



**HAL**  
open science

## **Risky future for Mediterranean forests unless they undergo extreme carbon fertilization**

Guillermo Gea-Izquierdo, Antoine Nicault, Giovanna Battipaglia, Isabel Dorado-Liñán, Emilia Gutiérrez, Montserrat Ribas, Joel Guiot

► **To cite this version:**

Guillermo Gea-Izquierdo, Antoine Nicault, Giovanna Battipaglia, Isabel Dorado-Liñán, Emilia Gutiérrez, et al.. Risky future for Mediterranean forests unless they undergo extreme carbon fertilization. *Global Change Biology*, 2017, 23 (7), pp.2915-2927. 10.1111/gcb.13597 . insu-02269704

**HAL Id: insu-02269704**

**<https://insu.hal.science/insu-02269704>**

Submitted on 23 Aug 2019

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 **Risky future for Mediterranean forests unless**  
2 **they undergo extreme carbon fertilization**

3 **Running head:** Mediterranean forests, climate change and CO<sub>2</sub>

4 <sup>1</sup>G Gea-Izquierdo, <sup>2</sup>A Nicault, <sup>3,4</sup>G Battipaglia, <sup>1</sup>I Dorado-Liñán, <sup>5</sup>E, Gutiérrez, <sup>5</sup>M  
5 Ribas, <sup>6</sup>J Guiot

6 <sup>1,7</sup>INIA-CIFOR. Ctra. La Coruña km. 7.5 28040 Madrid, Spain. <sup>2</sup>Aix-Marseille

7 Université/CNRS FR 3098 ECCOREV 13545 Aix-en-Provence, France. <sup>3</sup>Department of  
8 Environmental, Biological and Pharmaceutical Sciences and Technologies<sup>[L]</sup>Second

9 University of Naples<sup>[L]</sup>Via Vivaldi, 43 - 81100 Caserta, Italy. <sup>4</sup>Ecole Pratique des

10 Hautes Etudes (PALECO EPHE), Institut des Sciences de l'Evolution, University of

11 Montpellier 2, F-34090 Montpellier, France. <sup>5</sup>Departament <sup>4</sup>Departament d'Ecologia,

12 Universitat de Barcelona, Avda. Diagonal 643, 08028 Barcelona, Spain. <sup>6</sup>Aix-Marseille

13 Université/CNRS/IRD UM 34, CEREGE 13545 Aix-en-Provence. <sup>7</sup>Corresponding

14 author: phone 0034913471461; email gea.guillermo@inia.es

15  
16 **Abstract**

17 Forest performance is challenged by climate change but higher atmospheric [CO<sub>2</sub>] (c<sub>a</sub>)

18 could help trees mitigate the negative effect of enhanced water stress. Forest projections

19 using data-assimilation with mechanistic models are a valuable tool to assess forest

20 performance. Firstly, we used dendrochronological data from 12 Mediterranean tree

21 species (6 conifers, 6 broadleaves) to calibrate a process-based vegetation model at 77

22 sites. Secondly, we conducted simulations of gross primary production (GPP) and radial

23 growth using an ensemble of climate projections for the period 2010-2100 for the high-

24 emission RCP8.5 and low-emission RCP2.6 scenarios. GPP and growth projections

25 were simulated using climatic data from the two RCPs combined with: (i) expected c<sub>a</sub>;

26 (ii) constant  $c_a = 390$  ppm, to test a purely climate-driven performance excluding  
27 compensation from carbon fertilization. The model accurately mimicked the growth  
28 trends since the 1950s when, despite increasing  $c_a$ , enhanced evaporative demands  
29 precluded a global net positive effect on growth. Modeled annual growth and GPP  
30 showed similar long-term trends. Under RCP2.6 (i.e. temperatures below  $+2^\circ\text{C}$  with  
31 respect to preindustrial values) the forests showed resistance to future climate (as  
32 expressed by non-negative trends in growth and GPP) except for some coniferous sites.  
33 Using exponentially growing  $c_a$  and climate as from RCP8.5, carbon fertilization  
34 overrode the negative effect of the highly constraining climatic conditions under that  
35 scenario. This effect was particularly evident above 500 ppm (which is already over  
36  $+2^\circ\text{C}$ ), which seems unrealistic and likely reflects model miss-performance at high  $c_a$   
37 above the calibration range. Thus, forest projections under RCP8.5 preventing carbon  
38 fertilization displayed very negative forest performance at the regional scale. This  
39 suggests that most of western Mediterranean forests would successfully acclimate to the  
40 coldest climate change scenario but be vulnerable to a climate warmer than  $+2^\circ\text{C}$  unless  
41 the trees developed an exaggerated fertilization response to  $[\text{CO}_2]$ .

42

43 **Keywords:** Dendroecology; process-based models; carbon fertilization; climate change;  
44 MAIDEN; water stress; forest dynamics.

45 **Type of paper:** Original research article

## 46 **Introduction**

47 Future climate will trigger changes in ecosystem functioning, including  
48 enhancement in forest vulnerability to water stress (Giorgi & Lionello 2008; van der  
49 Molen et al. 2011; Anderegg et al. 2015). Understanding how forests will respond to  
50 warmer conditions but under higher than present  $c_a$  is crucial to assess future forest  
51 performance. Theoretically, plants should enhance growth and net primary productivity  
52 (NPP) by optimization of different functional traits in response to elevated  $[\text{CO}_2]$  (i.e.  $c_a$   
53 levels way above present values,  $e\text{CO}_2$ ) if this was a limiting factor. In practice, rising  $c_a$   
54 has enhanced intrinsic water-use efficiency (iWUE) in forests but this was not generally  
55 translated on a net increase of growth, meaning that other factors such as water stress  
56 and/or nutrient limitation have overridden the potential positive effect of  $\text{CO}_2$  (Peñuelas  
57 et al. 2011; Keenan et al. 2013; Reichstein et al. 2013; van der Sleen et al. 2014; Kim et  
58 al. 2016).

59 The net effect on tree growth of the interaction 'Climate x  $\text{CO}_2$ ' can depend  
60 nonlinearly on  $c_a$  levels (Reichstein et al. 2013). Observational data show evidence up  
61 to current  $c_a < 403$  ppm whereas future emission scenarios project  $c_a$  far above this level  
62 (IPCC 2014). Free-Air Carbon dioxide Enrichment (FACE) experiments were designed  
63 to address this issue. In these experiments  $[\text{CO}_2]$  was elevated up to 600-800 ppm but  
64 they were carried out under current environmental conditions mostly on temperate  
65 forests (Battipaglia et al. 2013; De Kauwe et al. 2013; Baig et al. 2015; Kim et al. 2016;  
66 Norby et al. 2016). Thus, the effect of  $e\text{CO}_2$  on forest performance in relation to climate  
67 and other environmental factors needs to be addressed in other biomes where more  
68 constraining (warmer and drier) conditions are expected for the future (Giorgi &  
69 Lionello 2008; García-Ruiz et al. 2011; IPCC 2014). The role that  $\text{CO}_2$  could play to  
70 compensate the negative impact of increasing water stress on forests has long been



71 debated. Plants can coordinate different functional traits in response to eCO<sub>2</sub> and  
72 drought-prone species at dry sites could benefit more from eCO<sub>2</sub> (Medlyn & De Kauwe  
73 2013; Duursma *et al.* 2016; Kelly *et al.* 2016). Leaf-level responses are easier to predict  
74 than canopy or ecosystem-level responses (Fatichi *et al.* 2015). Consequently, there is  
75 still uncertainty on how the forest carbon cycle will adjust in the future because multiple  
76 interactive factors determine the net response of forests at different scales (Breda *et al.*  
77 2006; Niinemets 2010; van der Molen *et al.* 2011; Kattge *et al.* 2011).

78 Vegetation models combine the effect of different stress factors on different  
79 functional traits to achieve a proper understanding of forest functioning. These models  
80 should be able to combine C-sink and C-source limitations to provide key information  
81 on how forests will develop in the future (Sala *et al.* 2012; McDowell *et al.* 2013;  
82 Fatichi *et al.* 2014; Anderegg *et al.* 2015; Walker *et al.* 2015). There is a constant need  
83 to improve the representation of hydrological, physical and biological processes in  
84 models. In addition, improvement of model performance needs to be achieved through  
85 benchmarking and data-assimilation (Peng *et al.* 2011; Pappas *et al.* 2013; Medlyn *et al.*  
86 2015; Prentice *et al.* 2015). Dendrochronological data have long been used to assess  
87 empirical relationships between climate and growth, which can be used as an indicator  
88 of tree fitness and performance (Fritts 1976). Process-based models can take into  
89 account the influence of CO<sub>2</sub> on plant functional acclimation. Thus they can help to  
90 reduce uncertainty in growth projections but need continuous feedback from multiproxy  
91 data to ensure realism (Guiot *et al.* 2014; Walker *et al.* 2015). Dendrochronological  
92 records can be used to improve complex process-based models and help to assess forest  
93 dynamics under global change (Babst *et al.* 2014). Assessing forest dynamics is  
94 particularly challenging in ecosystems like those under Mediterranean climate (Morales  
95 *et al.* 2005) where two stress periods (cold in winter and drought in summer) limit plant

96 performance. Warmer winters could enlarge the growing season and promote higher  
97 photosynthetic rates (just in evergreens) but also higher respiration rates, whereas  
98 warmer summers would exert a negative impact (higher water stress) on forests.  
99 Modeling the net effect on trees of the balance between these two periods is critical to  
100 assess the future forest response to climate change.

101 We analyzed the effect that forthcoming changes in climate and  $c_a$  will yield over  
102 Mediterranean forests, which are expected to face a high vulnerability to future climate  
103 (Giorgi & Lionello 2008; García-Ruiz *et al.* 2011; IPCC 2014). We calibrated a stand  
104 mechanistic model using a network of tree-ring growth chronologies including an  
105 ensemble of species covering a wide ecological and geographic range to ensure realism  
106 and biological robustness when simulating future forest performance at the regional  
107 scale under different climate and  $c_a$  scenarios. C-assimilation and C-allocation were  
108 explicitly controlled by climate and  $CO_2$  at different phenological stages (Misson 2004;  
109 Gea-Izquierdo *et al.* 2015). Importantly, the model includes a C storage pool to take into  
110 account carry-over effects and its daily scale can fit different limiting environmental  
111 conditions at different periods within and among years (Sala *et al.* 2012; Fatichi *et al.*  
112 2014). Thus, the net effect in response to the winter and summer stress periods was  
113 explicitly assessed. Forest projections were implemented using two contrasting  
114 representative concentration pathways (RCPs, van Vuuren *et al.* 2011). Using model  
115 simulations of future forest growth and GPP we addressed the following questions: (i)  
116 what will be the net effect of a warmer climate for Mediterranean forests?; (ii) to what  
117 extent could rising  $c_a$  help compensate the expected negative effect of climate warming  
118 on forest growth and productivity? (iii) how will Mediterranean forests perform in  
119 relation to the maximum temperature threshold for future climate (i.e. +2.0°C respect to  
120 preindustrial levels) agreed in the COP21 (<http://www.cop21paris.org/>)?

121

## 122 **Material and Methods**

### 123 *Forest sites: growth data for model calibration*

124 To calibrate the model at the regional scale we used dendrochronological data from  
125 77 forest sites including 12 Mediterranean tree species: 6 conifers and 6 broadleaves  
126 (App. 1). These data were either owned by the authors or obtained from databases  
127 (ITRDB, [https://www.ncdc.noaa.gov/data-access/paleoclimatology-data/datasets/tree-](https://www.ncdc.noaa.gov/data-access/paleoclimatology-data/datasets/tree-ring)  
128 [ring](https://www.ncdc.noaa.gov/data-access/paleoclimatology-data/datasets/tree-ring); DendroDB, <https://dendrodb.eccorev.fr/framedb.htm>). We explored the data to  
129 avoid chronologies where major site disturbances could have affected the decadal-to-  
130 multidecadal growth variations. The chronologies included were older than 80 years in  
131 order to avoid the effect of juvenile growth (data used for calibration started in 1950)  
132 and ended later than 1995. Exceptionally, some sites ending before 1995 in Algeria  
133 were included to ensure enough data from that region. The resulting calibration period  
134 slightly differed across sites due to the different time-span of chronologies, but always  
135 fell between 1950 and 2010 and was greater than 40 years. For the analysis, ring-width  
136 growth data were transformed to basal area increments (BAI,  $\text{cm}^2 \text{year}^{-1}$ ). One output  
137 from the model is C allocated to the tree stem ( $\text{g C m}^{-2} \text{year}^{-1}$ ). To make BAI and model  
138 output comparable for model calibration both data were normalized to unitless indices  
139 (Misson 2004; Gaucherel et al. 2008).

140

### 141 *Climate and $c_a$ : historical data and future scenarios*

142 Daily precipitation and temperature data used for model calibration for 1950-2010  
143 were either obtained from <http://www.meteo.unican.es/datasets/spain02> (Herrera *et al.*,  
144 2012) for Spain (20 km grid) or from <http://hydrology.princeton.edu/data.php> (Sheffield  
145 *et al.* 2006) for the rest ( $1^\circ$  grid). Data were downscaled to match mean climatic local

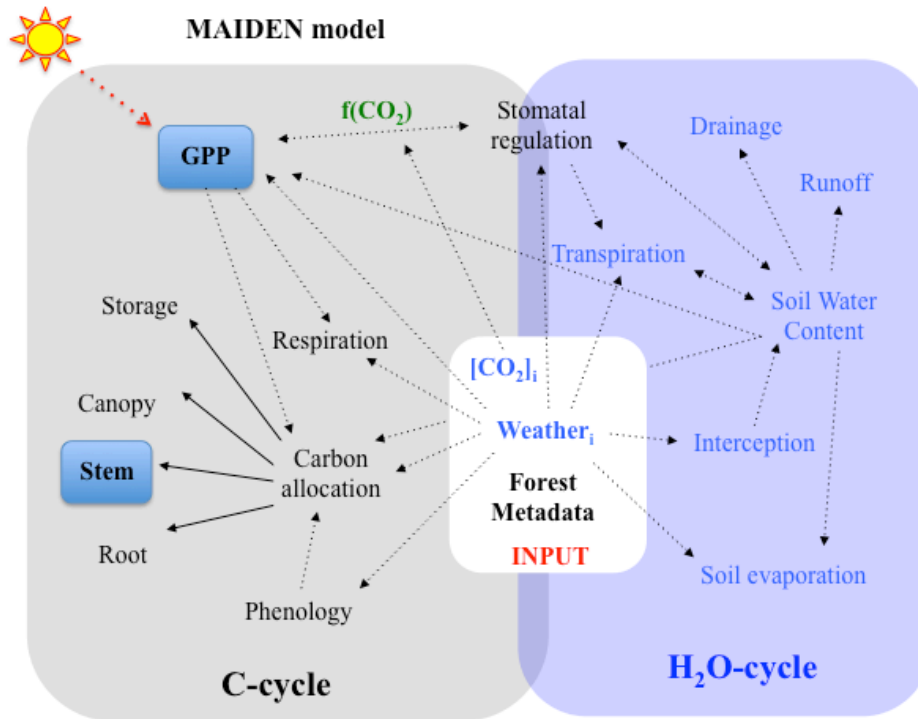
146 values where these were available. For the future forest projections we used two  
147 greenhouse gas (GHG) radiative forcing scenarios developed for the Fifth Report of the  
148 IPCC (IPCC 2014), called RCPs (Van Vuuren *et al.* 2011). RCP8.5 (GHGRF<8.5  
149 W/m<sup>2</sup>) is the “business as usual” scenario. RCP2.6 is the most optimistic and stringent  
150 among RCPs, corresponding to strong mitigation policies with a GHG radiative forcing  
151 constrained to remain <2.6 W/m<sup>2</sup>. RCP2.6 is the only RCP limiting global warming to  
152 +2°C relative to the pre-industrial level. We used a multimodel ensemble of 19  
153 simulations for RCP2.6 and 18 simulations for RCP8.5 performed by 13 climatic  
154 institutes (see App. 2). The global climate models have a coarse resolution, from one to  
155 more than five degrees depending on the model. There is often some mismatch between  
156 the stand level and input climatic data (Körner 2003; Potter *et al.* 2013; Pappas *et al.*  
157 2015). This was minimized as possible by downscaling climate scenarios to match the  
158 shared period of the historical data. Under RCP8.5 the projected climate for our study  
159 sites describes a relative increase in mean annual temperature (MAT) of +5.0°C and a  
160 decrease of over 40% in mean annual precipitation (MAP) by 2100 respect to current  
161 values. Under RCP2.6 the projected climate forecasts a mean increase in MAT of  
162 +1.0°C by 2068, stabilizing thereafter, with no decrease in MAP respect to current  
163 conditions (App. 3).

164

#### 165 *The process-based model MAIDEN*

166 The vegetation model MAIDEN (Fig. 1) was originally developed to be used with  
167 dendrochronological data by being calibrated to both time series of radial growth and  
168 estimates of transpiration from sap-flow experiments (Misson 2004). Recently, the  
169 model has been further developed to be used with evergreen Mediterranean taxa with a  
170 multiproxy approach using gross primary productivity (GPP) estimates from Eddy

171 covariance stations and plot growth data (Gea-Izquierdo *et al.* 2015). Inputs are daily  
 172 climatic data (precipitation, maximum and minimum temperatures) and  $c_a$ . In addition



173  
 174 Fig. 1. Outline of MAIDEN. Only GPP (gross primary production) and biomass  
 175 allocated to the stem (used to calibrate the model sitewise) estimates (blue boxes) are  
 176 reported along the manuscript. Daily output of each variable is generated based on the  
 177 input data of day  $i$ . 'Weather' corresponds to daily integrals of precipitation as well as  
 178 maximum and minimum daily temperatures. GPP and 'Stomatal conductance' are  
 179 functions of CO<sub>2</sub>, whereas variability within the other processes is mostly driven by  
 180 meteorological inputs directly or indirectly (e.g. SWC). For more details on the  
 181 functions and processes outlined see Misson (2004) and Gea-Izquierdo *et al.* (2015).

182

183 the model requires as input different site related physiographic characteristics and  
 184 species functional traits (see Gea-Izquierdo *et al.* 2015 for details). The processes within  
 185 the model are mainly functions of climate, CO<sub>2</sub> and soil water availability (hence water

186 stress). The model acts at the stand level calculating carbon and water fluxes (Fig. 1)  
187 using a coupled photosynthesis-stomatal conductance model. It uses the standard  
188 biochemical model of Farquhar *et al.* (1980) in which photosynthesis is driven by the  
189 most limiting between Rubisco-limited activity and electron-transport. Stomatal  
190 conductance is also estimated using a widely used equation as a function of vapor  
191 pressure deficit (VPD, Leuning 1995). After GPP and autotrophic respiration have been  
192 estimated carbon is allocated to different tree components. Photosynthesis and  
193 allocation are driven by decoupled non-linear (daily) functions of climate. Thus growth  
194 is not only a direct function of C availability and the model is designed to address in  
195 time not only C-source but also C-sink limitations, which is an important step required  
196 to achieve more robust and realistic vegetation models (Muller *et al.* 2011; Sala *et al.*  
197 2012; Fatichi *et al.* 2014). The model is particularly sensitive to water stress by  
198 implicitly modeling as functions of climate and water stress some functional and  
199 demographic traits such as leaf area, carbon allocation, leaf- and canopy-level  
200 photosynthesis and transpiration (Muller *et al.* 2011; Gea-Izquierdo *et al.* 2015;  
201 Duursma *et al.* 2016; Kelly *et al.* 2016). [CO<sub>2</sub>] only affects photosynthesis and stomatal  
202 conductance, i.e. leaf area or respiration are direct functions of climate but not CO<sub>2</sub>. A  
203 brief outline of the model is shown in Fig. 1.

204

#### 205 *Model calibration and ecological coherence of the parametric space*

206 Calibration of complex multiparametric models is necessary to improve model  
207 performance and because of the presence of collinearities between parameters and  
208 absence of an exact solution (Prentice *et al.* 2015). To ensure good model performance,  
209 it is important to assess the functional coherence of parameters to be calibrated. In  
210 addition to calibration, it is desirable to run independent validations particularly when

211 models are fitted for prediction purposes. Nevertheless, we could not run an  
212 independent crossvalidation for two reasons: (1) we calibrated against annual growth  
213 (i.e. we had a number of observations between 40 and 60) estimates by integrating  
214 annually the daily estimates from the model, therefore our data was too short to be split  
215 in two, (2) a jackknife was intractable both computationally and also because  
216 continuous (daily) time data series are needed to run the model, i.e. in case individual  
217 years were left out the model could not calculate the carbon and water dynamics needed  
218 to compute the complete time series at each site.

219 We implemented a species-specific approach rather than using plant functional  
220 types (PFTs) as it is often applied in ecosystem models (Kattge *et al.* 2011; Atkin *et al.*  
221 2015; Pappas *et al.* 2016). We applied the model at the regional scale and to different  
222 species to analyze forest performance under future climate and  $c_a$ . Data-assimilation  
223 was used to apply the model to different ecological conditions and species (Peng *et al.*  
224 2011; Medlyn *et al.* 2015). Overall, the growth data did not show a positive trend  
225 whereas  $c_a$  increased steadily in the calibration period (1951-2010). Therefore, by  
226 calibrating the model site-wise using non-detrended (but normalized) growth data and  
227 observed  $c_a$  levels we assured that the model excluded an artificial carbon fertilization  
228 effect on past growth. Additionally, to avoid overestimation of photosynthesis and get  
229 unbiased simulations (Schaefer *et al.* 2012), we ensured that maximum GPP daily  
230 integrals yielded within ranges given in Baldocchi *et al.* (2010): 4-6 g C m<sup>-2</sup> day<sup>-1</sup> for  
231 evergreens and of 10-14 g C m<sup>-2</sup> day<sup>-1</sup> for deciduous species. Similarly, we constrained  
232 annual GPP and NPP estimates to be within those measured for similar ecosystems (see  
233 Table 3 in Falge *et al.* 2002 and Table 3 in Luysaert *et al.* 2007). Species Specific leaf  
234 area (SLA) was obtained from Mediavilla *et al.* (2008) and Kattge *et al.* (2011).

235 Here, within the processes in Fig. 1 we show those functions with parameters  
 236 involved in the calibration phase. For more details, we refer to the original model in  
 237 Misson (2004) and the updated modified last version in Gea-Izquierdo *et al.* (2015).

$$238 \quad (i) \quad \theta_g(i) = \frac{1}{1 + \exp(\text{soil}_b \cdot (\text{SWC}(i) - \text{soil}_{ip}))} \quad [\text{E1}]$$

$$239 \quad (ii) \quad a_{31}(i) = (1 - \exp(p_{3moist} \cdot \text{SWC}(i)) \cdot \left( \exp\left(-0.5 \cdot \left(\frac{T_{max}(i) - p_{3temp}}{p_{3sd}}\right)^2\right)\right) \quad [\text{E2}]$$

$$240 \quad (iii) \quad a_{32}(i) = (1 - \exp(st_{3moist} \cdot \text{SWC}(i)) \cdot \left( \exp\left(-0.5 \cdot \left(\frac{T_{max}(i) - st_{3temp}}{st_{3sd\_temp}}\right)^2\right)\right) \quad [\text{E3}]$$

$$241 \quad (iv) \quad a_4(i) = (1 - \exp(st_{4temp} \cdot T_{max}(i)) \cdot \left( \exp\left(-0.5 \cdot \left(\frac{\text{SWC}(i)}{st_{4sd\_moist}}\right)^2\right)\right) \quad [\text{E4}]$$

242  $\theta_g$  is a soil water stress function affecting stomatal conductance.  $a_{31}$ ,  $a_{32}$  and  $a_4$  are  
 243 allocation functions for two different phenological periods (3 and 4).  $a_{31}$  is related to the  
 244 leaves and  $a_{32}$  to the stem, whereas  $a_4$  determines C allocation between the stem and  
 245 storage. SWC is soil water content and  $T_{max}$  is daily maximum temperature. We  
 246 calibrated  $\text{soil}_{ip}$  from [E1];  $p_{3moist}$  and  $p_{3temp}$  from [E2];  $st_{3moist}$  and  $st_{3temp}$  from  
 247 [E3]; and  $st_{4sd\_moist}$  and  $st_{4temp}$  from [E4]. The rest of parameters were set following  
 248 Gea-Izquierdo *et al.* (2015). All parameters except  $\text{soil}_{ip}$  (which is related to the  
 249 stomatal response) help to define carbon allocation in relation to soil water content and  
 250 air temperature during the active period.

251 We calibrated these model parameters taking into account variability in functional  
 252 traits and the response to climate of plant processes related to site and species. To  
 253 address the local phenotypic response of species (Montwé *et al.* 2016), some of those  
 254 parameters ( $\leq 7$ ) described in the previous paragraph were calibrated site-wise using  
 255 maximum likelihood principles and a global optimization algorithm (Gaucherel *et al.*  
 256 2008; Gea-Izquierdo *et al.* 2015). A maximum of 7 allocation parameters from [E1] to  
 257 [E4] were optimized depending on species by comparing normalized annual integrals of



258 modeled C allocation to the stem and normalized annual growth series. We calculated  
259 different statistics to check the goodness of fit: the coefficient of determination ( $R^2$ ), the  
260 linear correlation ( $\rho$ ), and the correlation ( $r_{low}$ ) between filtered (using splines with a  
261 50% frequency cutoff of 30 years) observed and modelled growth.  $r_{low}$  was calculated to  
262 analyse the model capability to mimic the interannual and decadal growth trends. To  
263 discuss the validity of our modelling exercise and since we could not run an  
264 independent verification to the calibration conducted, coherence of the intersite  
265 multiparametric space was analysed by exploring the ecological significance of  
266 parameters compared to different site characteristics including latitude, longitude,  
267 altitude, precipitation, temperature, and Penman-Monteith potential evapotranspiration  
268 (PET), which was calculated for each site following Allen *et al.* (1998). The  
269 relationship between the 7 model parameters fitted at each of the 77 sites and the mean  
270 site ecological covariates was explored through pointwise correlations: (i) using site  
271 individual indices; (ii) using the principal components (PCs) of the 7 x 77 matrix  
272 (Legendre & Legendre 1998).

273

#### 274 *Forest performance under climate change and different $c_a$ scenarios*

275 Once the model had been calibrated we implemented forest projections at the 77  
276 sites using simulated climatic data generated under RCP2.6 and RCP8.5. To discuss the  
277 effect of  $c_a$  on the net response of forests to climate change we compared two type of  
278 forest simulations driven by the multimodel climatic scenarios:

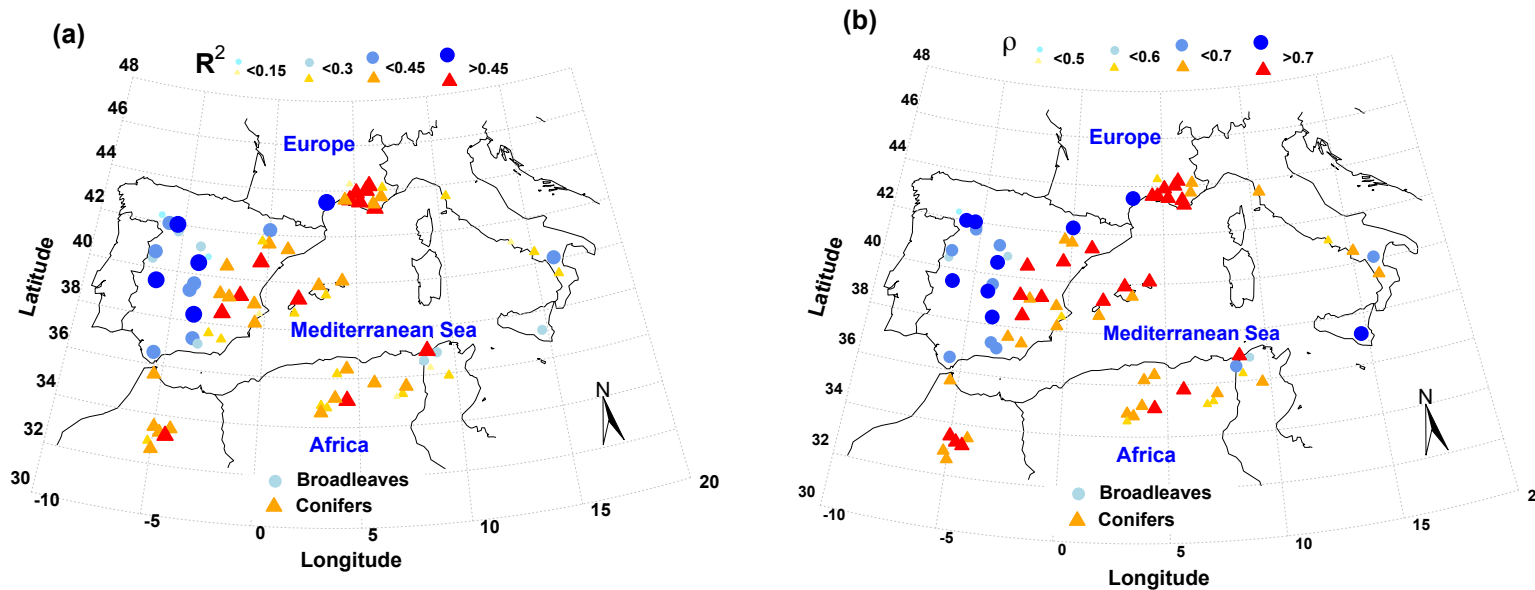
- 279 (i) ‘fertilization’ scenario: with  $c_a$  levels expected for RCP2.6 and RCP8.5.
- 280 (ii) ‘non-fertilization’ scenario: using climate from RCP2.6 and RCP8.5 but constant  $c_a$   
281 = 390 ppm after 2010.

282 We report GPP in addition to growth projections. Future growth trends were assessed

Group	n	$R^2$		r		$r_{low}$	
		Mean (sd)	Max (Min)	Mean (sd)	Max (Min)	Mean (sd)	Max (Min)
Broadleaves	22	0.343 (0.168)	0.643 (0.0)	0.675 (0.084)	0.821 (0.483)	0.728 (0.185)	0.929 (0.114)
Conifers	55	0.356 (0.141)	0.710 (0.073)	0.676 (0.070)	0.855 (0.537)	0.786 (0.156)	0.968 (0.249)
Total	77	0.353 (0.148)	0.710 (0.0)	0.675 (0.074)	0.855 (0.483)	0.769 (0.165)	0.968 (0.114)

283 Table 1. Mean values of goodness of fit statistics. n= number of forest sites;  $R^2$ =coefficient of determination; r=coefficient of correlation;

284  $r_{low}$ =coefficient of correlation of filtered data (see material and methods for details).



285

286 Fig. 2. Map showing the coefficient of determination ( $R^2$ ) and correlation ( $\rho$ ) between dendrochronological data and modeled stem growth data

287 using MAIDEN at the 77 forest sites.  $R^2$  is shown in (a) whereas  $\rho$  in (b). For  $R^2$  we split in four classes:  $0 \leq R^2 < 0.15$ ;  $0.15 \leq R^2 < 0.3$ ;

288  $0.3 \leq R^2 < 0.45$ ;  $0.45 \leq R^2$ . For  $\rho$  in:  $0.45 \leq \rho < 0.5$ ;  $0.5 \leq \rho < 0.6$ ;  $0.6 \leq \rho < 0.7$ ;  $0.7 \leq \rho$ . Triangles depict conifers, whereas circles broadleaves.

289 through the slope of simple regressions between simulated growth (or GPP) for a given  
290 scenario and site against year for the periods 2010-2100, 2010-2050 and 2051-2100.

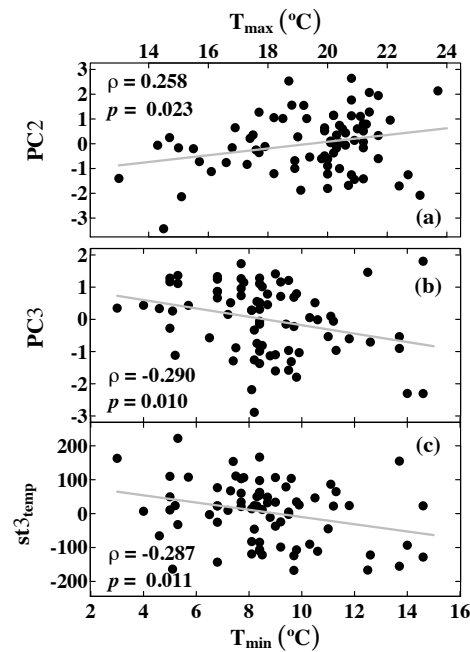
291

## 292 **Results**

### 293 *Model calibration across western Mediterranean forests*

294 The model fit against the calibration data is shown in Fig. 2 and Table 1 and the  
295 parameters fitted in App. 4. Importantly, data used to calibrate the model did not show  
296 an overall significant increase in growth, hence did not suggest evidence of a global net  
297 carbon fertilization effect during the last decades (Table 2). The observed growth trends  
298 were highly correlated with the model output ( $\rho$  and particularly  $r_{low}$  in Table 1).  
299 Correlation between model output (carbon allocated to the stem) and growth data was in  
300 average 0.67 whereas mean  $R^2$  was 0.36 (Table 1). The multiparametric space of the 77  
301 fitted models was explored through PCA. The eigenvalues corresponding to the three  
302 first principal components (PCs) were over the mean, hence significant according to the  
303 Kaiser-Guttman criterion (Legendre & Legendre 1998). These three PCs including the 7  
304 parameters fitted in the calibration phase explained 60.9 % of the variability (PC1 24.7  
305 %, PC2 19.7 % and PC3 16.6%). PC1 was mostly related to parameters linked to  
306 humidity: a positive relation with moisture parameters such as  $soil_{ip}$  and  $st_{4sd\_moist}$ , and  
307 a negative one with  $st_{3moist}$  (not shown). PC2 was mostly related to parameters linked  
308 to temperatures: positively with  $st_{4temp}$  and negatively with  $p_{3temp}$  and  $st_{3temp}$ . PC3  
309 was positively correlated with the humidity parameter  $p_{3moist}$  (not shown). The 77  
310 parameters and the PCs (2 and 3) showed some significant relationships with the site  
311 ecological characteristics (Fig.3). Most of the site-based relationships between the fitted  
312 parameters and the ecological characteristics of the 77 sites were linked to site  
313 temperatures (Fig. 3), whereas almost no significant relationships were found with the

314 other tested covariates (e.g. site precipitation or PET). These relationships suggest the  
 315 existence of some ecological coherence within the parametric space fitted for the 77  
 316 forest sites, which would support the robustness of the model parameterization used.



317 Fig. 3. Ecological coherence of model parameters at the 77 forest sites (only those  
 318 relationships that were significant are shown): (a) PC2 and site  $T_{\max}$ ; (b) PC3 and site  
 319  $T_{\min}$ ; (c)  $st3_{temp}$  and site  $T_{\min}$ .

321

### 322 *Growth-GPP projections under changing climate and $c_a$*

323 The model allocates carbon to different plant compartments driven by different non-  
 324 linear functions of environmental variability, hence it allows some decoupling between  
 325 GPP, NPP and secondary growth. In this sense, 59% and 80% of the simulations  
 326 presented correlations between GPP and growth higher than 0.5 for RCP2.6 and  
 327 RCP8.5, respectively (App. 5). Thus the majority of sites showed a good agreement  
 328 between interannual GPP and growth (i.e. generally the modeled interannual variability  
 329 of growth was driven by that of GPP). Furthermore, the growth trends (long-term,  
 330 multiannual) were of similar sign (positive, negative or neutral) as those of GPP (Table

Group	Observed past Growth	Growth projections 2010-2099							
		Allowing fertilization (i.e. predicted $c_a$ )				No fertilization (i.e. $c_a = 390$ ppm)			
		GPP		Growth		GPP		Growth	
		2010-2050	2051-2099	2010-2050	2051-2099	2010-2050	2051-2099	2010-2050	2051-2099
Broadleaves	0.79 (0.96)	0.60 c (2.03)	-0.62 c (0.38)	0.35 b (0.71)	-0.24 d (0.16)	-2.2 a (2.75)	0.26 a (0.47)	-0.65 b (0.89)	0.08 a (0.18)
Conifers	0.17 (0.71)	0.78 c (1.98)	-0.98 d (1.42)	0.14 c (0.31)	-0.08 c (0.13)	-3.64 b (3.92)	-0.02 b (0.82)	-0.34 a (0.38)	0.02 b (0.11)
Total	0.35 (0.83)	0.73 (1.98)	-0.88 (1.22)	0.20 (0.47)	-0.13 (0.16)	-3.21 (3.67)	0.06 (0.74)	-0.43 (0.58)	0.04 (0.13)
Broadleaves	0.79 (0.96)	2.11 b (2.20)	3.21 b (2.79)	0.93 a (0.96)	1.37 a (1.13)	-4.32 c (3.14)	-5.48 c (2.42)	-1.30 c (1.11)	-1.05 d (0.73)
Conifers	0.17 (0.71)	3.38 a (1.74)	3.75 a (2.51)	0.37 b (0.45)	0.78 b (0.62)	-5.68 d (3.77)	-6.4 d (3.21)	-0.61 b (0.46)	-0.57 c (0.53)
Total	0.35 (0.83)	3.02 (1.96)	3.60 (2.58)	0.53 (0.68)	0.95 (0.83)	-5.29 (3.63)	-6.11 (3.02)	-0.81 (0.77)	-0.71 (0.63)

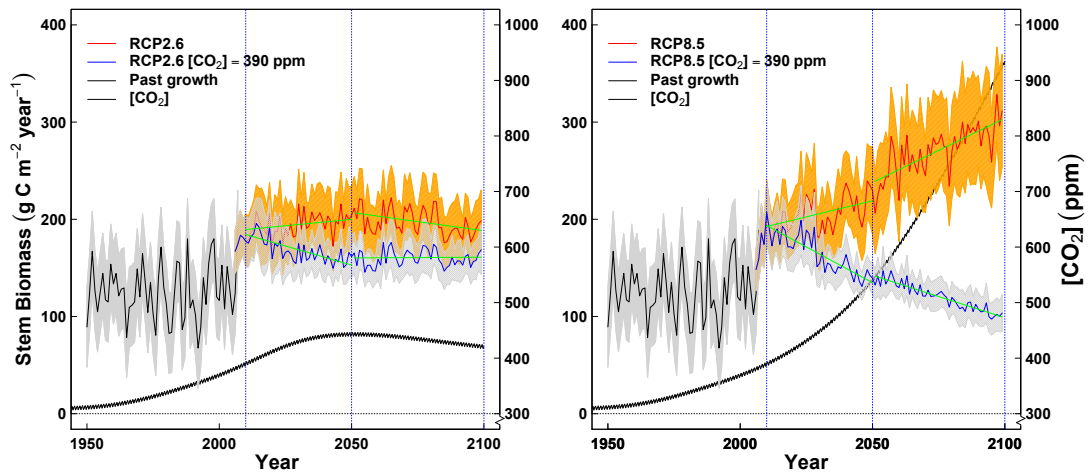
331

332 Table 2. Growth trends as estimated by the slopes of linear regressions between growth (slopes in  $\text{cm}^2 \cdot \text{year}^{-2}$ ) or GPP (slopes in  $\text{g C m}^{-2} \text{ year}^{-2}$ )

333 and year. Mean slopes are shown for observed past growth and for projected growth and GPP for the periods 2010-2050 and 2051-2099.

334 Standard deviations are between parentheses. One-way ANOVA differences between broadleaves and conifers (RCP2.6 and RCP85, i.e. 4 levels)

335 within columns are depicted with different letters.



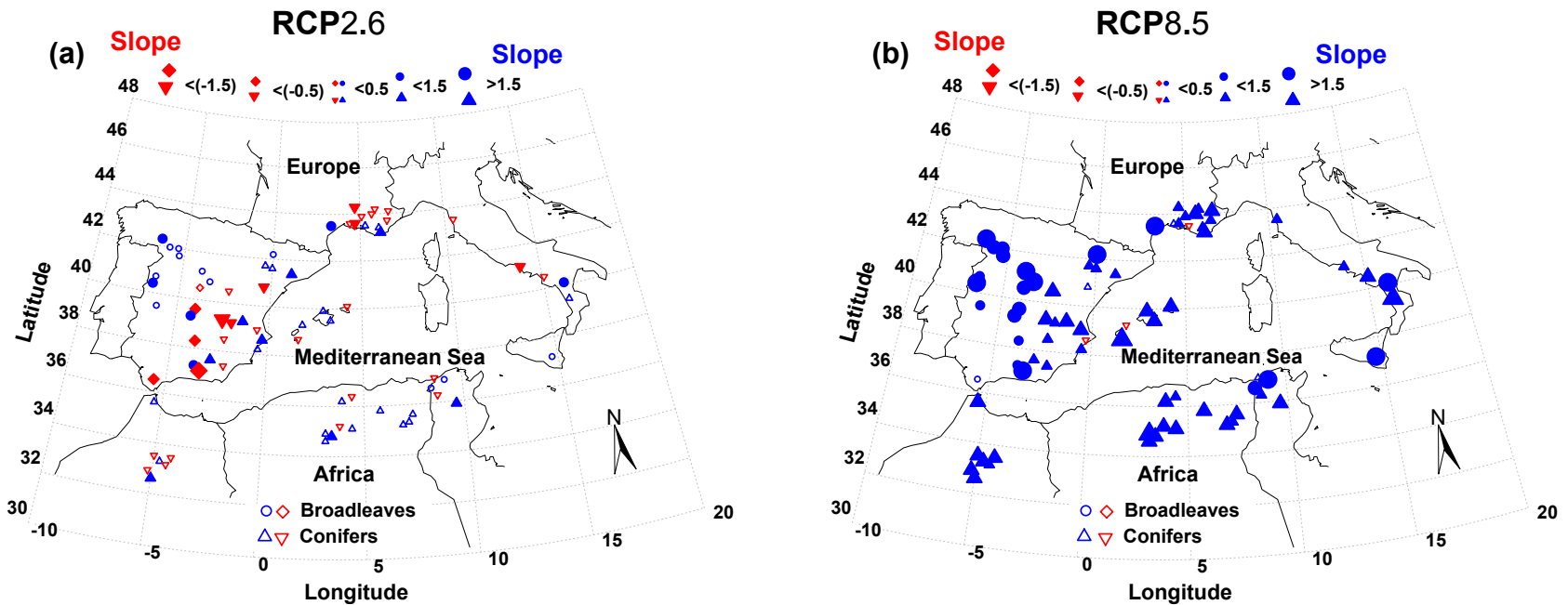
336  
 337 Fig. 4. Example of growth projections at one *Quercus pyrenaica* site (QUPY3 in  
 338 App.1). Trends (i.e. linear regressions between mean growth and year) for 2010-2050,  
 339 and 2051-2099 are shown with green lines for the ‘fertilization’ (red line) and ‘non-  
 340 fertilization’ (blue line) scenarios. These trends correspond to the slopes reported in  
 341 Table 2, Fig. 5 and Fig. 6. Shaded areas behind annual mean growth values ( $\hat{y}$ ; thick  
 342 black line is mean past growth) correspond to the confidence intervals for the mean  
 343 calculated as  $\hat{y} \pm 1.96 \cdot sd_{\hat{y}}/\sqrt{n}$  ( $sd_{\hat{y}}$  is the combined standard deviation of the model  
 344 estimates and the variability among climatic scenarios;  $n$  is sample size).  $c_a$  values  
 345 ( $[CO_2]$ ) corresponding to the two scenarios considered (i.e. RCP2.6, RCP8.5) are shown  
 346 as thin black lines.

347  
 348 2). GPP projections exhibited steeper trends (both positive and negative) for  
 349 Mediterranean conifers (all evergreen species), whereas the growth trends were steeper  
 350 for broadleaves (mostly deciduous) than for conifers (Table 2). An example of a model  
 351 simulation and how the reported trends (i.e. slopes) were calculated is depicted in Fig.  
 352 4. Model simulations yielded the greatest GPP and growth in more mesic sites, as  
 353 expected (App. 6).

354 According to our projections forest growth would not be much altered when  
355 assuming the low emission RCP2.6 scenario with predicted  $c_a$  (Fig. 5). Under the  $c_a$   
356 pathway from RCP2.6, which reaches the  $c_a$  maximum (446 ppm) in 2051, the model  
357 simulates a slight increase in growth up to 2050 followed by a slight decrease (Table 2).  
358 The resulting overall trend up to 2100 depends on site conditions: most forests exhibited  
359 non-significant or reduced trends under RCP2.6 for both  $c_a$  scenarios (Fig. 5, 6).  
360 However, for the ‘non-fertilization’ scenario, model simulations suggested significant  
361 negative growth trends for some coniferous sites e.g. in Southern France and Eastern  
362 Spain (Fig. 6). Therefore the results of the model mostly suggested that forests would  
363 acclimate at the regional scale to the climate proposed by RCP2.6. Yet, negative local  
364 impacts for some coniferous species would pop up when constraining the carbon  
365 fertilization effect.

366 Climate simulations under RCP8.5 forecast a much warmer scenario with less  
367 precipitation than RCP2.6 (App. 3). In response, forest growth projections under this  
368 scenario showed a different picture to that described for RCP2.6. For the ‘non-  
369 fertilization’ scenario (i.e. constant 390 ppm) future forest growth trends would be  
370 negative across all the western Mediterranean. Both conifers and broadleaves would  
371 suffer huge decreases in GPP and growth concurrent with the increase in PET expected  
372 under RCP8.5. These negative trends were much steeper than those for RCP2.6 (Table  
373 2) and in some cases converged towards zero. In contrast, under the coherent  $c_a$  pathway  
374 (exponential increase in  $c_a$  to 935 ppm in 2100) for RCP8.5, the model suggested that  
375 plants would not only compensate the more stressing climate but also that growth and  
376 GPP would be enhanced across the study region regardless of species (Fig. 6; Table 2).

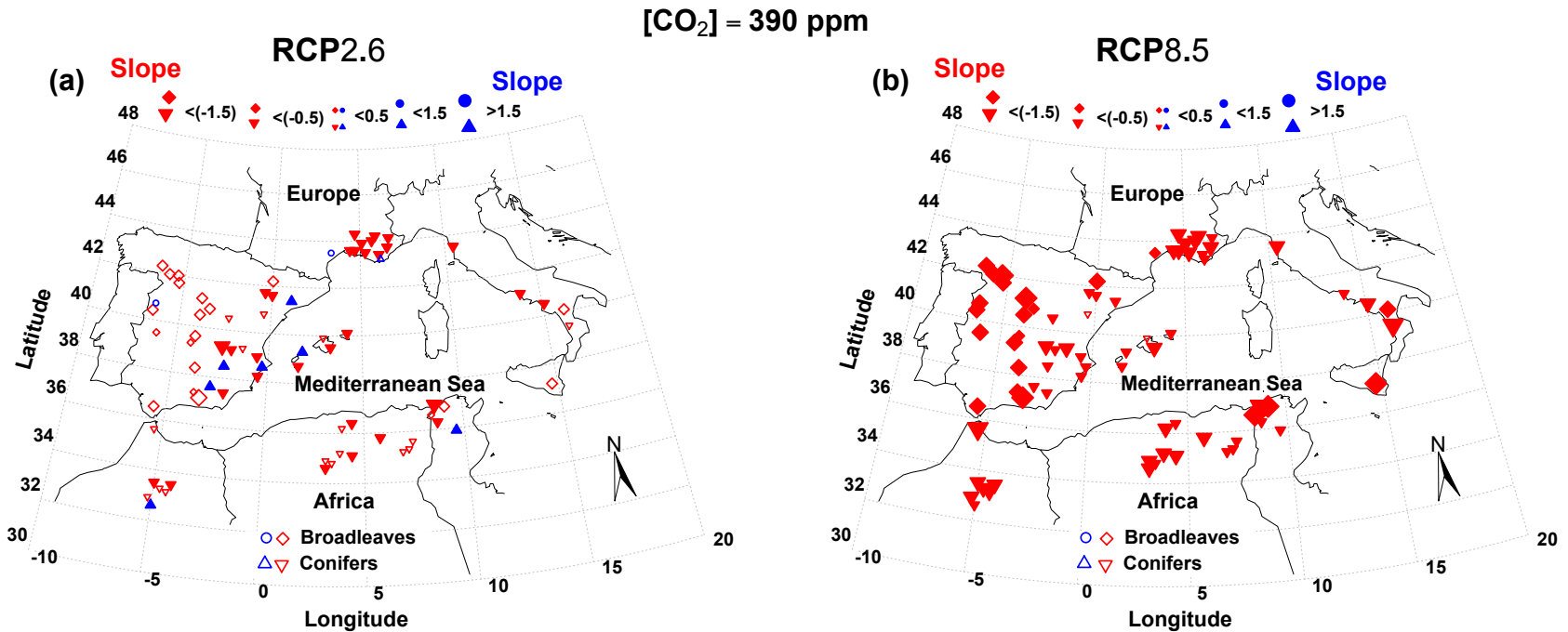
377 In average RCP8.5 predicts for the studied area in average a +2°C warmer climate  
378 (with  $c_a$  = 504 ppm) and a slight MAP reduction by 2050 (App. 3) compared to present



379

380 Fig. 5. Future growth simulations trends using climatic scenarios RCP2.6 (a) and RCP8.5 (b), with predicted (increasing)  $c_a$  between 2010 and  
 381 2099. Trends are estimated as the slopes of the linear regressions between stem biomass growth and year (see Fig. 4). Symbols are scaled as a  
 382 function of the slope value. Red symbols correspond to negative trends whereas blue symbols to positive trends. Solid symbols correspond to  
 383 significant trends ( $\alpha=0.05$ ) whereas empty symbols to non-significant trends.





384  
 385 Fig. 6. Future growth simulations trends between 2010 and 2099 using climatic scenarios RCP2.6 (a) and RCP8.5 (b), with constant  $c_a = 390$   
 386 ppm. Trends are estimated as the slopes of linear regressions between stem biomass growth and year (see Fig. 4). Symbols are scaled as a  
 387 function of the slope value. Red symbols correspond to negative trends whereas blue symbols to positive trends. Solid symbols correspond to  
 388 significant trends ( $\alpha=0.05$ ) whereas empty symbols to non-significant trends.

389 values (i.e. +2.8°C compared to preindustrial levels). This is over the reduction goal in  
390 greenhouse emissions for COP21 (<http://www.cop21paris.org/>) established below +2°C,  
391 which otherwise would be achieved in RCP2.6. Simulations under RCP8.5 suggested a  
392 negative impact in growth and GPP of climate unless this was compensated by an  
393 exaggerated fertilization effect of eCO<sub>2</sub>. The higher the temperature, the more evident  
394 and widespread this negative impact would become (Fig. 5, 6; Table 2). In contrast,  
395 when allowing fertilization, the greatest positive growth trends (i.e. a greater net  
396 fertilization effect) would arise after 2050 with c<sub>a</sub> levels >500 ppm (Table 2). Growth  
397 trends after 2050 were steeper than those before 2050 for the ‘fertilization’ scenario  
398 (mean difference 0.42, p<0.001) but not for the ‘non-fertilization’ scenario (mean  
399 difference 0.09, p=0.403). This highlights the larger influence in simulated growth and  
400 GPP of eCO<sub>2</sub> (positive) compared to that of expected high temperatures (negative).

401

## 402 **Discussion**

403 *Forest future in a warmer western Mediterranean region: what is the role of c<sub>a</sub>?*

404 Trees can enhance productivity and modify some anatomical and physiological  
405 traits (e.g. iWUE) in response to eCO<sub>2</sub> but it is not known how they will perform under  
406 future climate and c<sub>a</sub> (Medlyn & De Kauwe 2013; Duursma *et al.* 2016; Kelly *et al.*  
407 2016). A positive net effect of eCO<sub>2</sub> on trees can be hampered by the limiting effect of  
408 other environmental constraints such as nitrogen (N) availability and water stress (De  
409 Kauwe *et al.* 2013; Reichstein *et al.* 2013; Fernández-Martínez *et al.* 2014; Walker *et al.*  
410 *et al.* 2015; Kim *et al.* 2016). A positive feedback of eCO<sub>2</sub>, e.g. in leaf area (if a steady-  
411 state has not been achieved yet; Körner 2006) and NPP, has been reported under current  
412 climate conditions in temperate forests where non-climatic factors such as N availability  
413 were limiting (Medlyn *et al.* 2015; Walker *et al.* 2015; Kim *et al.* 2016). In contrast,

414 Duursma *et al.* (2016) did not observe any change in leaf area in response to eCO<sub>2</sub> when  
415 leaf area was limited by water availability (i.e. like in our model). Therefore, depending  
416 on the most limiting factor, different ecosystems seem to express different responses to  
417 eCO<sub>2</sub> under current climatic conditions.

418 As reflected by our model during the observational period (see Gea-Izquierdo *et al.*  
419 2015 for iWUE), the effect of recent rising c<sub>a</sub> has generally produced an enhancement in  
420 iWUE but not in growth rates (e.g. Peñuelas *et al.* 2011; Keenan *et al.* 2013; Saurer *et*  
421 *al.* 2014; van der Sleen *et al.* 2014). Furthermore, studies reporting past growth  
422 enhancement within the last 150 years (e.g. at species-specific high-elevation sites) did  
423 not consider c<sub>a</sub> as the main factor triggering that growth increase (Salzer *et al.* 2009;  
424 Gea-Izquierdo & Cañellas 2014). According to our results, most Mediterranean forests  
425 would mitigate the optimistic RCP2.6 scenario either with or without C-fertilization.  
426 Hence, forests would mostly endure the +2°C warming limit set within the Paris'  
427 Agreement (<http://www.cop21paris.org/>). In contrast, projections under high-emission  
428 RCP8.5 would forecast big changes in forest performance. Simulations reflected a very  
429 negative impact of climatic conditions under RCP8.5 and a non-fertilization scenario,  
430 whereas they suggested a dominant positive effect of eCO<sub>2</sub> at the regional scale  
431 (particularly for c<sub>a</sub> > 500 ppm) when allowing fertilization even under the very limiting  
432 climatic conditions of RCP8.5. The observed positive trends following a Temperature x  
433 eCO<sub>2</sub> interaction were not unexpected, because the Farquhar model ensures large direct  
434 responses to eCO<sub>2</sub> since Rubisco-limited photosynthesis responds fast to enhanced c<sub>a</sub>  
435 (Reichstein *et al.* 2013; Friend *et al.* 2014; Baig *et al.* 2015; Walker *et al.* 2015; Norby  
436 *et al.* 2016). Yet, this fertilization effect both for GPP and growth under RCP8.5 looks  
437 unrealistic (Körner 2006; Friend *et al.* 2014; Baig *et al.* 2015; Kelly *et al.* 2016).

438       The future long-term response of forests is uncertain but we expected no positive  
439 net effect of eCO<sub>2</sub> on plant growth under very strong water-limitations (van der Molen  
440 *et al.* 2011; Girardin *et al.* 2012; Baig *et al.* 2015). Nevertheless, this was only reflected  
441 for RCP8.5 by the non-fertilization scenario. Photosynthesis is a saturating function of  
442 intercellular CO<sub>2</sub> (Kelly *et al.* 2016) and according to Körner (2006) saturation is  
443 expected at levels similar to those of RCP8.5 in 2100 (circa 1000 ppm). Likely, the  
444 model does not downregulate assimilation enough under high c<sub>a</sub> or underestimates the  
445 limiting effect of other interacting factors (e.g. light, nutrients, water stress, hydraulics)  
446 on e.g. maximum carboxylation. This was addressed empirically by setting a limit at  
447 390 ppm but a more detailed understanding of the physiological processes in relation to  
448 eCO<sub>2</sub> would definitely help to improve model forecasts. A global negative response of  
449 Mediterranean forests to intense warming unless there is an exaggerated C-fertilization  
450 effect is, thus, evident in our results. Importantly, this implies negative consequences  
451 for forest performance and means that a positive effect of milder winters (e.g. earlier  
452 growing season or enhanced winter assimilation in evergreens) would not counteract the  
453 negative effect of longer stressing summers. There is an ample debate on the actual  
454 factors causing tree death, but it seems that a combination of interrelated C-related  
455 traits, hydraulically-related features and climate-related impacts of biotic agents should  
456 govern the forest decline and mortality processes (Sala *et al.* 2012; McDowell *et al.*  
457 2013; Aguiar *et al.* 2015; Anderegg *et al.* 2015). Regardless of the final causal factor/s,  
458 steep negative growth trends like those under RCP8.5 and no-fertilization strongly  
459 suggest changes in stand dynamics and composition and ultimately enhanced mortality  
460 at some sites (Bigler *et al.* 2006; van der Molen *et al.* 2011; Gea-Izquierdo *et al.* 2014).  
461  
462 *Forest growth projections under climate change and eCO<sub>2</sub>: utilities and uncertainties*

463 Models need continuous refinement to achieve robustness, reliability and realism  
464 (Prentice et al. 2015). Forecasts of tree growth are a valuable tool to understand forest  
465 performance under climate change but there are many sources of uncertainty within  
466 model performance and data-assimilation that need discussion (Friend et al. 2014).  
467 Caveats in models include: (i) current knowledge of the physiological processes; (ii)  
468 model implementation and parameterization (including scale-dependent constraints);  
469 (iii) uncertainty of model outputs outside the calibration range (Pappas *et al.* 2013;  
470 Prentice *et al.* 2015). Additionally, changes in plasticity of functional traits and plant  
471 acclimation processes can bias model projections (Muller et al. 2011). We assumed  
472 uniformitarianism (i.e. temporal invariance of the modeled relationships) in model  
473 projections but it could be that threshold related responses arise after the calibration  
474 boundaries are surpassed. In this sense (iii) is minimized by explicitly modeling  
475 processes but not eliminated as a result of (i) and (ii). Despite inherent limitations in  
476 models and observational data (Babst *et al.* 2014), by using data-assimilation we  
477 maximized the likelihood of getting unbiased past long-term trends to increase  
478 reliability of growth projections (Peng *et al.* 2011; Medlyn *et al.* 2015). Forest  
479 productivity was analyzed assuming present steady-state stand conditions (e.g.  
480 composition, leaf area, root mass) constrained by water stress (Körner 2006). This could  
481 bias the analysis of forest dynamics (Körner 2003; Friend et al. 2014; Pappas *et al.*  
482 2015), particularly for mixed stands (Pappas et al. 2013). However, our aim was to  
483 analyze performance of the present stands under future environmental conditions, with  
484 emphasis on species long-term trends. Changes in inter-species dynamics are away from  
485 the scope of our analysis and need to be studied complementary.

486 The model does not include nutrient dynamics but focuses on the water and carbon  
487 cycles and the effect of water stress at different functional levels. This is because

488 nutrient availability is generally considered secondary in Mediterranean ecosystems  
489 compared to water stress, which in addition is expected to increase in the future (Giorgi  
490 & Lionello 2008; García-Ruiz *et al.* 2011; IPCC 2014). Thus, the limiting effect of  
491 factors such as availability of N, P, or hydraulic constraints could invalidate a C  
492 fertilization effect on net growth under eCO<sub>2</sub> (Körner 2006; Norby *et al.* 2010; Fatichi  
493 *et al.* 2014; Fernández *et al.* 2014; Friend *et al.* 2014; Baig *et al.* 2015). Our model is  
494 simpler than ecosystem models including nutrient and stand dynamics (e.g. Reichstein  
495 *et al.* 2013; Walker *et al.* 2015). However, our scale is finer (stand, species-specific  
496 compared to PFTs) and, most important, it is driven by actual growth data to ensure  
497 unbiased estimation of short- and long-term trends. Other factors such as differences in  
498 carry-over effects between conifers and broadleaves, changes in species composition  
499 and demography, competition and tree-related traits such as ontogeny and size, partly  
500 modulate the forest response to climate (van der Molen *et al.* 2011). However, long-  
501 term stand productivity seems to change slightly under moderate disturbances such as  
502 those produced by silvicultural treatments or insect infestation (Vesala *et al.* 2005;  
503 Amiro *et al.* 2010). Therefore, we believe that climate effects are dominant and the  
504 reported long-term trends are robust in relation to variability within these other factors,  
505 which would affect mostly in the short-term. The model fit reported was in the range of  
506 that in similar studies (Misson 2004; Li *et al.* 2014; Gea-Izquierdo *et al.* 2015; Girardin  
507 *et al.* 2016). Different goodness-of-fit at different sites could result on differences in  
508 model performance. However, in App. 7A we show how variability in R<sup>2</sup> did not  
509 influence the estimated trends (future projections). Particularly, when the trees exhibited  
510 long-term trends in the past (observational period), these showed a very high agreement  
511 with the modeled growth trends (App. 7B; Table 1). As mentioned, the robustness of  
512 our approach relies on the spatial (regional) scale where the model was applied, which

513 expands model implementation at a broader scale than that where it was previously  
514 applied (Misson *et al.* 2004; Gaucherel *et al.* 2008; Gea-Izquierdo *et al.* 2015).

515 Model simulations can be improved by better addressing the influence of CO<sub>2</sub> on  
516 variability of different plant traits (Kattge *et al.* 2011; Atkin *et al.* 2015; Pappas *et al.*  
517 2016). Leaf photosynthesis is negatively affected by drought through several  
518 mechanisms including changes in stomatal and mesophyll conductance or reductions in  
519 biochemical efficiency (Flexas *et al.* 2005). Both photosynthesis and stomatal  
520 conductance were modeled as direct functions of CO<sub>2</sub>. However, warming and  
521 increasing water stress in relation to eCO<sub>2</sub> could differently affect autotrophic  
522 respiration and photosynthetic capacity, or enhance photorespiration more than  
523 photosynthesis (Baig *et al.* 2015; Girardin *et al.* 2016; Rowland *et al.* 2015; Varone &  
524 Gratani 2015). Our model takes into account the short-term acclimation of leaf  
525 photosynthesis and stomatal conductance to CO<sub>2</sub>, while respiration is set as a function  
526 of temperature and GPP (Gea-Izquierdo *et al.* 2015). Different sources of interspecific  
527 variability under eCO<sub>2</sub> in autotrophic respiration not explained by models (Atkin *et al.*  
528 2015) or other factors such as species-specific variability in  $J_{\max}/V_{c_{\max}}$  could partly  
529 impair our results. Trees modify other traits such as leaf area and sapwood area to  
530 withstand xericity (Breda *et al.* 2006; Martin-StPaul *et al.* 2013; Duursma *et al.* 2016;  
531 Kelly *et al.* 2016). In our model, intra-annual, inter-annual and long-term structural  
532 acclimation of leaf area and allocation rules rely on climate and SWC but not on CO<sub>2</sub>.  
533 Thus, addressing the influence of CO<sub>2</sub> on leaf area dynamics by setting SLA and  
534 allocation rules also as functions of CO<sub>2</sub> could also help to better assess whether the  
535 reported fertilization effect is unrealistic (Duursma *et al.* 2016; Medlyn *et al.* 2015).

536 In summary, modeled forest growth reflected the observed absence of an overall net  
537 positive effect of enhanced  $c_a$  under increased temperatures (i.e. PET) in the recent past.

538 According to model projections, western Mediterranean forests would mostly mitigate  
539 the negative effect of a climate remaining below the maximum warming levels (+2°C)  
540 agreed in COP21 (i.e. scenario RCP2.6) but the situation would be very different above  
541 those levels (as represented by RCP8.5). Our results suggest that fertilization could  
542 override the negative effect of stressing climatic conditions under high-emission  
543 RCP8.5 but this fertilization effect of eCO<sub>2</sub> looks unrealistic according to the literature,  
544 being most likely a result of miss-performance of models way above the calibration c<sub>a</sub>  
545 levels. Consequently, simulations precluding fertilization under high-emission scenarios  
546 show very negative forest performance at the regional scale in the future for both  
547 conifers and broadleaves. Our results suggest that western Mediterranean forests would  
548 not resist the stressing conditions of a much warmer climate unless species exhibited an  
549 exaggerated C fertilization effect. It is necessary to include c<sub>a</sub> variability in forest  
550 models but it is not enough. We still need a better understanding of the physiological  
551 processes governing the capacity of acclimation of different plant traits (e.g. V<sub>cmax</sub>) to  
552 the interaction between water stress, eCO<sub>2</sub> and nutrient availability. In this sense, our  
553 simulations precluding a fertilization effect seems more realistic than those allowing  
554 fertilization under c<sub>a</sub> levels way above those used to calibrate models. Our study  
555 provides a comprehensive data-driven analysis of the likely performance of western  
556 Mediterranean forests under predicted climate change and c<sub>a</sub>. Yet model performance  
557 still needs to be refined under high c<sub>a</sub> as expected in the future.

558

## 559 **Acknowledgements**

560 This study was funded by project AGL2014-61175-JIN of the Spanish Ministry of  
561 Economy and Competitiveness and the Labex OT-Med (n° ANR-11-LABEX-0061)  
562 from the «Investissements d’Avenir» program of the French National Research Agency



563 through the A\*MIDEX project (n° ANR-11-IDEX-0001-02). The climate simulations  
564 have been extracted from the CMIP5 site ([https://esgf-index1.ceda.ac.uk/projects/esgf-  
ceda/](https://esgf-index1.ceda.ac.uk/projects/esgf-<br/>565 ceda/)) and processed with the help of Romain Suarez.

566

## 567 **References**

568 Aguadé D, Poyatos R, Gómez M, Oliva J, Martínez-Vilalta J (2015) The role of  
569 defoliation and root rot pathogen infection in driving the mode of drought-related  
570 physiological decline in Scots pine (*Pinus sylvestris* L.). *Tree Physiology*, **35**, 229–  
571 242.

572 Allen RG, Pereira LS, Raes D, Smith M (1998) *Crop Evapotranspiration – Guidelines  
573 for Computing Crop Water Requirements*, FAO Irrigation and Drainage Paper 56,  
574 FAO, 1998, ISBN 92-5-104219-5.

575 Amiro BD, Barr AG, Barr JG *et al.* (2010) Ecosystem carbon dioxide fluxes after  
576 disturbance in forests of North America. *Journal of Geophysical Research:  
577 Biogeosciences*, **115**, G00K02, doi:[10.1029/2010JG001390](https://doi.org/10.1029/2010JG001390).

578 Anderegg WRL, Schwalm C, Biondi F *et al.* (2015) Pervasive drought legacies in forest  
579 ecosystems and their implications for carbon cycle models. *Science*, **1**, 528–532.

580 Atkin OK, Bloomfield KJ, Reich PB *et al.* (2015) Global variability in leaf respiration  
581 in relation to climate, plant functional types and leaf traits. *New Phytologist*, **206**,  
582 614–636.

583 Babst F, Alexander MR, Szejner P *et al.* (2014) A tree-ring perspective on the terrestrial  
584 carbon cycle. *Oecologia*, **176**, 307–322.

585 Baig S, Medlyn BE, Mercado LM, Zaehle S (2015) Does the growth response of woody  
586 plants to elevated CO<sub>2</sub> increase with temperature? A model-oriented meta-analysis.  
587 *Global Change Biology*, **21**, 4303–4319.

588 Baldocchi DD, Ma SY, Rambal S *et al.* (2010) On the differential advantages of  
589 evergreenness and deciduousness in mediterranean oak woodlands: a flux  
590 perspective. *Ecological Applications*, **20**, 1583–1597.

591 Battipaglia G, Saurer M, Cherubini P, Calfapietra C, McCarthy HR, Norby RJ, Cotrufo  
592 MF (2013) Elevated CO<sub>2</sub> increases tree-level intrinsic water use efficiency:  
593 insights from carbon and oxygen isotope analyses in tree rings across three forest  
594 FACE sites. *New Phytologist*, **197**, 544-554.

595 Bigler C, Braker OU, Bugmann H, Dobbertin M, Rigling A (2006) Drought as an  
596 inciting mortality factor in Scots pine stands of the Valais, Switzerland.  
597 *Ecosystems*, **9**, 330–343.

598 Breda N, Huc R, Granier A, Dreyer E (2006) Temperate forest trees and stands under  
599 severe drought: a review of ecophysiological responses, adaptation processes and  
600 long-term consequences. *Annals of Forest Science*, **63**, 625–644.

601 De Kauwe MG, Medlyn BE, Zaehle S *et al.* (2013) Forest water use and water use  
602 efficiency at elevated CO<sub>2</sub>: a model-data intercomparison at two contrasting  
603 temperate forest FACE sites. *Global Change Biology*, **19**, 1759–1779.

604 Duursma RA, Gimeno TE, Boer MM, Crous KY, Tjoelker MG, Ellsworth DS (2016)  
605 Canopy leaf area of a mature evergreen Eucalyptus woodland does not respond to  
606 elevated atmospheric [CO<sub>2</sub>] but tracks water availability. *Global Change Biology*,  
607 **22**, 1666-1676.

608 Falge E, Baldocchi D, Tenhunen J *et al.* (2002) Seasonality of ecosystem respiration  
609 and gross primary production as derived from FLUXNET measurements.  
610 *Agricultural and Forest Meteorology*, **113**, 53–74.

611 Farquhar GD, von Caemmerer S, Berry JA (1980) A biochemical model of  
612 photosynthetic CO<sub>2</sub> assimilation in leaves of C<sub>3</sub> species, *Planta*, **149**, 78–90.

613 Fatichi S, Leuzinger S, Körner C (2014) Moving beyond photosynthesis : from carbon  
614 source to sink-driven vegetation modeling. *New Phytologist*, **201**, 1086–1095.

615 Fatichi S, Pappas C, Ivanov VY (2015) Modeling plant-water interactions: an  
616 ecohydrological overview from the cell to the global scale. *Wiley Interdisciplinary*  
617 *Reviews Water* 3, 327–368.

618 Fernández-Martínez M, Vicca S, Janssens IA et al. (2014) Nutrient availability as the  
619 key regulator of global forest carbon balance. *Nature Climate Change*, **4**, 471–476.

620 Flexas J, Bota J, Galmes J, Medrano H, Ribas-Carbo M (2006) Keeping a positive  
621 carbon balance under adverse conditions: responses of photosynthesis and  
622 respiration to water stress. *Physiologia Plantarum*, **127**, 343–352.

623 Friend AD, Lucht W, Rademacher TT et al. (2014) Carbon residence time dominates  
624 uncertainty in terrestrial vegetation responses to future climate and atmospheric  
625 CO<sub>2</sub>. *Proceedings of the Natural Academy of Sciences of the United States of*  
626 *America*, **111**, 3280–5.

627 Fritts HC (1976) *Tree Rings and Climate*. Blackburn Press, p. 567.

628 García-Ruiz JM, Ignacio Lopez-Moreno J, Vicente-Serrano SM, Lasanta-Martinez T,  
629 Begueria S (2011) Mediterranean water resources in a global change scenario.  
630 *Earth-Science Reviews*, **105**, 121–139.

631 Gaucherel C, Guiot J, Misson L (2008) Evolution of the potential distribution area of  
632 french mediterranean forests under global warming. *Biogeosciences*, **5**, 1493–  
633 1504.

634 Gea-Izquierdo G, Cañellas I. (2014) Contrasting instability in growth trends of  
635 Mediterranean oaks at opposite distributional limits. *Ecosystems*, **17**, 228-241.

636 Gea-Izquierdo G, Viguera B, Cabrera M, Cañellas I (2014) Drought induced decline  
637 could portend widespread pine mortality at the xeric ecotone in managed  
638 mediterranean pine-oak woodlands. *Forest Ecology and Management*, **320**, 70–82.

639 Gea-Izquierdo G, Guibal F, Joffre R, Ourcival J-M, Simioni G, Guiot J. (2015)  
640 Modelling the climatic drivers determining photosynthesis and carbon allocation in  
641 evergreen Mediterranean forests using multiproxy long time series.  
642 *Biogeosciences*, **12**, 3695-3712.

643 Giorgi F, Lionello P (2008) Climate change projections for the Mediterranean region.  
644 *Global and Planetary Change*, **63**, 90–104.

645 Girardin MP, Bernier PY, Raulier F, Tardif JC, Conciatori F, Guo XJ (2012) Testing for  
646 a CO<sub>2</sub> fertilization effect on growth of Canadian boreal forests. *Journal of*  
647 *Geophysical Research-Biogeosciences*, **116**, G01012, doi:[10.1029/2010JG001287](https://doi.org/10.1029/2010JG001287).

648 Girardin MP, Hogg EH, Bernier PY, Kurz WA, Guo XJ, Cyr G (2016) Negative  
649 impacts of high temperatures on growth of black spruce forests intensify with the  
650 anticipated climate warming. *Global Change Biology*, **22**, 627–643.

651 Guiot J, Boucher E, Gea-Izquierdo G (2014) Process models and model-data fusion in  
652 dendroecology. *Frontiers in Ecology and Evolution*, **2**, 1–12.

653 Herrera S, Gutierrez JM, Ancell R, Pons MR, Frias MD, Fernandez J (2012)  
654 Development and Analysis of a 50 year high-resolution daily gridded precipitation  
655 dataset over Spain (Spain02). *International Journal of Climatology*, **32**, 74-85.

656 IPCC (2014) Climate Change 2014: Synthesis Report. Contribution of Working Groups  
657 I, II and III to the Fifth Assessment Report of the Intergovernmental Panel on  
658 Climate Change [Core Writing Team, R.K. Pachauri and L.A. Meyer (eds.)].  
659 IPCC, Geneva, Switzerland, 151 pp.

660 Kattge J, Díaz S, Lavorel S *et al.* (2011) TRY - a global database of plant traits. *Global*  
661 *Change Biology*, **17**, 2905–2935.

662 Keenan TF, Hollinger DY, Bohrer G, Dragoni D, Munger JW, Schmid HP, Richardson  
663 AD (2013) Increase in forest water-use efficiency as atmospheric carbon dioxide  
664 concentrations rise. *Nature*, **499**, 324–7.

665 Kelly JWG, Duursma RA, Atwell BJ, Tissue DT, Medlyn BE (2016) Drought x CO<sub>2</sub>  
666 interactions in trees: A test of the low-intercellular CO<sub>2</sub> concentration (C<sub>i</sub>)  
667 mechanism. *New Phytologist*, **209**, 1600–1612.

668 Kim D, Oren R, Qian SS (2016) Response to CO<sub>2</sub> enrichment of understory vegetation  
669 in the shade of forests. *Global Change Biology*, **22**, 944–956.

670 Körner (2003) Slow in, rapid out - Carbon flux studies and Kyoto targets. *Science*, **300**,  
671 1242–1243 ST – Slow in, rapid out – Carbon flux s.

672 Körner C (2006) Plant CO<sub>2</sub> responses: an issue of definition , time and resource supply.  
673 *New Phytologist*, **172**, 393–411.

674 Legendre P, Legendre L (1998) *Numerical Ecology*. 853 pp. Elsevier, Amsterdam.

675 Leuning R (1995) A critical appraisal of a combined stomatal- photosynthesis model for  
676 C<sub>3</sub> plants. *Plant, Cell & Environment*, **18**, 339–355.

677 Li G, Harrison SP, Prentice IC, Falster D (2014) Simulation of tree ring-widths with a  
678 model for primary production, carbon allocation and growth. *Biogeosciences*, **11**,  
679 6711–6724.

680 Luysaert S, Inglima I, Jung M *et al.* (2007) CO<sub>2</sub> balance of boreal, temperate, and  
681 tropical forests derived from a global database. *Global Change Biology*, **13**, 2509–  
682 2537.

683 Martin-Stpaul NK, Limousin J-M, Vogt-Schilb H, Rodríguez-Calcerrada J, Rambal S,  
684 Longepierre D, Misson L (2013) The temporal response to drought in a

685 Mediterranean evergreen tree: comparing a regional precipitation gradient and a  
686 throughfall exclusion experiment. *Global change biology*, **19**, 2413–26.

687 McDowell NG, Fisher RA, Xu C *et al.* (2013) Evaluating theories of drought-induced  
688 vegetation mortality using a multimodel – experiment framework. *New*  
689 *Phytologist*, **200**, 304–321.

690 Mediavilla S, Garcia-Ciudad a., Garcia-Criado B, Escudero a. (2008) Testing the  
691 correlations between leaf life span and leaf structural reinforcement in 13 species  
692 of European Mediterranean woody plants. *Functional Ecology*, **22**, 787–793.

693 Medlyn B, De Kauwe B (2013) Carbon dioxide and water use in forests. *Nature*, **499**,  
694 20–22.

695 Medlyn BE, Zaehle S, De Kauwe MG *et al.* (2015) Using ecosystem experiments to  
696 improve vegetation models. *Nature Climate Change*, **5**, 528–534.

697 Misson L (2004) MAIDEN: a model for analyzing ecosystem processes in  
698 dendroecology. *Canadian Journal of Forest Research*, **34**, 874–887.

699 Montwé D, Isaac-Renton M, Hamann A, Spiecker H (2016) Drought tolerance and  
700 growth in populations of a wide-ranging tree species indicate climate change risks  
701 for the boreal north. *Global Change Biology*, **22**, 806–815.

702 Morales P, Sykes MT, Prentice IC *et al.* (2005) Comparing and evaluating process-  
703 based ecosystem model predictions of carbon and water fluxes in major European  
704 forest biomes. *Global Change Biology*, **11**, 2211–2233.

705 Muller B, Pantin F, Génard M, Turc O, Freixes S, Piques M, Gibon Y (2011) Water  
706 deficits uncouple growth from photosynthesis, increase C content, and modify the  
707 relationships between C and growth in sink organs. *Journal of Experimental*  
708 *Botany*, **62**, 1715–1729.

709 Niinemets U (2010) Responses of forest trees to single and multiple environmental  
710 stresses from seedlings to mature plants: Past stress history, stress interactions,  
711 tolerance and acclimation. *Forest Ecology and Management*, **260**, 1623–1639.

712 Norby RJ, Warren JM, Iversen CM, Medlyn BE, McMurtrie RE (2010) CO<sub>2</sub>  
713 enhancement of forest productivity constrained by limited nitrogen availability.  
714 *Proceedings of the National Academy of Sciences USA*, **107**, 19368–19373.

715 Norby RJ, De Kauwe MG, Domingues TF *et al.* (2016) Model-data synthesis for the  
716 next generation of forest free-air CO<sub>2</sub> enrichment (FACE) experiments. *New*  
717 *Phytologist*, **209**, 17–28.

718 Pappas C, Fatichi S, Leuzinger S, Wolf A, Burlando P (2013) Sensitivity analysis of a  
719 process-based ecosystem model: Pinpointing parameterization and structural  
720 issues. *Journal of Geophysical Research: Biogeosciences*, **118**, 505–528.

721 Pappas C, Fatichi S, Rimkus S, Burlando P, Huber M (2015) The role of local scale  
722 heterogeneities in terrestrial ecosystem modeling. *Journal of Geophysical*  
723 *Research: Biogeosciences*, **120**, 341–360.

724 Pappas C, Fatichi S, Burlando P (2016) Modeling terrestrial carbon and water dynamics  
725 across climatic gradients: Does plant trait diversity matter? *New Phytologist*, **209**,  
726 137–151.

727 Peng CH, Guiot J, Wu HB, Jiang H, Luo YQ (2011) Integrating models with data in  
728 ecology and palaeoecology: advances towards a model-data fusion approach.  
729 *Ecology Letters*, **14**, 522–536.

730 Peñuelas J, Canadell JG, Ogaya R (2011) Increased water-use efficiency during the 20th  
731 century did not translate into enhanced tree growth. *Global Ecology and*  
732 *Biogeography*, **20**, 597–608.

733 Potter KA, Arthur Woods H, Pincebourde S (2013) Microclimatic challenges in global  
734 change biology. *Global Change Biology*, **19**, 2932–2939.

735 Prentice IC, Liang X, Medlyn BE, Wang YP (2015) Reliable, robust and realistic: The  
736 three R's of next-generation land-surface modelling. *Atmospheric Chemistry and*  
737 *Physics*, **15**, 5987–6005.

738 Reichstein M, Bahn M, Ciais P *et al.* (2013) Climate extremes and the carbon cycle.  
739 *Nature*, **500**, 287–295.

740 Rowland L, Lobo-do-Vale RL, Christoffersen BO *et al.* (2015) After more than a  
741 decade of soil moisture deficit, tropical rainforest trees maintain photosynthetic  
742 capacity, despite increased leaf respiration. *Global Change Biology*, **21**, 4662–  
743 4672.

744 Sala A, Woodruff DR, Meinzer FC (2012) Carbon dynamics in trees: feast or famine?  
745 *Tree physiology*, **32**, 764–75.

746 Salzer MG, Hughes MK, Bunn AG, Kipfmüller KF (2009) Recent unprecedented tree-  
747 ring growth in bristlecone pine at the highest elevations and possible causes.  
748 *Proceedings of the National Academy of Sciences USA*, **106**, 20348–20353.

749 Saurer M, Spahni R, Frank DC *et al.* (2014) Spatial variability and temporal trends in  
750 water-use efficiency of European forests. *Global change biology*, **20**, 3700–3712.

751 Schaefer K, Schwalm CR, Williams C *et al.* (2012) A model-data comparison of gross  
752 primary productivity: Results from the North American Carbon Program site  
753 synthesis. *Journal of Geophysical Research*, **117**, G03010,  
754 doi:10.1029/2012JG001960.

755 Sheffield J, Goteti G, Wood EF (2006) Development of a 50-yr high-resolution global  
756 dataset of meteorological forcings for land surface modeling. *Journal of Climate*,  
757 **19** (13), 3088-3111.



758 Stephenson NL, Das a J, Condit R *et al.* (2014) Rate of tree carbon accumulation  
759 increases continuously with tree size. *Nature*, **507**, 90–3.

760 van Der Molen MK, Dolman AJ, Ciais P *et al.* (2011) Drought and ecosystem carbon  
761 cycling. *Agricultural and Forest Meteorology*, **151**, 765–773.

762 van der Sleen P, Groenendijk P, Vlam M *et al.* (2014) No growth stimulation of tropical  
763 trees by 150 years of CO<sub>2</sub> fertilization but water-use efficiency increased. *Nature*  
764 *Geoscience*, **8**, 24–28.

765 van Vuuren DP, Edmonds J, Kainuma M *et al.* (2011) The representative concentration  
766 pathways: An overview. *Climatic Change*, **109**, 5–31.

767 Varone L, Gratani L (2015) Leaf respiration responsiveness to induced water stress in  
768 Mediterranean species. *Environmental and Experimental Botany*, **109**, 141–150.

769 Vesala T, Suni T, Rannik Ü *et al.* (2005) Effect of thinning on surface fluxes in a  
770 boreal forest. *Global Biogeochemical Cycles*, **19**, 1–11.

771 Walker AP, Zaehle S, Medlyn BE *et al.* (2015) Predicting long-term carbon  
772 sequestration in response to CO<sub>2</sub> enrichment: How and why do current ecosystem  
773 models differ? *Global Biogeochemical Cycles*, **29**, 476–495.

774 App. 1. List of chronologies used and their source. Lat=latitude; Long=longitude; Altit=altitude; MAP = mean annual precipitation; MAT= mean  
 775 annual temperature (°C); PET<sub>pm</sub>=Penman-Monteith annual evapotranspiration; P=annual precipitation.

#	Site	Species	Country	Lat (°)	Long (°)	Altit (m)	MAP (mm)	MAT (°C)	PET <sub>pm</sub> (mm)	P-PET (mm)	P/PET	Source
1	QUPY2			42.2	-6.7	1300	985	9.65	962.4	22.6	1.02	
2	QUPY3			41.9	-6.2	760	453.9	12.3	1108.9	-655.0	0.41	
3	QUPY4			40.3	-6.8	900	1135.7	14.1	1213.3	-77.6	0.94	
4	QUPY5			40.7	-3.7	1300	580	10.1	1017.0	-437.0	0.57	
6	QUPY7	<i>Quercus pyrenaica</i> Willd.	Spain	39.5	-4.3	900	496.5	14.3	1228.9	-732.4	0.40	Gea-Izquierdo <i>et al.</i> (2014)
6	QUPY8			38.2	-4.1	890	474	15.0	1356.4	-882.4	0.35	
7	QUPY9			37.2	-4.0	1486	509.8	11.3	1164.8	-655.0	0.44	
8	QUPY10			37.0	-3.7	1550	619.3	9.8	984.5	-365.2	0.63	
9	QUPY11			40.8	-4.2	1056	599.9	10.1	1001.8	-401.9	0.6	
10	QUIL1			41.6	-5.6	740	433.8	12.5	1139.7	-705.9	0.38	
11	QUIL2			40.6	-6.7	700	562.6	12.9	1220.6	-658.0	0.46	
12	QUIL3	<i>Quercus ilex</i> L.	Spain	40.4	-4.2	600	516.6	13.3	1169.5	-652.9	0.44	Gea-Izquierdo <i>et al.</i> (2011)
13	QUIL4			39.4	-6.4	390	544.1	16.1	1269.4	-725.3	0.43	
14	QUIL5			43.2	3.0	270	1009.1	12.8	1001.5	7.6	1.01	
15	QUFG1			41.9	-5.7	680	453.9	12.3	1107.8	-653.9	0.41	Gea-Izquierdo (Unpublished data)
16	QUFG2	<i>Quercus faginea</i> Lam.	Spain	39.2	-4.5	900	496.5	14.3	1231.7	-735.2	0.40	
17	QUFG3			41.5	-0.3	550	326.1	15.3	1272.3	-946.2	0.26	<a href="https://www.ncdc.noaa.gov/paleo/study/10475">https://www.ncdc.noaa.gov/paleo/study/10475</a>
18	QUPU1	<i>Quercus pubescens</i> Willd.	Italy	37.1	14.4	430	537.1	15.8	964.2	-427.1	0.56	Garfi (2000)
19	QUCE1	<i>Quercus cerris</i> L.	Italy	40.4	15.8	590	767.3	11.2	878.5	-111.2	0.87	Battipaglia (unpublished data)
20	QUCA1			36.4	-5.9	330	988.4	16.1	1195.4	-207.0	0.83	
21	QUCA5	<i>Quercus canariensis</i> Willd.	Spain	36.8	8.8	758	649.3	16.5	1319	-669.7	0.49	Gea-Izquierdo <i>et al.</i> (2010)

22	QUCA6			36.5	8.1	760	625.2	15.9	1338.8	-713.6	0.47			
23	PINI2	<i>Pinus nigra</i> J.F. Arnold	Algeria	36.3	4.1	1560	580	14.8	1201.0	-621.0	0.48	Touchan <i>et al.</i> (2011)		
24	PINI5	<i>Pinus heldreichi</i> H.Christ	Italy	39.3	15.9	1430	806.1	12.4	901.1	-95.0	0.89	<a href="https://dendrodb.eccorev.fr/framedb.htm">https://dendrodb.eccorev.fr/framedb.htm</a>		
25	PIPI2	<i>Pinus pinaster</i> Ait.	Morocco	35.5	-5.7	900	611.3	16.4	1156.4	-545.1	0.53	Touchan <i>et al.</i> (2011)		
26	PIPNI			40.4	-2.6	1055	443.3	11.4	1217.1	-773.8	0.36	<a href="https://www.ncdc.noaa.gov/paleo/study/2863">https://www.ncdc.noaa.gov/paleo/study/2863</a>		
27	PIPNI2		Spain	39.1	-1.7	705	379.8	14.0	1269.4	-889.6	0.30	<a href="https://www.ncdc.noaa.gov/paleo/study/2866">https://www.ncdc.noaa.gov/paleo/study/2866</a>		
28	PIPNI3	<i>Pinus pinea</i> L.		39.2	-2.3	720	264.3	14.1	1330.4	-1066.1	0.20	<a href="https://www.ncdc.noaa.gov/paleo/study/2867">https://www.ncdc.noaa.gov/paleo/study/2867</a>		
29	PIPNI4			39.2	-2.8	700	343.3	14.5	1347.4	-1004.1	0.25	<a href="https://www.ncdc.noaa.gov/paleo/study/2865">https://www.ncdc.noaa.gov/paleo/study/2865</a>		
30	PIPNI8		Italy		43.4	10.2	10	962.1	15.3	1019.4	-57.3	0.94	<a href="https://www.ncdc.noaa.gov/paleo/study/16755">https://www.ncdc.noaa.gov/paleo/study/16755</a>	
31	PIPNI9				41.0	13.6	9	571.8	17.7	974.1	-402.3	0.59	Battipaglia <i>et al.</i> (2016)	
32	PIHA1		France	43.1	5.9	420	645.9	14.7	1201.2	-555.3	0.54	Gea-Izquierdo <i>et al.</i> (2015)		
33	PIHA2			36.9	8.3	23	630.3	18.7	1283.0	-652.7	0.49	Gea-Izquierdo (unpublished data)		
34	PIHA3		Tunisia	36.2	8.4	950	619.8	15.1	1297.7	-677.9	0.48			
35	PIHA4			35.8	9.3	800	542.3	14.7	1322.1	-779.8	0.41	Touchan <i>et al.</i> (2011)		
36	PIHA5			34.8	2.8	1380	349.6	14.5	1315.3	-965.7	0.27			
37	PIHA8			35.2	6.9	1300	380.8	14.6	1330.3	-949.5	0.29			
38	PIHA9			35.7	5.5	1200	431.2	15.6	1292.6	-861.4	0.33			
39	PIHA10			35.3	7.1	1650	380.8	14.6	1333.9	-953.1	0.29			
40	PIHA11	<i>Pinus halepensis</i> Mill.	Algeria	35.0	4.1	1100	390.2	15.6	1304.3	-914.1	0.30	Safar (1994), Safar <i>et al.</i> (1992)		
41	PIHA12				35.1	3.5	1060	417.2	15.4	1304.8	-887.6		0.32	
42	PIHA13				34.7	3.1	1410	350.8	14.3	1305.5	-954.7		0.27	
43	PIHA14				34.7	2.8	1350	350.8	14.3	1306.5	-955.7		0.27	
44	PIHA15				35.1	6.6	1450	380.8	14.6	1333.1	-952.3		0.29	
45	PIHA16				44.1	5.6	600	845.3	11.4	916.2	-70.9		0.92	
46	PIHA17			France	43.5	4.4	190	743.9	14.0	1029.2	-285.3		0.72	Nicault (1999)
47	PIHA18				43.8	4.8	330	788.9	13.2	1004.1	-215.2		0.79	
48	PIHA19				43.4	6.3	300	742.4	13.2	973.0	-230.6		0.76	

49	PIHA20		43.2	5.8	200	713.2	14.6	1015.3	-302.1	0.70	
50	PIHA21		43.7	5.4	170	816.3	11.8	944.8	-128.5	0.86	
51	PIHA22		44.0	6.4	600	1073.5	9.4	816.0	257.5	1.32	
52	PIHA23		43.4	5.0	150	749.1	13.6	986.4	-237.3	0.76	
53	PIHA28		44.2	4.4	350	888	12.2	966.6	-78.6	0.92	
54	PIHA31		43.8	4.1	300	743.6	14.2	1023.0	-279.4	0.73	
55	PIHA32		38.6	-2.6	1000	350.6	14.7	1203.4	-852.8	0.29	
56	PIHA34		37.3	-2.5	1280	287.6	14.7	1311.1	-1023.5	0.22	
57	PIHA35		38.9	-0.9	800	333.7	13.8	1270.0	-936.3	0.26	
58	PIHA36		40.8	-0.7	850	453.3	13.7	1152.7	-699.4	0.39	
59	PIHA37		38.1	-0.8	10	276.2	18.0	998.5	-722.3	0.28	
60	PIHA38		38.5	-0.6	900	380.1	14.2	1194.3	-814.2	0.32	
61	PIHA39		37.5	-3.2	1150	380.6	12.8	1362.8	-982.2	0.28	
62	PIHA40	Spain	41.6	-0.3	500	335.9	14.4	1203.3	-867.4	0.28	Ribas (2006)
63	PIHA41		41.3	0.8	750	304	14.2	1197.8	-893.8	0.25	
64	PIHA42		41.8	-0.7	350	310.2	14.2	1194.5	-884.3	0.26	
65	PIHA43		38.7	1.3	80	362.4	18.2	1177.3	-814.9	0.31	
66	PIHA44		40.0	3.9	0	486.9	16.7	1071.7	-584.8	0.45	
67	PIHA45		39.4	3.0	250	430.5	17.9	1080.0	-649.5	0.4	
68	PIHA46		39.8	2.6	700	548.7	14.3	1228.7	-680	0.45	
69	PIHA47		39.1	1.5	175	327.6	17.1	1303.5	-975.9	0.25	
70	PIHA48	Italy	40.3	14.8	8	711.5	15.1	1000.7	-289.2	0.71	Battipaglia <i>et al.</i> (2014)
71	CEAT1		32.2	-5.5	1920	483	9.9	1255.1	-772.1	0.38	Touchan <i>et al.</i> (2011)
72	CEAT2		33.1	-4.5	2180	526.5	15.2	1266.8	-740.3	0.42	
73	CEAT3	<i>Cedrus atlantica</i> (Endl.) Manetti ex Carrière	32.3	-5.3	2100	463.7	12.2	1339.1	-875.4	0.35	Esper <i>et al.</i> (2007)
74	CEAT4		33.2	-5.3	1830	539.4	10.5	1213.3	-673.9	0.44	
75	CEAT5		32.6	-5.0	2200	492.2	13.5	1360.5	-868.3	0.36	

76	CEAT6		32.6	-4.7	2200	492.2	13.5	1360.5	-868.3	0.36	
77	CEAT7	Algeria	36.3	3.6	1520	576.1	14.7	1196.2	-620.1	0.50	Touchan <i>et al.</i> (2011)

776

## 777 **References**

- 778 Battipaglia G, Strumia S, Esposito A, Giuditta E, Sirignano C, Altieri S, Rutigliano FA (2014) The effects of prescribed burning on *Pinus*  
779 *halepensis* Mill. as revealed by dendrochronological and isotopic analyses. *Forest Ecology and Management*, **334**, 201-208.
- 780 Battipaglia G, Savi T, Ascoli D, Castagneri D, Esposito A, Mayr S, Nardini A. (2016). Effects of prescribed burning on ecophysiological,  
781 anatomical and stem hydraulic properties in *Pinus pinea* L. *Tree physiology* (in press)
- 782 Esper J, Frank DC, Büntgen U, Verstege A, Luterbacher J, Xoplaki E (2007) Long-term drought severity variations in Morocco. *Geophysical*  
783 *Research Letters* 34, doi: 10.1029/2007GL030844.
- 784 Garfi, G. (2000). Climatic signal in tree-rings of *Quercus pubescens* and *Celtis australis* in South-eastern Sicily. *Dendrochronologia* 18, 41-51.
- 785 Gea-Izquierdo G, Cherubini P, Cañellas I. 2011. Tree-rings reflect the impact of climate change on *Quercus ilex* L. along a temperature gradient  
786 in Spain over the last 100 years. *Forest Ecology and Management* 262, 1807-1816.
- 787 Gea-Izquierdo G, Fonti P, Cherubini P, Martín-Benito D, Chaar H, Cañellas I. 2012. Xylem hydraulic adjustment and growth response of  
788 *Quercus canariensis* Willd. to climatic variability. *Tree Physiology* 32, 401-413.

- 789 Gea-Izquierdo G, Cañellas I. 2014 Contrasting instability in growth trends of Mediterranean oaks at opposite distributional limits. *Ecosystems*  
790 17, 228-241.
- 791 Nicault A. (1999). Analyse de l'influence du climat sur les variations inter- et intra-annuelles de la croissance radiale du pin d'Alep (*Pinus*  
792 *halepensis* Mill.) en Provence calcaire. Thèse de l'Université d'Aix-Marseille III, 254 p.
- 793 Ribas M. (2006). Dendroecología de *Pinus halepensis* Mill. en Este de la Península Ibérica e Islas Baleares: Sensibilidad y grado de adaptación a  
794 las condiciones climáticas. PhD dissertation. University of Barcelona. <http://diposit.ub.edu/dspace/handle/2445/35321>
- 795 Safar W. Serre-Bachet, F. & Tessier L. (1992). Les plus vieux pins d'Alep vivants connus. *Dendrochronologia*, 10, pp. 41-52.
- 796 Safar W. (1994). Contribution à l'étude dendroécologique du pin d'Alep (*Pinus halepensis* Mill) dans une région semi-aride d'Algérie : l'Atlas  
797 Saharien (Ouled Naïl - Aurès - Hodna). Thèse de l'université d'Aix - Marseille III., 215 p.
- 798 Touchan, R., K.J. Anchukaitis, D.M. Meko, M. Sabir, S. Attalah, and A. Aloui. (2011) Spatiotemporal drought variability in northwestern Africa  
799 over the last nine centuries. *Climate Dynamics* 37, 237-252.

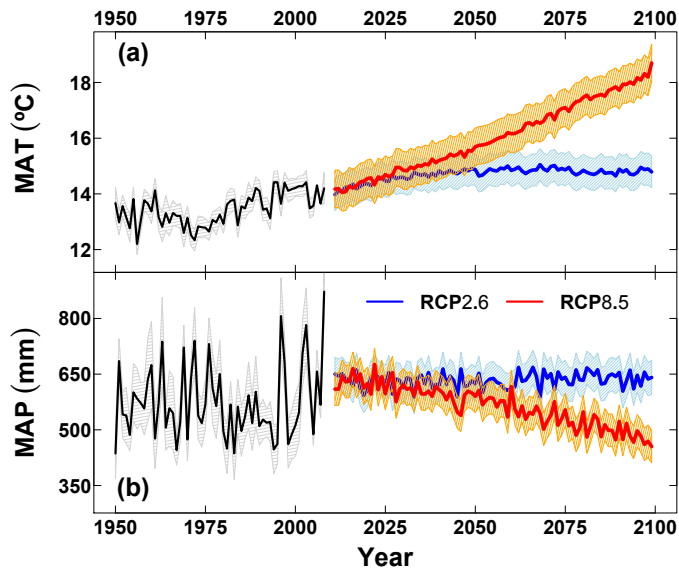
800 App. 2. Model code and Institute of CMIP5 climate simulations included in the study.

801 Crosses indicate when RCP scenarios were selected for a certain Model.

Model	Institute	RCP2.6	RCP8.5
bcc-csm1-1	Beijing Climate Center, China	X	X
bcc-csm1-1-m	Meteorological Administration	X	
BNU-ESM	College of Global Change and Earth System Science, Beijing Normal University	X	X
CanESM2	Canadian Centre for Climate modelling, Canada	X	X
CNRM-CM5	Centre National de Recherches Météorologiques / Centre Européen de Recherche et Formation Avancée en Calcul Scientifique, France	X	X
CSIRO-MK3-6-0	Commonwealth Scientific and Industrial Research Organization in collaboration with Queensland Climate Change Centre of Excellence	X	X
EC-EARTH	EC-EARTH consortium	X	X
FGOALS-g2	LASG, Institute of Atmospheric Physics, Chinese Academy of Sciences and CESS, Tsinghua University	X	X
GFDL-CM3	NOAA Geophysical Fluid Dynamics Laboratory, USA	X	X
GFDL-ESM2G		X	X
GFDLESM2M		X	
HadGEM2-ES	Met-Office – Hadley Center, contributed by Instituto Nacional de Pesquisas Espaciais, Spain	X	X
IPSL-CM5A-LR	Institut Pierre-Simon Laplace, France	X	X
IPSL-CM5A-MR		X	X
MIROC-ESM	Japan Agency for Marine-Earth Science and Technology, Atmosphere and Ocean Research Institute (The University of Tokyo), and National MIROC-ESM-CHEM Institute for Environmental Studies	X	X
MIROC-ESM-CHEM		X	X
MIROC5	Atmosphere and Ocean Research Institute (The University of Tokyo), National Institute for MIROC MIROC4h Environmental Studies, and Japan Agency for MIROC5 Marine-Earth Science and Technology	X	X
MPI-ESM-LR	Max-Planck Inst. Für Meteorologie, Germany	X	X
MRI-CGCM3	Meteorological Research Institute, Japan	X	X
NorESM1-M	Norwegian Climate Centre		X
# Simulations		19	18

802

803 App. 3. Mean daily annual temperature (MAT) and mean annual precipitation (MAP) of  
804 the climate model simulations at each of the 77 sites for the RCP 2.6 (a) and RCP 8.5  
805 (b). Shaded areas are confidence intervals at  $\alpha=0.05$  for annual means.



806  
807



808 App. 4. Mean values and standard deviations of 7 fitted parameters to the 77 sites.

809

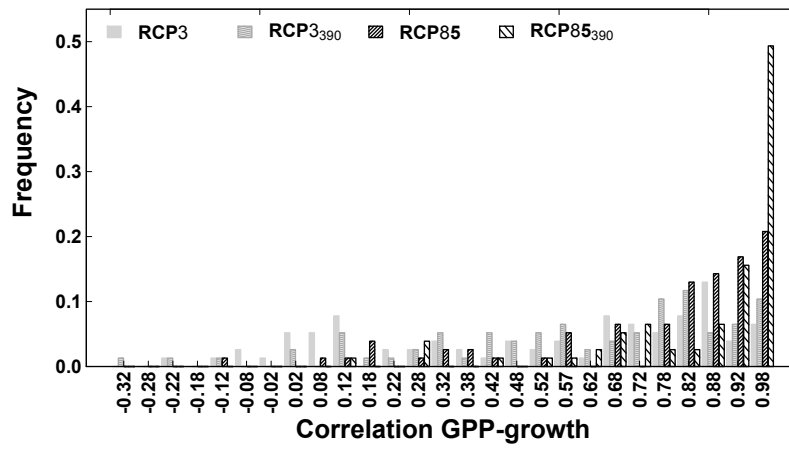
<b>Parameter</b>	<b>Mean</b>	<b>Standard deviation</b>	<b>Minimum</b>	<b>Maximum</b>
soil <sub>ip</sub>	139.0	96.7	3.1	392.9
p <sub>3moist</sub>	-0.255	0.211	-0.700	-0.001
p <sub>3temp</sub>	26.7	77.8	-153.7	181.2
st <sub>3moist</sub>	-0.226	0.197	-0.699	-0.006
st <sub>3temp</sub>	4.8	88.2	-167.4	222.0
st <sub>4sd_moist</sub>	-0.321	0.230	-0.900	-0.009
st <sub>4temp</sub>	334.9	212.2	49.7	1080.2

810

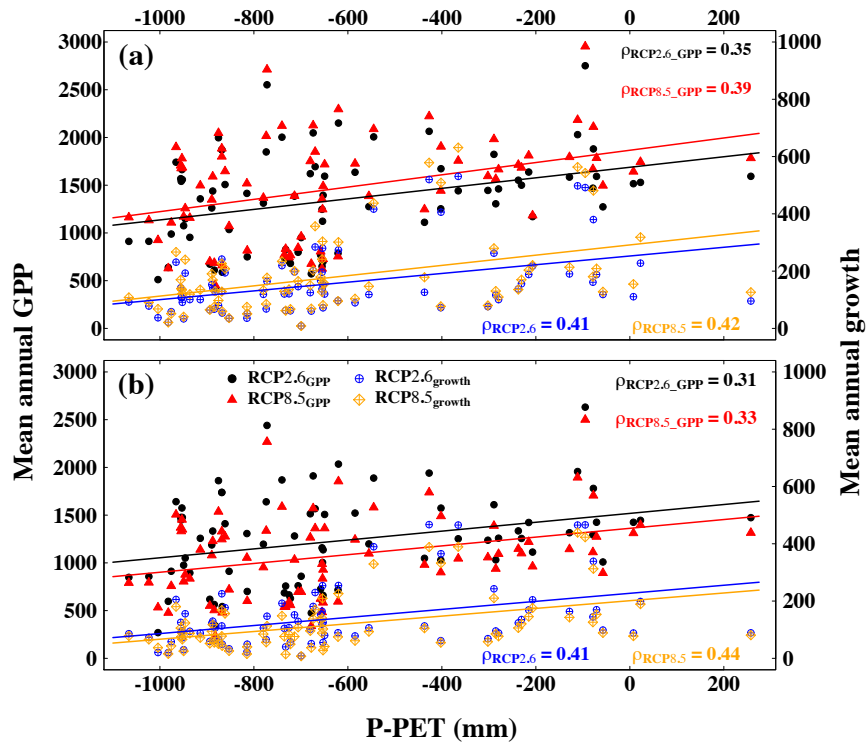
811 App. 5. Distribution of the relationship between GPP and carbon allocation to the stem  
812 (i.e. radial growth) as estimated by a linear correlation.

813

814



815 App.6. Mean projected GPP and growth (2010-2100) and site 'Annual precipitation  
 816 minus Penman-Monteith potential evapotranspiration' (P-PET, in mm) for RCP2.6 and  
 817 RCP8.5 scenarios: (a) fertilization scenario (i.e. predicted  $c_a$ ); (b) non-fertilization  
 818 scenario (i.e.  $c_a=390$  ppm). All relationships ( $\rho$ =correlation) are significant at  $p<0.01$ .



819  
 820

821 App. 7. In this figure we show two graphs to support the robustness of our modeling  
 822 approach and the future projections implemented at the regional scale. (A) The  
 823 coefficient of determination ( $R^2$ ) is shown as a function of the projected growth trends  
 824 to demonstrate that there is no relationship between the estimated trends and the  
 825 calibration  $R^2$ . In addition, regardless of  $R^2$ , all projections under RCP8.5 are either  
 826 positive (fertilization, in blue) or negative (non-fertilization, in red), whereas variability  
 827 in RCP2.6 slopes estimated is independent of the goodness of fit (as estimated by  $R^2$ ,  
 828 see (A)). In (B), to illustrate the model capacity to fit the interannual (decadal) growth  
 829 trends, we show the slopes estimated on past growth observations in function of the  $r_{low}$   
 830 statistic (correlations between filtered series).  $r_{low}$  was greater than 0.7 in 73% of the  
 831 models fitted (B). Most importantly,  $r_{low}$  was greater than 0.6 in all cases when there was  
 832 a significant past trend (in (B) we highlight those sites where past  $|\text{slope}| \geq 0.35$   
 833  $\text{cm}^2 \cdot \text{year}^{-2}$ ), which means that the model was able to mimic the long-term growth-trends  
 834 when trees exhibited some positive or negative trend in the past.

