to numerical diffusion is, however, likely to be of minor importance in comparison to the lack of realistic wildfire emission injection heights.

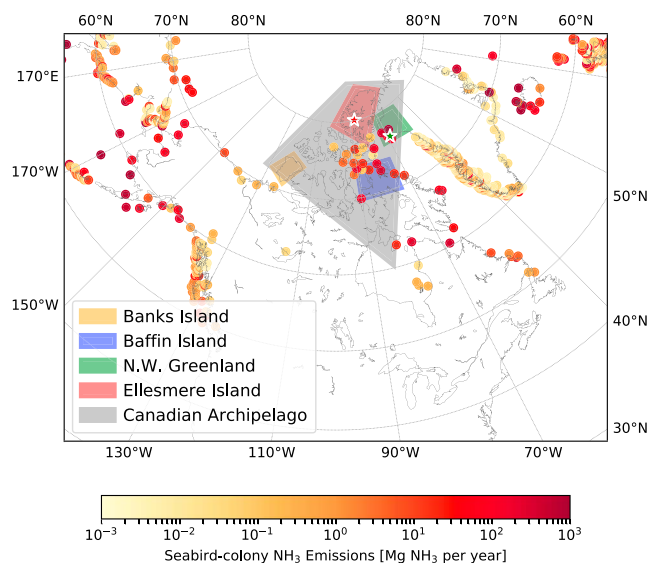


Figure 9. Locations chosen to evaluate spatially averaged GEOS-Chem NH_3 and NH_4^+ surface concentrations and total columns for the fire-affected period from 15–23 August 2017. Seabird colony NH_3 emissions in GEOS-Chem are also shown.

3.5. GEOS-Chem Comparison to IASI Measurements

To investigate the influence of possible transport and emissions errors in the GEOS-Chem model, a qualitative comparison of GEOS-Chem CO and NH_3 from the fire-only scenario to measurements from IASI was performed over the entire domain of Canada and Greenland, from 40°N , 180°W to 86°N , 15°E . For comparison of the GEOS-Chem model to IASI observations, the following method was applied. For each day, the GEOS-Chem model output was interpolated to the overpass time and locations of each individual IASI observation, and IASI observations were subsequently gridded to the GEOS-Chem horizontal grid. The interpolated GEOS-Chem total columns and gridded IASI total columns were then daily averaged over the entire spatial domain. The results are shown in Figures 7 and 8 for CO and NH_3 , respectively, for selected days during the simulation period.

The results are shown for four days from 10 August 2017, corresponding to the initial start of the fire plume to 23 August 2017, when measurements at Eureka and Thule both return to ambient levels. On 10 August, the initial plume is observed to have originated in BC in the lower left of the domain of Figure 7. From the comparison of IASI to GEOS-Chem CO in Figure 7, it is seen that the spatial distribution of the fire source is well represented by the model using the GFAS emission inventory, although the model tends to underestimate the CO concentration within the plume. The second main wildfire source, located south of Great Slave Lake in the NWT, is also observed in both IASI and GEOS-Chem where good agreement is found between the two. On subsequent days, transport of both plumes poleward is observed. The general shape of the fire plume from the BC wildfires is well represented in GEOS-Chem, but is lower in magnitude than IASI. The NWT wildfires appear to be underestimated by the model in comparison to IASI on 14 August. Due to the influence of the BC wildfire plume, it is difficult to distinguish the individual contribution of the BC and NWT wildfires on the observed CO plume. Due to the vertical sensitivity of IASI, which tends to peak in the middle to upper troposphere (George et al., 2009; Hurtmans et al., 2012; Kerzenmacher et al., 2012), near-surface CO from the NWT wildfires may be underestimated by IASI observations on 14 August, while the transport of the plume from the BC wildfires becomes more visible as the plume is advected vertically. However, it is found that the CO plume originating from the NWT is generally of lower concentration in the model than IASI. This observation is consistent with the FLEXPART sensitivity analysis in section 3.2 and the comparisons of GEOS-Chem to the FTIR measurements in section 3.4. In section 3.2, it was shown that measurements at Eureka were more strongly influenced by the NWT wildfires than Thule, while a low bias in GEOS-Chem CO at Eureka, in comparison to the FTIR measurements, was shown in Figure 5, which is consistent with an underestimation of emissions from the NWT wildfires.

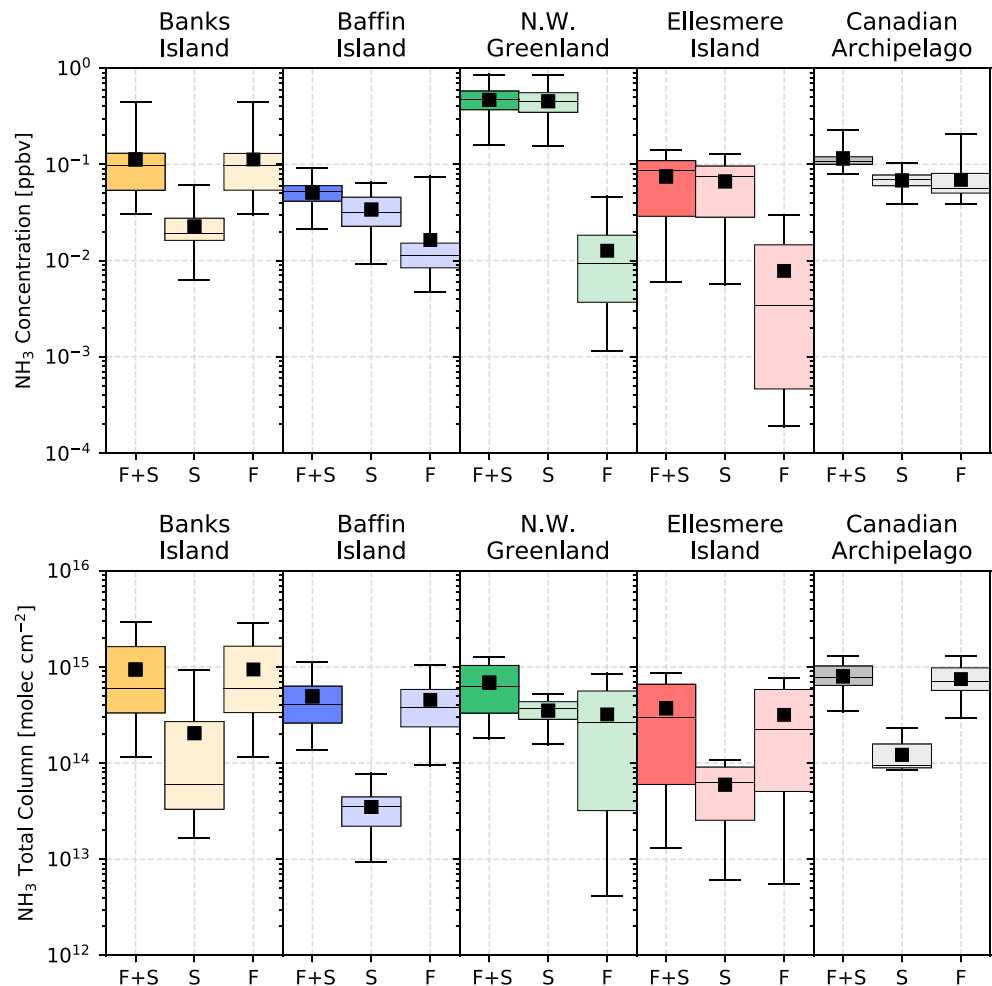


Figure 10. Box-and-whiskers plots of GEOS-Chem spatially averaged surface layer concentrations (top) and total columns (bottom) of NH_3 from 15–23 August 2017 for the fire and seabird emissions (F+S), seabird-only (S), and fire-only (F) simulations. The box indicates the quartile values of the data and the whiskers represent the range of the data. The black horizontal line indicates the median value and the black square is the mean of all data.

From 10 to 14 August in Figure 7, in both the IASI observations and model, the smoke plume originating in BC is transported poleward across central Canada and passes over the NWT wildfires. The combined plume is then transported north into the Canadian Arctic. On 17 August, the plume reaches Eureka, which is consistent with the FTIR enhancements. Transport of the plume toward Thule is also observed during this time, arriving at Thule on 19 August. From 20 August onward, the plume passes both Eureka and Thule before the CO total columns at both sites return to ambient levels on 24 August.

The emissions and transport of GEOS-Chem NH_3 show similar spatial and temporal variability to that of CO, as illustrated in Figure 8. While the transport of the NH_3 plume as observed by IASI is well represented in GEOS-Chem, the modeled NH_3 total columns are low in comparison to IASI. Within the plume, the GEOS-Chem NH_3 may be a factor of 2 lower than IASI. On 10 August, GEOS-Chem modeled NH_3 total columns are substantially less influenced by the BC wildfires. On subsequent days, the NH_3 plume dissipates rapidly, resulting in decreased total columns as the plume is transported poleward. On 14 August, the loss of NH_3 is particularly evident as the plume originating from the BC wildfires has reached the NWT wildfire source. As was observed in the IASI CO observations shown in Figure 7, the plume originating in BC is indistinguishable from the NWT wildfire source region. For NH_3 , IASI shows similar results with high values in the regions exceeding 5×10^{16} molec/cm². GEOS-Chem NH_3 does not exhibit these high values within the plume although the NWT wildfire source can clearly be distinguished from the plume in the

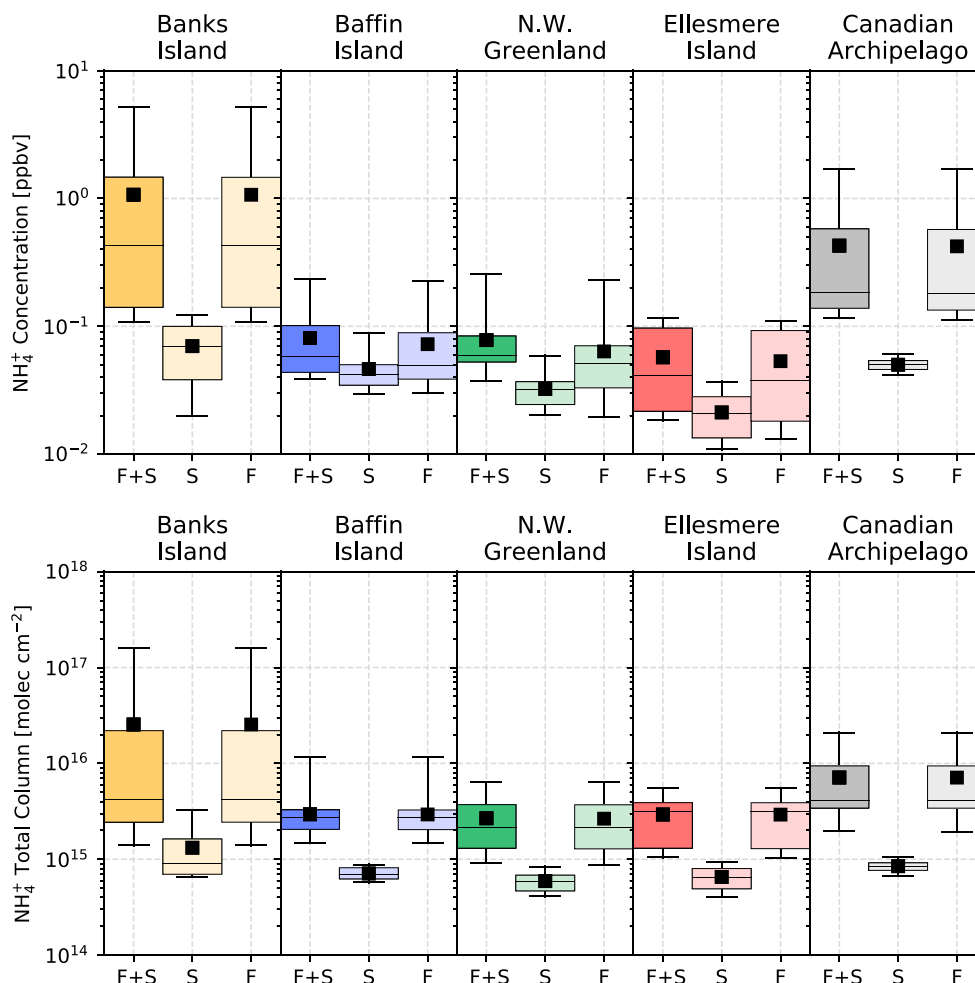


Figure 11. Same as Figure 10 but for NH_4^+ .

model. On following days, the modeled NH_3 plume dissipates as it is transported poleward, resulting in the underestimation of GEOS-Chem in comparison to the FTIR measurements, as was shown in section 3.4.

3.6. GEOS-Chem NH_3 and NH_4^+ in the Arctic

In this section, GEOS-Chem is used to estimate the influence of wildfire NH_3 on surface layer and total column concentrations in the Canadian Arctic. As was shown in sections 3.4 and 3.5, the model underestimates the transport of wildfire NH_3 in comparison to FTIR and IASI measurements. As such, we would expect modeled wildfire NH_3 in the Arctic to also have a low bias. We also consider the influence of NH_4^+ . The sum of NH_3 and NH_4^+ is defined as total ammonia ($\text{NH}_x \equiv \text{NH}_3 + \text{NH}_4^+$) and is an important contribution to reactive nitrogen. After emission, NH_3 will partition rapidly to an equilibrium state with NH_4^+ . The relative amounts of NH_3 and NH_4^+ are dependent on the availability of reactants and temperature. For wildfire emissions, it has been found that a large fraction (up to 30%) of emitted NH_3 is converted to NH_4^+ within 1.4 hr (Yokelson et al., 2009). Due to the longer lifetime of NH_4^+ , from several days to weeks, this would enable long-range transport and therefore NH_4^+ could be an important contributor to reactive nitrogen in the Arctic (Karlsson et al., 2013).

To examine the surface influence of wildfire NH_3 and NH_4^+ in the Arctic, the following five regions of interest were chosen: Banks Island, Baffin Island, Northwestern Greenland, Ellesmere Island, and the Canadian Archipelago (shown in Figure 9). These regions represent locations within the high Arctic, both with and without seabird colonies. Banks Island does not contain any considerable seabird colonies and is closest in proximity to the wildfire source regions. Ellesmere Island includes the Eureka FTIR site and is free of any considerable seabird colonies. While some seabird colonies are present in the Baffin Island region, it is also

Table 2
GEOS-Chem NH₃ and NH₄⁺ Surface Concentrations and Total Columns in the Regions Defined in Figure 9 Temporally Averaged From 15–23 August 2017

	NH ₃			NH ₄ ⁺		
	Fires+Seabirds	Seabirds	Fires	Fires+Seabirds	Seabirds	Fires
Surface concentration (ppbv)						
Banks Island	0.11 (0.08)	0.02 (0.01)	0.11 (0.08)	1.07 (1.37)	0.07 (0.03)	1.07 (1.37)
Baffin Island	0.05 (0.01)	0.03 (0.01)	0.02 (0.01)	0.08 (0.05)	0.05 (0.02)	0.07 (0.05)
Northwestern Greenland	0.47 (0.16)	0.45 (0.15)	0.01 (0.01)	0.08 (0.05)	0.03 (0.01)	0.06 (0.05)
Ellesmere Island	0.07 (0.04)	0.07 (0.04)	0.01 (0.01)	0.06 (0.04)	0.02 (0.01)	0.05 (0.04)
Canadian Archipelago	0.12 (0.03)	0.07 (0.01)	0.07 (0.03)	0.43 (0.45)	0.05 (0.00)	0.42 (0.45)
Total column (10 ¹⁵ molec /cm ²)						
Banks Island	0.94 (0.78)	0.20 (0.28)	0.94 (0.79)	25.44 (43.68)	1.32 (0.86)	25.44 (43.67)
Baffin Island	0.49 (0.28)	0.03 (0.02)	0.45 (0.27)	2.95 (1.48)	0.71 (0.10)	2.93 (1.48)
Northwestern Greenland	0.69 (0.37)	0.35 (0.10)	0.32 (0.29)	2.68 (1.63)	0.59 (0.12)	2.65 (1.64)
Ellesmere Island	0.37 (0.30)	0.06 (0.03)	0.32 (0.28)	2.93 (1.43)	0.65 (0.17)	2.92 (1.43)
Canadian Archipelago	0.80 (0.24)	0.12 (0.05)	0.75 (0.25)	7.12 (6.07)	0.85 (0.11)	7.11 (6.07)

Note. The values in parentheses indicate the standard deviation of the mean. Results are shown for the three GEOS-Chem simulations: fire and seabird NH₃ emissions, fire-only, and seabird-only emissions.

strongly influenced by seabird colony NH₃ emissions from Greenland. The Northwestern Greenland region was chosen to include the location of the Thule FTIR site and the seabird colonies of the northwestern coast of Greenland and therefore represents the extreme case of high seabird colony NH₃ emissions. Lastly, the Canadian Archipelago region is representative of the majority of the Canadian high Arctic and includes all of the aforementioned regions. For each of these five regions, the GEOS-Chem grid boxes that lie within the defined boundaries illustrated in Figure 9 are spatially averaged for each 2-hourly GEOS-Chem output timestep from 15–23 August 2017, which corresponds to the period of fire-affected FTIR measurements at Eureka and Thule. We consider the following three GEOS-Chem simulation scenarios: (1) fire emissions only, (2) seabird colony NH₃ emissions, (3) fire and seabird colony NH₃ emissions. The results are shown in Figures 10 and 11 and are tabulated in Table 2.

As shown in Figure 10 and Table 2, it is found that the greatest NH₃ surface concentrations occur in northwestern Greenland with a mean value of 0.47 ppbv in the GEOS-Chem simulation with fire and seabird colony NH₃ emissions. From the seabird-only simulation, it is evident that the seabird colony NH₃ emissions are the dominant contribution in this region, with a mean surface layer contribution of 0.45 ppbv. Ellesmere Island is also strongly influenced by seabird colony NH₃ emissions, with a mean surface concentration of 0.07 ppbv with a standard deviation of 0.04 ppbv from the seabird-only simulation and 0.01 ppbv with a standard deviation of 0.01 ppbv from the fire-only simulation. Baffin Island shows nearly equal contributions from the seabird colony (0.02 ppbv) and wildfires (0.03 ppbv). Banks Island is predominantly influenced by wildfire emissions, with a mean surface concentration from the fire-only simulation of 0.11 ppbv. These results are consistent with the loss of NH₃ in the plume by deposition in the low-latitude regions (Banks Island and Baffin Island), with decreasing surface layer influence in the high-latitude regions (Northwestern Greenland and Ellesmere Island). For the period of fire-affected measurements from 15–23 August, wildfire and seabird colony NH₃ had equal contributions 0.07 ppbv to surface layer NH₃ over the Canadian Archipelago. However, in all regions, total column concentrations of NH₃ were predominantly due to wildfire NH₃ emissions, with the exception of Northwestern Greenland. In this region, the seabird colony and wildfire NH₃ emissions had nearly equivalent contributions of 0.35 and 0.32 × 10¹⁵ molec/cm², respectively.

The surface layer concentrations of NH₄⁺ show somewhat similar results to those for NH₃, with greater influence of wildfire NH₄⁺ for the lower-latitude regions, as seen in Figure 11 and Table 2. In all regions, wildfire emissions are the dominate source of NH₄⁺, with mean surface layer concentrations ranging from 1.07 ppbv for Banks Island and to 0.05 ppbv at Ellesmere Island. Over the Canadian Archipelago, a mean concentration of 0.42 ppbv was found, with large variation (standard deviation of 0.45 ppbv) due to the spatial and temporal variability of the plume during the fire-affected period. As was the case for NH₃, total column amounts

of NH_4^+ are greatly enhanced due to the transport of wildfire emissions, with a mean wildfire contribution of 7.11×10^{15} molec/cm² in the Canadian Archipelago for the fire-only simulation.

It can therefore be concluded that the 2017 Canadian wildfires events had a significant influence on surface layer NH_3 and NH_4^+ . It should be noted that although these enhancements occurred over a short period of time, from 15–23 August, the enhanced surface layer concentrations of NH_3 were comparable to the persistent contribution of seabird colony NH_3 . For NH_4^+ , the influence of wildfire emissions greatly enhanced both surface layer and total column concentrations in all regions. As was discussed in sections 3.4 and 3.5, GEOS-Chem was shown to underestimate the transport of wildfire NH_3 and CO in comparison to FTIR and IASI observations. It is therefore likely that the modeled influence of transported wildfire emissions on Arctic NH_3 and NH_4^+ presented here is a low estimate.

The influence of wildfires on the reactive nitrogen budget of the high Arctic remains unknown. It has been shown that the Arctic has been undergoing an increase in the areal extent of vegetated land as a result of warming in the Arctic (Keenan & Riley, 2018; Myneni et al., 1997; Zhu et al., 2016) and increased community plant height has also been observed as a result of this warming (Bjorkman et al., 2018). Plant growth in the Arctic is strongly limited by nutrient availability (Schimel & Bennett, 2004), therefore the wildfire contribution to nitrogen deposition in the form of NH_x could potentially facilitate an increase in plant growth, especially if extreme wildfire events such as these become more frequent. The effects on nitrogen deposition in the Canadian high Arctic are still uncertain, although it has been noted that high Arctic tundra has a high capacity for nitrogen pollution retention, resulting in a major source of eutrophication (Choudhary et al., 2016), which could have a negative impact on biodiversity.

4. Conclusions

The 2017 BC and NWT wildfires resulted in the greatest observed enhancements of total column NH_3 , CO, HCN, and C_2H_6 in the decade-long time series of FTIR measurements at Eureka, Nunavut, and Thule, Greenland. The magnitude of these enhancements was on the order of 2–5 times the ambient levels of these species at each site. The observed NH_3 enhancements provide evidence for the importance of wildfires as an episodic source of NH_3 in the summertime Arctic, supporting the results of Lutsch et al. (2016). Enhancement ratios of NH_3 , HCN, and C_2H_6 were calculated with respect to CO at both sites. It was found that variations in these enhancement ratios were due to the influence of multiple fire plumes and variations in the burning phases of the fire events. Evidence for this was provided by FLEXPART sensitivities which showed the influence of the BC and NWT plumes, with the influence of each fire source varying between measurement days at both sites. Eureka FTIR measurements were found to be most sensitive to the NWT wildfires, and Thule FTIR measurements were most sensitive to the BC wildfires, based on the FLEXPART sensitivity analysis.

The GEOS-Chem CTM was used to simulate the transport of wildfire emissions to the Arctic. Through comparisons of GEOS-Chem NH_3 and CO to FTIR measurements at Eureka and Thule, and to observations from IASI, it was shown that the model underestimates the transport of wildfire emissions to the Arctic for the observations examined here. The underestimation of modeled NH_3 and CO in comparison to Eureka FTIR measurements suggests an underestimation of emissions from the NWT fire source. This is consistent with the model comparisons to IASI measurements, which suggested an underestimation of emissions from the NWT wildfires in the model. At Thule, GEOS-Chem showed good agreement to the FTIR measurements for CO, but underestimated the wildfire NH_3 contribution. The reason for the underestimation of GEOS-Chem CO and NH_3 in comparison to FTIR and IASI measurements remains unclear and is likely the result of several factors including the underestimation of emissions, plume height, chemistry, and transport errors in the model. These aspects of the model should be investigated in future studies. Increased temporal and spatial sampling of NH_3 and NH_4^+ by in situ, ground-based and satellite instruments is recommended to better evaluate model performance and infer wildfire impacts on NH_3 in the Arctic. Tundra NH_3 emissions should be examined in these studies. Furthermore, the use of satellite observations, such as the studies of Kharol et al. (2018) and Adams et al. (2019), should also be the focus of future work in order to quantify boreal wildfire NH_3 emissions and transport to the Arctic.

Simulated transport of wildfire emissions in GEOS-Chem was found to be a significant contributor to reactive nitrogen in the form of NH_3 and NH_4^+ in the high Arctic for the fire-affected period of 15–23 August 2017. GEOS-Chem surface layer NH_3 and NH_4^+ was shown to be greatly enhanced during the fire-affected

period. A mean wildfire contribution of 0.07 ppbv in the Canadian Archipelago from 15–23 August 2017 was equivalent to the local background from seabird-colony NH_3 sources. Surface layer NH_4^+ was greatly enhanced during this period as well, with a mean concentration of 0.42 ppbv in the Canadian Archipelago in the fire-only simulation, whereas for the seabird-only simulation a mean concentration of 0.05 ppbv was observed. Due to the influence of transported wildfire emissions, total column NH_3 and NH_4^+ were found to be enhanced several times greater than background levels associated with the seabird colony sources. The strong episodic influence of the BC and NWT wildfires on the reactive nitrogen budget of the high Arctic could have significant impacts on nutrient availability and eutrophication of this sensitive ecosystem.

Data Availability

The FTIR data are available from the NDACC data repository at <ftp://ftp.cpc.ncep.noaa.gov/ndacc/>. IASI is a joint mission of EUMETSAT and the Centre National d'Etudes Spatiales (CNES, France). The IASI Level-1C data are distributed in near real time by Eumetsat through the EumetCast system distribution. The authors acknowledge the AERIS data infrastructure (<https://www.aeris-data.fr/>) for providing access to the IASI Level-1C data and Level-2 CO and NH_3 data used in this study. We acknowledge the free use of IASI NH_3 data from the Atmospheric Spectroscopy Group at Université libre de Bruxelles (ULB). The MODIS MCD64A1 data product was retrieved from the online Data Pool, courtesy of the NASA Land Processes Distributed Active Archive Center (LP DAAC), USGS/Earth Resources Observation and Science (EROS) Center, Sioux Falls, South Dakota, https://lpdaac.usgs.gov/data_access/data_pool. The authors acknowledge the use of the GFASv1.2 emissions inventory, which contains modified Copernicus Atmosphere Monitoring Service information (2018). The GEOS data used in this study have been provided by the Global Modeling and Assimilation Office (GMAO) at NASA Goddard Space Flight Center. The authors also acknowledge the use of the FLEXPART Lagrangian dispersion model (<https://www.flexpart.eu/wiki/FpDownloads>) and the use of the Pflexible Python module (<https://bitbucket.org/jfburkhart/pflexible>) developed by John F. Burkhart, which was modified here to plot the FLEXPART sensitivities in this paper. The NCEP CFSv2 data used are listed in the references. Any additional data may be obtained from Erik Lutsch (elutsch@physics.utoronto.ca).

References

- Abbott, J. P. D., Leaitch, W. R., Aliabadi, A. A., Bertram, A. K., Blanchet, J.-P., Boivin-Rioux, A., et al. (2019). Overview paper: New insights into aerosol and climate in the Arctic. *Atmospheric Chemistry and Physics*, *19*(4), 2527–2560. <https://doi.org/10.5194/acp-19-2527-2019>
- Adams, C., McLinden, C. A., Shephard, M. W., Dickson, N., Damers, E., Chen, J., et al. (2019). Satellite-derived emissions of carbon monoxide, ammonia, and nitrogen dioxide from the 2016 Horse River wildfire in the Fort McMurray area. *Atmospheric Chemistry and Physics*, *19*(4), 2577–2599.
- Akagi, S. K., Yokelson, R. J., Wiedinmyer, C., Alvarado, M. J., Reid, J. S., Karl, T., et al. (2011). Emission factors for open and domestic biomass burning for use in atmospheric models. *Atmospheric Chemistry and Physics*, *11*(9), 4039–4072.
- Alvarado, M. J., Cady-Pereira, K. E., Xiao, Y., Millet, D. B., & Payne, V. H. (2011). Emission ratios for ammonia and formic acid and observations of peroxy acetyl nitrate (PAN) and ethylene in biomass burning smoke as seen by the Tropospheric Emission Spectrometer (TES). *Atmosphere*, *2*(4), 633–654.
- Andreae, M. O., & Merlet, P. (2001). Emission of trace gases and aerosols from biomass burning. *Global Biogeochemical Cycles*, *15*(4), 955–966. <https://agupubs.onlinelibrary.wiley.com/doi/pdf/10.1029/2000GB001382>
- Aneja, V. P., Bunton, B., Walker, J. T., & Malik, B. P. (2001). Measurement and analysis of atmospheric ammonia emissions from anaerobic lagoons. *Atmospheric Environment*, *35*(11), 1949–1958.
- Asman, Willem A. H., Sutton, M. A., & Schjorring, J. K. (1998). Ammonia: emission, atmospheric transport and deposition. *The New Phytologist*, *139*(1), 27–48. <https://nph.onlinelibrary.wiley.com/doi/pdf/10.1046/j.1469-8137.1998.00180.x>
- BC Wildfire Service (2017). Wildfire season summary—Province of British Columbia. <https://www2.gov.bc.ca/gov/content/safety/wildfire-status/about-bcws/wildfire-history/wildfire-season-summary>, Accessed: 2019-1-30
- Batchelor, R. L., Strong, K., Lindenmaier, R., Mittermeier, R. L., Fast, H., Drummond, J. R., & Fogal, P. F. (2009). A new Bruker IFS 125HR FTIR spectrometer for the polar environment atmospheric research laboratory at Eureka, Nunavut, Canada: Measurements and comparison with the existing Bomem DA8 spectrometer. *Journal of Atmospheric and Oceanic Technology*, *26*(7), 1328–1340. <https://doi.org/10.1175/2009JTECHA1215.1>
- Beaudoin, A., Bernier, P.-Y., Guindon, L., Villemaire, P., Guo, X. J., Stinson, G., et al. (2014). Mapping attributes of Canada's forests at moderate resolution through kNN and MODIS imagery.
- Bjorkman, A. D., Myers-Smith, I. H., Elmendorf, S. C., Normand, S., Rüger, N., Beck, P. S. A., et al. (2018). Plant functional trait change across a warming tundra biome. *Nature*, *562*, 57–62.
- Blackall, T. D., Wilson, L. J., Theobald, M. R., Milford, C., Nemitz, E., Bull, J., et al. (2007). Ammonia emissions from seabird colonies. *Geophysical Research Letters*, *34*, L10801. <https://doi.org/10.1029/2006GL028928>
- Bobbink, R., Hicks, K., Galloway, J., Spranger, T., Alkemade, R., Ashmore, M., et al. (2010). Global assessment of nitrogen deposition effects on terrestrial plant diversity: A synthesis. *Ecological applications: a publication of the Ecological Society of America*, *20*(1), 30–59.
- Bobbink, R., Hornung, M., & Roelofs, J. G. M. (1998). The effects of air-borne nitrogen pollutants on species diversity in natural and semi-natural European vegetation. *The Journal of Ecology*, *86*(5), 717–738.

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- Boulanger, Y., Gauthier, S., & Burton, P. J. (2014). A refinement of models projecting future Canadian fire regimes using homogeneous fire regime zones. *Canadian Journal of Forest Research. Journal Canadien de la Recherche Forestiere*, 44(4), 365–376.
- Bouwman, A. F., Boumans, L. J. M., & Batjes, N. H. (2002). Estimation of global NH₃ volatilization loss from synthetic fertilizers and animal manure applied to arable lands and grasslands: Ammonia emissions from fertilizers. *Global Biogeochemical Cycles*, 16(2), 1024. <https://doi.org/10.1029/2000GB001389>
- Bouwman, A. F., Lee, D. S., Asman, W. A. H., Dentener, F. J., Van Der Hoek, K. W., & Olivier, J. G. J. (1997). A global high-resolution emission inventory for ammonia. *Global Biogeochemical Cycles*, 11(4), 561–587. <https://doi.org/10.1029/97GB02266>
- Burling, I. R., Yokelson, R. J., Akagi, S. K., Urbanski, S. P., Wold, C. E., Griffith, D. W. T., et al. (2011). Airborne and ground-based measurements of the trace gases and particles emitted by prescribed fires in the United States. *Atmospheric Chemistry and Physics*, 11(23), 12,197–12,216.
- Burling, I. R., Yokelson, R. J., Griffith, D. W. T., Johnson, T. J., Veres, P., Roberts, J. M., et al. (2010). Laboratory measurements of trace gas emissions from biomass burning of fuel types from the southeastern and southwestern United States. *Atmospheric Chemistry and Physics*, 10(22), 11,115–11,130.
- Choudhary, S., Blaud, A., Osborn, A. M., Press, M. C., & Phoenix, G. K. (2016). Nitrogen accumulation and partitioning in a high Arctic tundra ecosystem from extreme atmospheric N deposition events. *The Science of the total environment*, 554–555, 303–310.
- Cicerone, R. J., & Zellner, R. (1983). The atmospheric chemistry of hydrogen cyanide (HCN). *Journal of Geophysical Research*, 88(C15), 10689.
- Circumpolar seabird data portal (2018). <http://axiom.seabirds.net/maps/js/seabirds.php>, Accessed: 2018-10-30
- Clarisse, L., Shephard, M. W., Dentener, F., Hurtmans, D., Cady-Pereira, K., Karagulian, F., et al. (2010). Satellite monitoring of ammonia: A case study of the San Joaquin Valley. *Journal of Geophysical Research*, 115, D13302. <https://doi.org/10.1029/2009JD013291>
- Clerbaux, C., Boynard, A., Clarisse, L., George, M., Hadji-Lazarou, J., Herbin, H., et al. (2009). Monitoring of atmospheric composition using the thermal infrared IASI/MetOp sounder. *Atmospheric Chemistry and Physics*, 9(16), 6041–6054.
- Coheur, P.-F., Clarisse, L., Turquety, S., Hurtmans, D., & Clerbaux, C. (2009). IASI measurements of reactive trace species in biomass burning plumes. *Atmospheric Chemistry and Physics*, 9(15), 5655–5667.
- Croft, B., Martin, R. V., Leaitch, W. R., Burkart, J., Chang, R. Y.-W., Collins, D. B., et al. (2019). Arctic marine secondary organicaerosol contributes significantly to summertime particle size distributions in the Canadian Arctic archipelago. *Atmospheric Chemistry and Physics*, 19(5), 2787–2812. <https://doi.org/10.5194/acp-19-2787-2019>
- Croft, B., Wentworth, G. R., Martin, R. V., Leaitch, W. R., Murphy, J. G., Murphy, B. N., et al. (2016). Contribution of Arctic seabird-colony ammonia to atmospheric particles and cloud-albedo radiative effect. *Nature communications*, 7, 13444.
- Dammers, E., Palm, M., Van Damme, M., Vigouroux, C., Smale, D., Conway, S., et al. (2016). An evaluation of IASI-NH₃ with ground-based Fourier transform infrared spectroscopy measurements. *Atmospheric Chemistry and Physics*, 16(16), 10,351–10,368.
- Dammers, E., Shephard, M. W., Palm, M., Cady-Pereira, K., Capps, S., Lutsch, E., et al. (2017). Validation of the CrIS fast physical NH₃ retrieval with ground-based FTIR. *Atmospheric Measurement Techniques*, 10(7), 2645–2667. <https://doi.org/10.5194/amt-10-2645-2017>
- Dammers, E., Vigouroux, C., Palm, M., Mahieu, E., Warneke, T., Smale, D., et al. (2015). Retrieval of ammonia from ground-based FTIR solar spectra. *Atmospheric Chemistry and Physics*, 15(22), 12,789–12,803. <https://doi.org/10.5194/acp-15-12789-2015>
- Dentener, F. J., & Crutzen, P. J. (1994). A three-dimensional model of the global ammonia cycle. *Journal of Atmospheric Chemistry*, 19(4), 331–369.
- Eastham, S. D., & Jacob, D. J. (2017). Limits on the ability of global Eulerian models to resolve intercontinental transport of chemical plumes. *Atmospheric Chemistry and Physics*, 17(4), 2543–2553.
- Erisman, J. W., Galloway, J., Seitzinger, S., Bleeker, A., & Butterbach-Bahl, K. (2011). Reactive nitrogen in the environment and its effect on climate change. *Current Opinion in Environmental Sustainability*, 3(5), 281–290.
- Eyring, V., Waugh, D. W., Bodeker, G. E., Cordero, E., Akiyoshi, H., Austin, J., et al. (2007). Multimodel projections of stratospheric ozone in the 21st century. *Journal of Geophysical Research*, 112, D03304. <https://doi.org/10.1029/2006JD008332>
- Fisher, J. A., Jacob, D. J., Wang, Q., Bahreini, R., Carouge, C. C., Cubison, M. J., et al. (2011). Sources, distribution, and acidity of sulfate-ammonium aerosol in the Arctic in winter-spring. *Atmospheric Environment*, 45(39), 7301–7318.
- Flannigan, M. D., Krawchuk, M. A., de Groot, W. J., Mike Wotton, B., & Gowman, L. M. (2009). Implications of changing climate for global wildland fire. *International Journal of Wildland Fire*, 18(5), 483–507.
- Fountoukis, C., & Nenes, A. (2007). ISORROPIA II: A computationally efficient thermodynamic equilibrium model for K⁺-Ca²⁺-Mg²⁺-NH₄⁺-Na⁺-SO₄²⁻-NO₃-Cl-H₂O aerosols. *Atmospheric Chemistry and Physics*, 7(17), 4639–4659.
- Galloway, J. N., Aber, J. D., Erisman, J. W., Seitzinger, S. P., Howarth, R. W., Cowling, E. B., & Cosby, B. J. (2003). The nitrogen cascade. *Bioscience*, 53(4), 341–356. [https://doi.org/10.1641/0006-3568\(2003\)053\[0341:TNC\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2003)053[0341:TNC]2.0.CO;2)
- George, M., Clerbaux, C., Hurtmans, D., Turquety, S., Coheur, P.-F., Pommier, M., et al. (2009). Carbon monoxide distributions from the IASI/METOP mission: Evaluation with other space-borne remote sensors. *Atmospheric Chemistry and Physics*, 9(21), 8317–8330.
- Giglio, L., Boschetti, L., Roy, D. P., Humber, M. L., & Justice, C. O. (2018). The collection 6 MODIS burned area mapping algorithm and product. *Remote Sensing of the Environment*, 217, 72–85.
- Hannigan, J. W., Coffey, M. T., & Goldman, A. (2009). Semiautonomous FTS observation system for remote sensing of stratospheric and tropospheric gases. *Journal of Atmospheric and Oceanic Technology*, 26(9), 1814–1828. <https://doi.org/10.1175/2009JTECHA1230.1>
- Holtlag, A. A. M., & Boville, B. A. (1993). Local versus nonlocal boundary-layer diffusion in a global climate model. *Journal of climate*, 6(10), 1825–1842. [https://doi.org/10.1175/1520-0442\(1993\)006h1825:LVNBLD2.0.CO;2](https://doi.org/10.1175/1520-0442(1993)006h1825:LVNBLD2.0.CO;2)
- Hurtmans, D., Coheur, P.-F., Wespes, C., Clarisse, L., Scharf, O., Clerbaux, C., et al. (2012). FORLI radiative transfer and retrieval code for IASI. *Journal of quantitative spectroscopy & radiative transfer*, 113(11), 1391–1408.
- Ireland, G., & Petropoulos, G. P. (2015). Exploring the relationships between post-fire vegetation regeneration dynamics, topography and burn severity: A case study from the Montane Cordillera Ecozones of Western Canada. *Applied Geography*, 56, 232–248.
- Janssens-Maenhout, G., Pagliari, V., Guizzardi, D., & Muntean, M. (2013). Global emission inventories in the Emission Database for Global Atmospheric Research (EDGAR)-Manual (I). <https://doi.org/10.2788/81454>
- Kaiser, J. W., Heil, A., Andreae, M. O., Benedetti, A., Chubarova, N., Jones, L., et al. (2012). Biomass burning emissions estimated with a global fire assimilation system based on observed fire radiative power. *Biogeosciences*, 9(1), 527–554. <https://doi.org/10.5194/bg-9-527-2012>
- Karllsson, P. E., Ferm, M., Tømmervik, H., Hole, L. R., Pihl Karllsson, G., Ruoho-Airola, T., et al. (2013). Biomass burning in eastern Europe during spring 2006 caused high deposition of ammonium in northern Fennoscandia. *Environmental pollution*, 176, 71–79.
- Keenan, T. F., & Riley, W. J. (2018). Greening of the land surface in the world's cold regions consistent with recent warming. *Nature Climate Change*, 8(9), 825–828.

- Kerzenmacher, T., Dils, B., Kumps, N., Blumenstock, T., Clerbaux, C., Coheur, P.-F., et al. (2012). Validation of IASI FORLI carbon monoxide retrievals using FTIR data from NDACC. *Atmospheric Measurement Techniques*, 5(11), 2751–2761.
- Kharol, S. K., Shephard, M. W., McLinden, C. A., Zhang, L., Sioris, C. E., O'Brien, J. M., et al. (2018). Dry deposition of reactive nitrogen from satellite observations of ammonia and nitrogen dioxide over North America. *Geophysical Research Letters*, 45, 1157–1166. <https://doi.org/10.1002/2017GL075832>
- Khaykin, S. M., GodinâBeekmann, S., Hauchecorne, A., Pelon, J., Ravetta, F., & Keckhut, P. (2018). Stratospheric smoke with unprecedentedly high backscatter observed by lidars above southern France. *Geophysical Research Letters*, 45, 1639–1646. <https://doi.org/10.1002/2017GL076763>
- Kirchmeier-Young, M. C., Gillett, N. P., Zwiers, F. W., Cannon, A. J., & Anslow, F. S. (2018). Attribution of the influence of human induced climate change on an extreme fire season. *Earth's Future*, 7, 2–10. <https://doi.org/10.1029/2018EF001050>
- Klonecki, A. (2003). Seasonal changes in the transport of pollutants into the Arctic troposphere-model study. *Journal of Geophysical Research*, 108(D4), 1479. <https://doi.org/10.1029/2002JD002199>
- Lefer, B. L., Talbot, R. W., Harriss, R. H., Bradshaw, J. D., Sandholm, S. T., Olson, J. O., et al. (1994). Enhancement of acidic gases in biomass burning impacted air masses over Canada. *Journal of Geophysical Research*, 99(D1), 1721.
- Lefer, B. L., Talbot, R. W., & Munger, J. W. (1999). Nitric acid and ammonia at a rural northeastern U.S. site. *Journal of Geophysical Research*, 104(D1), 1645–1661.
- Li, Q., Jacob, D. J., Bey, I., Yantosca, R. M., Zhao, Y., Kondo, Y., & Notholt, J. (2000). Atmospheric hydrogen cyanide (HCN): Biomass burning source, ocean sink? *Geophysical Research Letters*, 27(3), 357–360.
- Li, Q., Jacob, D. J., Yantosca, R. M., Heald, C. L., Singh, H. B., Koike, M., et al. (2003). A global three-dimensional model analysis of the atmospheric budgets of HCN and CH₃ CN: Constraints from aircraft and ground measurements. *Journal of Geophysical Research*, 108(D21), 8827. <https://doi.org/10.1029/2002JD003075>
- Liu, H., Jacob, D. J., Bey, I., & Yantosca, R. M. (2001). Constraints from ²¹⁰Pb and ⁷Be on wet deposition and transport in a global three-dimensional chemical tracer model driven by assimilated meteorological fields. *Journal of Geophysical Research*, 106(D11), 12,109–12,128.
- Lutsch, E., Dammers, E., Conway, S., & Strong, K. (2016). Long-range transport of NH₃, CO, HCN, and C₂H₆ from the 2014 Canadian wildfires. *Geophysical Research Letters*, 43, 8286–8297. <https://doi.org/10.1002/2016GL070114>
- Macias Fauria, M., & Johnson, E. A. (2008). Climate and wildfires in the North American boreal forest. *Philosophical transactions of the Royal Society of London. Series B, Biological sciences*, 363, 2317–2329.
- Myneni, R. B., Keeling, C. D., Tucker, C. J., Asrar, G., & Nemani, R. R. (1997). Increased plant growth in the northern high latitudes from 1981 to 1991. *Nature*, 386, 698–702.
- Natural Resources Canada (2013). Forest classification | Natural Resources Canada. <https://www.nrcan.gc.ca/forests/measuring-reporting/classification/13179>. (Accessed: 2019-1-30).
- Park, R. J. (2004). Natural and transboundary pollution influences on sulfate-nitrate-ammonium aerosols in the United States: Implications for policy. *Journal of Geophysical Research*, 109, D15204. <https://doi.org/10.1029/2003JD004473>
- Park, R. J., Jacob, D. J., Kumar, N., & Yantosca, R. M. (2006). Regional visibility statistics in the United States: Natural and transboundary pollution influences, and implications for the regional haze rule. *Atmospheric Environment*, 40(28), 5405–5423.
- Paton-Walsh, C., Deutscher, N. M., Griffith, D. W. T., Forgan, B. W., Wilson, S. R., Jones, N. B., & Edwards, D. P. (2010). Trace gas emissions from savanna fires in northern Australia. *Journal of Geophysical Research*, 115, D16314. <https://doi.org/10.1029/2009JD013309>
- Paton-Walsh, C., Jones, N. B., Wilson, S. R., Haverd, V., Meier, A., Griffith, D. W. T., & Rinsland, C. P. (2005). Measurements of trace gas emissions from Australian forest fires and correlations with coincident measurements of aerosol optical depth. *Journal of Geophysical Research*, 110, D24305. <https://doi.org/10.1029/2005JD006202>
- Paulot, F., Paynter, D., Ginoux, P., Naik, V., Whitburn, S., Van Damme, M., et al. (2017). Gas-aerosol partitioning of ammonia in biomass burning plumes: Implications for the interpretation of spaceborne observations of ammonia and the radiative forcing of ammonium nitrate. *Geophysical Research Letters*, 44, 8084–8093. <https://doi.org/10.1002/2017GL074215>
- Peterson, D. A., Campbell, J. R., Hyer, E. J., Fromm, M. D., Kablick, G. P., Cossuth, J. H., & DeLand, M. T. (2018). Wildfire-driven thunderstorms cause a volcano-like stratospheric injection of smoke. *NPJ Climate and Atmospheric Science*, 1(1), 30.
- Petetin, H., Sauvage, B., Parrington, M., Clark, H., Fontaine, A., Athier, G., et al. (2018). The role of biomass burning as derived from the tropospheric CO vertical profiles measured by AGOS aircraft in 2002–2017. *Atmospheric Chemistry and Physics*, 18(23), 17,277–17,306.
- Philip, S., Martin, R. V., & Keller, C. A. (2016). Sensitivity of chemistry-transport model simulations to the duration of chemical and transport operators: A case study with GEOS-Chem v10-01. *Geoscientific Model Development*, 9(5), 1683–1695.
- Pope, C. A. 3rd, Ezzati, M., & Dockery, D. W. (2009). Fine-particulate air pollution and life expectancy in the United States. *The New England Journal of Medicine*, 360(4), 376–386. <https://doi.org/10.1056/NEJMSa0805646>
- R'Honi, Y., Clarisse, L., Clerbaux, C., Hurtmans, D., Dufлот, V., Turquety, S., et al. (2013). Exceptional emissions of NH₃ and HCOOH in the 2010 Russian wildfires. *Atmospheric Chemistry and Physics*, 13(8), 4171–4181.
- Rastigejev, Y., Park, R., Brenner, M. P., & Jacob, D. J. (2010). Resolving intercontinental pollution plumes in global models of atmospheric transport. *Journal of Geophysical Research*, 115, D02302. <https://doi.org/10.1029/2009JD012568>
- Riddick, S. N., Dragosits, U., Blackall, T. D., Daunt, F., Wanless, S., & Sutton, M. A. (2012). The global distribution of ammonia emissions from seabird colonies. *Atmospheric Environment*, 55, 319–327.
- Rodgers, C. D., & Connor, B. J. (2003). Intercomparison of remote sounding instruments. *Journal of Geophysical Research*, 108(D3), 4116. <https://doi.org/10.1029/2002JD002299>
- Rothman, L. S., Gordon, I. E., Barbe, A., Benner, D. C., Bernath, P. F., Birk, M., et al. (2009). The HITRAN 2008 molecular spectroscopic database. *Journal of quantitative spectroscopy & radiative transfer*, 110(9), 533–572.
- Saha, S., Moorthi, S., Wu, X., Wang, J., Nadiga, S., Tripp, P., et al. (2011). *NCEP Climate Forecast System version 2 (CFSv2) 6-hourly products*. Boulder CO: Research Data Archive at the National Center for Atmospheric Research, Computational and Information Systems Laboratory.
- Schimel, J. P., & Bennett, J. (2004). Nitrogen mineralization: Challenges of a changing paradigm. *Ecology*, 85(3), 591–602. <https://esajournals.onlinelibrary.wiley.com/doi/pdf/10.1890/03-8002>
- Schlesinger, W. H., & Hartley, A. E. (1992). A global budget for atmospheric NH₃. *Biogeochemistry*, 15(3), 191–211.
- Sekimoto, K., Koss, A. R., Gilman, J. B., Selimovic, V., Coggon, M. M., Zarzana, K. J., et al. (2018). High- and low-temperature pyrolysis profiles describe volatile organic compound emissions from western US wildfire fuels. *Atmospheric Chemistry and Physics*, 18(13), 9263–9281.
- Shephard, M. W., & Cady-Pereira, K. E. (2015). Cross-track Infrared Sounder (CrIS) satellite observations of tropospheric ammonia. *Atmospheric Measurement Techniques*, 8(3), 1323–1336. <https://doi.org/10.5194/amt-8-1323-2015>

- Shephard, M. W., Cady-Pereira, K. E., Luo, M., Henze, D. K., Pinder, R. W., Walker, J. T., & Clarisse, L. (2011). TES ammonia retrieval strategy and global observations of the spatial and seasonal variability of ammonia. *Atmospheric Chemistry and Physics*, *11*(20), 10,743–10,763. <https://doi.org/10.5194/acp-11-10743-2011>
- Shephard, M. W., McLinden, C. A., Cady-Pereira, K. E., Luo, M., Moussa, S. G., Leithead, A., et al. (2015). Tropospheric Emission Spectrometer (TES) satellite observations of ammonia, methanol, formic acid, and carbon monoxide over the Canadian oil sands: Validation and model evaluation. *Atmospheric Measurement Techniques*, *8*(12), 5189–5211. <https://doi.org/10.5194/amt-8-5189-2015>
- Shi, Y., Matsunaga, T., Saito, M., Yamaguchi, Y., & Chen, X. (2015). Comparison of global inventories of CO₂ emissions from biomass burning during 2002–2011 derived from multiple satellite products. *Environmental Pollution*, *206*, 479–487.
- Shindell, D. T., Chin, M., Dentener, F., Doherty, R. M., Faluvegi, G., Fiore, A. M., et al. (2008). A multi-model assessment of pollution transport to the Arctic. *Atmospheric Chemistry and Physics*, *8*(17), 5353–5372.
- Singh, H. B., & Zimmerman, P. B. (1992). Atmospheric distribution and sources of nonmethane hydrocarbons, *Gaseous pollutants: Characterization and cycling* pp. 177–235. New York: Jon Wiley and Sons.
- Stein, O., & Rudolph, J. (2007). Modeling and interpretation of stable carbon isotope ratios of ethane in global chemical transport models. *Journal of Geophysical Research*, *112*, D14308. <https://doi.org/10.1029/2006JD008062>
- Stohl, A. (2006). Characteristics of atmospheric transport into the Arctic troposphere. *Journal of Geophysical Research*, *111*, D11306. <https://doi.org/10.1029/2005JD006888>
- Stohl, A., Forster, C., Frank, A., Seibert, P., & Wotawa, G. (2005). Technical note: The Lagrangian particle dispersion model FLEXPART version 6.2. *Atmospheric Chemistry and Physics*, *5*(9), 2461–2474. <https://doi.org/10.5194/acp-5-2461-2005>
- Toon, G. C., Blavier, J.-F., Sen, B., Margitan, J. J., Webster, C. R., May, R. D., et al. (1999). Comparison of MkIV balloon and ER-2 aircraft measurements of atmospheric trace gases. *Journal of Geophysical Research*, *104*(D21), 26,779–26,790.
- Turquety, S., Logan, J. A., Jacob, D. J., Hudman, R. C., Leung, F. Y., Heald, C. L., et al. (2007). Inventory of boreal fire emissions for North America in 2004: Importance of peat burning and pyroconvective injection. *Journal of Geophysical Research*, *112*, 512. <https://doi.org/10.1029/2006JD007281>
- Urbanski, S. P. (2013). Combustion efficiency and emission factors for wildfire-season fires in mixed conifer forests of the northern Rocky Mountains, US. *Atmospheric Chemistry and Physics*, *13*(14), 7241–7262.
- Val Martin, M., Logan, J. A., Kahn, R. A., Leung, F.-Y., Nelson, D. L., & Diner, D. J. (2010). Smoke injection heights from fires in North America: Analysis of 5 years of satellite observations. *Atmospheric Chemistry and Physics*, *10*(4), 1491–1510.
- Van Damme, M., Clarisse, L., Dammers, E., Liu, X., Nowak, J. B., Clerbaux, C., et al. (2015). Towards validation of ammonia (NH₃) measurements from the IASI satellite. *Atmospheric Measurement Techniques*, *8*(3), 1575–1591. <https://doi.org/10.5194/amt-8-1575-2015>
- Van Damme, M., Clarisse, L., Heald, C. L., Hurtmans, D., Ngadi, Y., Clerbaux, C., et al. (2014). Global distributions, time series and error characterization of atmospheric ammonia (NH₃) from IASI satellite observations. *Atmospheric Chemistry and Physics*, *14*(6), 2905–2922. <https://doi.org/10.5194/acp-14-2905-2014>
- Van Damme, M., Clarisse, L., Whitburn, S., Hadji-Lazaro, J., Hurtmans, D., Clerbaux, C., & Coheur, P.-F. (2018). Industrial and agricultural ammonia point sources exposed. *Nature*, *564*, 99–103.
- Van Damme, M., Erisman, J. W., Clarisse, L., Dammers, E., Whitburn, S., Clerbaux, C., et al. (2015). Worldwide spatiotemporal atmospheric ammonia (NH₃) columns variability revealed by satellite. *Geophysical Research Letters*, *42*, 8660–8668. <https://doi.org/10.1002/2015GL065496>
- Van Damme, M., Whitburn, S., Clarisse, L., Clerbaux, C., Hurtmans, D., & Coheur, P.-F. (2017). Version 2 of the IASI NH₃ neural network retrieval algorithm: Near-real-time and reanalysed datasets. *Atmospheric Measurement Techniques*, *10*(12), 4905–4914.
- Viatte, C., Strong, K., Hannigan, J., Nussbaumer, E., Emmons, L. K., Conway, S., et al. (2015). Identifying fire plumes in the Arctic with tropospheric FTIR measurements and transport models. *Atmospheric Chemistry and Physics*, *15*(5), 2227–2246.
- Viatte, C., Strong, K., Paton-Walsh, C., Mendonca, J., O'Neill, N. T., & Drummond, J. R. (2013). Measurements of CO, HCN, and C₂H₆ total columns in smoke plumes transported from the 2010 Russian boreal forest fires to the Canadian high Arctic. *Atmosphere-Ocean*, *51*(5), 522–531.
- Viatte, C., Strong, K., Walker, K. A., & Drummond, J. R. (2014). Five years of CO, HCN, C₂H₆, C₂H₂, CH₃OH, HCOOH and H₂CO total columns measured in the Canadian high Arctic. *Atmospheric Measurement Techniques*, *7*(6), 1547–1570. <https://doi.org/10.5194/amt-7-1547-2014>
- Vigouroux, C., Stavrakou, T., Whaley, C., Dils, B., Dufflot, V., Hermans, C., et al. (2012). FTIR time-series of biomass burning products (HCN, C₂H₆, C₂H₂, CH₃OH, and HCOOH) at reunion island (21°S, 55°E) and comparisons with model data. *Atmospheric Chemistry and Physics*, *12*(21), 10367–10385.
- Ward, D. E., & Hardy, C. C. (1991). Smoke emissions from wildland fires. *Environment International*, *17*(2), 117–134.
- Warner, J. X., Wei, Z., Strow, L. L., Dickerson, R. R., & Nowak, J. B. (2016). The global tropospheric ammonia distribution as seen in the 13-year AIRS measurement record. *Atmospheric Chemistry and Physics*, *16*(8), 5467–5479. <https://doi.org/10.5194/acp-16-5467-2016>
- Wentworth, G. R., Murphy, J. G., Croft, B., Martin, R. V., Pierce, J. R., Côté, J.-S., et al. (2016). Ammonia in the summertime Arctic marine boundary layer: Sources, sinks, and implications. *Atmospheric Chemistry and Physics*, *16*(4), 1937–1953.
- Wesely, M. L. (1989). Parameterization of surface resistances to gaseous dry deposition in regional-scale numerical models. *Atmospheric Environment*, *23*(6), 1293–1304.
- Whaley, C. H., Makar, P. A., Shephard, M. W., Zhang, L., Zhang, J., Zheng, Q., et al. (2018). Contributions of natural and anthropogenic sources to ambient ammonia in the Athabasca oil sands and north-western Canada. *Atmospheric Chemistry and Physics*, *18*(3), 2011–2034.
- Whitburn, S., Van Damme, M., Clarisse, L., Bauduin, S., Heald, C. L., Hadji-Lazaro, J., et al. (2016). A flexible and robust neural network IASI-NH₃ retrieval algorithm. *Journal of Geophysical Research: Atmospheres*, *121*, 6581–6599. <https://doi.org/10.1002/2016JD024828>
- Whitburn, S., Van Damme, M., Clarisse, L., Hurtmans, D., Clerbaux, C., & Coheur, P.-F. (2017). IASI-derived NH₃ enhancement ratios relative to CO for the tropical biomass burning regions. *Atmospheric Chemistry and Physics*, *17*(19), 12,239–12,252.
- Whitburn, S., Van Damme, M., Clarisse, L., Turquety, S., Clerbaux, C., & Coheur, P.-F. (2016). Doubling of annual ammonia emissions from the peat fires in Indonesia during the 2015 El Niño. *Geophysical Research Letters*, *43*, 11,007–11,014. <https://doi.org/10.1002/2016GL070620>
- Whitburn, S., Van Damme, M., Kaiser, J. W., van der Werf, G. R., Turquety, S., Hurtmans, D., et al. (2015). Ammonia emissions in tropical biomass burning regions: Comparison between satellite-derived emissions and bottom-up fire inventories. *Atmospheric Environment*, *121*, 42–54.
- Wotton, B. M., Flannigan, M. D., & Marshall, G. A. (2017). Potential climate change impacts on fire intensity and key wildfire suppression thresholds in Canada. *Environmental Research Letters*, *12*(9), 095003.

- Wotton, B. M., Nock, C. A., & Flannigan, M. D. (2010). Forest fire occurrence and climate change in Canada. *International Journal of Wildland Fire*, 19(3), 253–271.
- Xiao, Y., Logan, J. A., Jacob, D. J., Hudman, R. C., Yantosca, R., & Blake, D. R. (2008). Global budget of ethane and regional constraints on U.S. sources. *Journal of Geophysical Research*, 113, D21306. <https://doi.org/10.1029/2007JD009415>
- Yokelson, R. J., Andreae, M. O., & Akagi, S. K. (2013). Pitfalls with the use of enhancement ratios or normalized excess mixing ratios measured in plumes to characterize pollution sources and aging. *Atmospheric Measurement Techniques*, 6(8), 2155–2158.
- Yokelson, R. J., Crounse, J. D., DeCarlo, P. F., Karl, T., Urbanski, S., Atlas, E., et al. (2009). Emissions from biomass burning in the Yucatan. *Atmospheric Chemistry and Physics*, 9(15), 5785–5812.
- York, D., Evensen, N. M., Martinez, M. L., & De Basabe Delgado, J. (2004). Unified equations for the slope, intercept, and standard errors of the best straight line. *American journal of physics*, 72(3), 367–375. <https://doi.org/10.1119/1.1632486>
- Zhang, L., Gong, S., Padro, J., & Barrie, L. (2001). A size-segregated particle dry deposition scheme for an atmospheric aerosol module. *Atmospheric Environment*, 35(3), 549–560. [https://doi.org/10.1016/S1352-2310\(00\)00326-5](https://doi.org/10.1016/S1352-2310(00)00326-5)
- Zhao, Y. (2002). Spectroscopic measurements of tropospheric CO, C₂H₆, C₂H₂, and HCN in northern Japan. *Journal of Geophysical Research*, 107(D18), 4343. <https://doi.org/10.1029/2001JD000748>
- Zhu, L., Henze, D., Bash, J., Jeong, G.-R., Cady-Pereira, K., Shephard, M., et al. (2015). Global evaluation of ammonia bidirectional exchange and livestock diurnal variation schemes. *Atmospheric Chemistry and Physics*, 15(22), 12,823–12,843.
- Zhu, Z., Piao, S., Myneni, R. B., Huang, M., Zeng, Z., Canadell, J. G., et al. (2016). Greening of the Earth and its drivers. *Nature Climate Change*, 6, 791.