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Key Points:

- The model reproduces gradients of physical and biogeochemical properties in the SCB, including seasonal, interannual, and ENSO variability
- Resolving submesoscale currents increases the vertical flux of nutrients and allows realistic reproduction of wastewater plume circulation
- Wastewater and riverine inputs are consistent with intensified nutrients, productivity, export, and remineralization nearshore

Supporting Information:

Supporting Information may be found in the online version of this article.

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Abstract The Southern California Bight (SCB), an eastern boundary upwelling system, is impacted by global warming, acidification, and oxygen loss and receives anthropogenic nutrients from a coastal population of 20 million people. We describe the configuration, forcing, and validation of a realistic, submesoscale-resolving ocean model as a tool to investigate coastal eutrophication. This modeling system represents an important achievement because it strikes a balance of capturing the forcing by U.S. Pacific Coast-wide phenomena, while representing the bathymetric features and submesoscale circulation that affect the transport of nutrients from natural and human sources. Moreover, the model allows simulations at time scales that approach the interannual frequencies of ocean variability. The model simulation is evaluated against a broad suite of observational data throughout the SCB, showing realistic depiction of the mean state and its variability with satellite and in situ measurements of state variables and biogeochemical rates. The simulation reproduces the main structure of the seasonal upwelling front, the mean current patterns, the dispersion of wastewater plumes, as well as their seasonal variability. Furthermore, it reproduces the mean distributions of key biogeochemical and ecosystem properties and their variability. Biogeochemical rates reproduced by the model, such as primary production and nitrification, are also consistent with measured rates. This validation exercise demonstrates the utility of using fine-scale resolution modeling and local observations to identify, investigate, and communicate uncertainty to stakeholders to support management decisions on local anthropogenic nutrient discharges to coastal zones.

Plain Language Summary We applied and validated an ocean numerical model to investigate the effects of land-based and atmospheric nutrient loading on coastal eutrophication and its effects on carbon, nitrogen, and oxygen cycles of the Southern California Bight, an upwelling-dominated marine embayment on the U.S. West Coast. The model is capable of high-resolution, multiyear hindcast simulations, which enable investigations to disentangle natural variability, climate change, and local human pressures that accelerate land-based and atmospheric nutrient loads. The model performance assessment illustrates that it faithfully reproduces monitored ocean properties related to algal blooms, oxygen, and water acidity, among others, that can be traced to land-based and atmospheric inputs of nutrients and carbon from human activities. The model performance assessment helps to constrain uncertainties in predictions to support ongoing conversations on approaches to reduce the effects of climate change, including considerations of management of local nutrient and carbon inputs.

1. Introduction

Human-driven eutrophication has resulted in profound impacts to coastal ecosystems around the world. These impacts are arguably the best studied in estuaries and enclosed bays (e.g., Chesapeake Bay; Boesch et al., 2001; Cerco & Cole, 1993) and semienclosed seas such as the Baltic Sea (Cederwall & Elmgren, 1990; Savchuk & Wulff, 2007), the Mediterranean Sea (Arhonditsis et al., 2000), and the Gulf of Mexico (Justić et al., 2005; Laurent et al., 2018). To date, few investigations of coastal eutrophication have occurred in eastern boundary upwelling systems (EBUS). While strong upwelling and vigorous surface currents would generally limit the extent

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to which coastal eutrophication could occur (Fennel & Testa, 2019), such investigations have also been limited by coupled physical–biogeochemical numerical modeling approaches that can adequately resolve fine-resolution bathymetry and the complexities of submesoscale circulation (Dauhajre et al., 2019; McWilliams, 2016), while simulating a sufficient duration (several years) to distinguish oceanic versus terrestrial forcing. These submesoscale circulation features, including fine-scale eddies and filaments <5 km in horizontal resolution, strongly control the magnitude and variability of nearshore upwelling and associated nutrient transport. Thus, high-resolution, submesoscale-resolving numerical models are a necessary prerequisite for mechanistic modeling studies and source attribution of oceanic versus terrestrial drivers of coastal eutrophication in EBUS. Inadequate modeling system and lack of numerical model validation have been identified as significant barriers to effective, evidence-based solutions to coastal eutrophication (Boesch, 2019).

All the necessary ingredients are present to motivate a numerical modeling investigation of the role of coastal eutrophication in driving ocean acidification and oxygen loss in the Southern California Bight (SCB), a large marine open embayment found in the California Current System (CCS) on the U.S. Pacific Coast. First, the SCB is a biologically productive region, and thus of high economic and ecological importance. Seasonal upwelling of nutrient-rich deep water maintains high rates of biological productivity over broad scales. At the same time, upwelling draws water masses that are naturally low in dissolved oxygen, pH, and carbonate saturation state (Ω_{Ar}) onto the shelf and into the photic zone (Sutton et al., 2017). Second, the SCB has one of the most spatially comprehensive and longest-running coastal observational systems in the world. Several physical and biogeochemical variables are sampled regularly and extensively, creating an ideal setting for model-data comparisons. Third, the SCB is home to one of the most densely populated coastal regions in North America, where the discharges of primary or secondary treated wastewater from a population of 20 million people are released to the coastal zone via ocean outfalls, along with the urban and agricultural runoff from 75 rivers. These nutrient sources rival natural upwelling in magnitude (Howard et al., 2014), roughly doubling available nitrogen to nearshore coastal waters. Intensifying ocean acidification, oxygen loss, and harmful algal blooms have motivated California policy makers to consider reducing anthropogenic nutrients as a climate change mitigation strategy (Ocean Protection Council, 2018), but wastewater treatment plant upgrades and methods to increase control or reduce nonpoint sources would cost billions. A numerical modeling approach is needed to disentangle the effects of natural upwelling and climate change from anthropogenic nutrient loading from land-based and atmospheric sources.

To support such investigations, the Regional Oceanic Model System (ROMS, Shchepetkin & McWilliams, 2005) coupled to the Biogeochemical Elemental Cycling (BEC) model (Moore et al., 2004) has been recently adapted for the CCS (Deutsch et al., 2021; Renault et al., 2021). A downscaled model domain was established, scaling from a 4 km horizontal resolution configuration spanning the entire CCS, to a 1-km resolution grid covering the much of the California coast (latitude <40.25°N), to a 0.3-km grid in the SCB, where investigations of local anthropogenic inputs were focused. Modeling experiments investigating submesoscale transport (captured at model resolutions ≤1 km) have demonstrated an up to tenfold increase in the magnitude of instantaneous vertical N fluxes (Kessouri, Bianchi, et al., 2020) relative to mesoscale transport represented by a 4 km model (Section 2.2). Furthermore, a finer horizontal resolution of bathymetry improves the representation of coastal currents, submesoscale circulation, and coast–offshore connectivity (Dauhajre et al., 2019). For this reason, investigations of coastal eutrophication are simulated here at 0.3 km horizontal resolution. Simulations conducted with the 4 km ROMS-BEC model domain have been validated for regional-scale atmospheric forcing, physics, and biogeochemistry, including O_2, carbonate saturation state, primary productivity, and hydrographic parameters, demonstrating that the model captures broad patterns of critical properties in the CCS (Deutsch et al., 2021; Renault et al., 2021). However, additional focused validation of nearshore, anthropogenically enhanced gradients in nutrients, primary production, oxygen, and pH in model simulations conducted at 0.3-km resolution are needed to gauge model utility to investigate the impacts of coastal eutrophication on ocean acidification and oxygen loss.

We employed this downscaled, submesoscale-resolving physical–biogeochemical model to investigate the effects of land-based and atmospheric nutrient inputs in driving coastal eutrophication and ocean acidification and oxygen loss (Kessouri et al., 2021). The aim of this manuscript is to (a) document the SCB ROMS-BEC model configuration, including the effects of land-based and atmospheric inputs of nutrients and organic carbon, intended to support investigations of coastal eutrophication and (b) present a validation of SCB ROMS-BEC simulations against available observations, focusing on anthropogenically enhanced gradients in nutrients, primary production, oxygen, and pH.

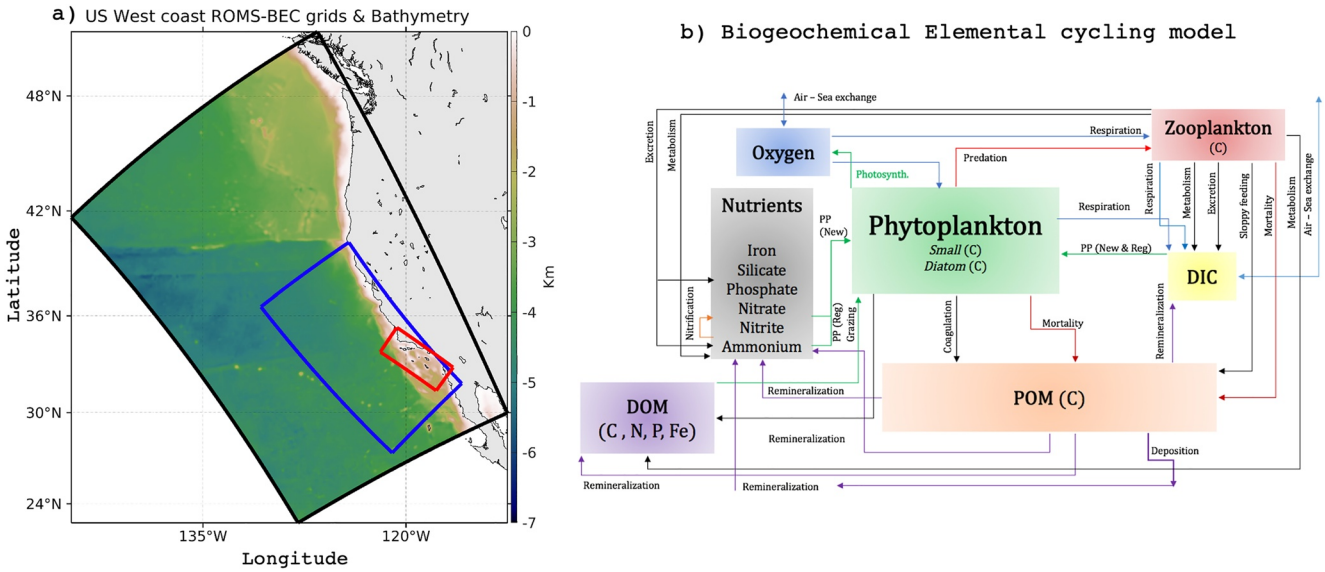

Figure 1. (a) Regional Oceanic Modeling System-Biogeochemical Elemental Cycling (ROMS-BEC) model configurations. dx = 4 km is the black box, dx = 1 km is the blue box, and dx = 0.3 km is the red box. Background color shading show the topography from dx = 4 km. (b) Schematic of the Biogeochemical Elemental Cycling model. The schematic shows state variables (boxes) and biogeochemical rates and feedback (arrows).

2. SCB Coupled Physical and Biogeochemical Model Description, Configuration, and Forcing

2.1. Model Description

2.1.1. Ocean Hydrodynamics

Ocean hydrodynamics is modeled with ROMS, a free-surface, terrain-following coordinate model with 3-D curvilinear coordinates that solves the primitive equations with split-explicit time steps. It contains state-of-art numerical algorithms that provide an accurate and stable representation of physical processes down to scales of tens of meters and allows for offline downscaling of high-resolution subdomains within larger domains. The offline downscaling is based on the Orlanski scheme for the baroclinic mode (Marchesiello et al., 2001) and a modified Flather scheme for the barotropic mode (Mason et al., 2010). Vertical mixing in the boundary layers is represented by a K-profile parameterization (W. G. Large et al., 1994). The U.S. West Coast hindcast model has been successfully run over two decades (between 1997 and 2017) at 1 and 4 km horizontal resolution using high-resolution spatial and temporal atmospheric forcing that represent the effects of wind dropoff, the current feedback on the surface stress, and high-frequency (HF) wind fluctuations (Renault, Hall, et al., 2016; Renault, Molemaker, McWilliams, et al., 2016). For this study, we further downscale to 0.3-km resolution to capture submesoscale processes.

2.1.2. Biogeochemistry

Ocean biogeochemical modeling approaches can have a broad range of complexities, ranging from few functional groups (e.g., NPZD models, Fasham, 1993), to multiple biogeochemical cycles (e.g., C, N, and O) and plankton functional groups. To provide a representation of biogeochemical cycles, ROMS is dynamically coupled to the BEC model (Deutsch et al., 2021; Gruber, 2004; Gruber et al., 2011; Moore et al., 2004). A schematic of BEC is shown in Figure 1b. BEC is a multielement (C, N, P, O, Fe, and Si) and multiplankton model that includes three explicit phytoplankton functional groups (picoplankton, silicifying diatoms, and N-fixing diazotrophs), one zooplankton group, and dissolved and sinking organic detritus. The impacts of calcifying phytoplankton (coccolithophores) on the carbon system are represented implicitly. Remineralization of sinking organic material follows the multiphase mineral ballast parameterization of Armstrong et al. (2001), and sedimentary processes have also been expanded. Particulate organic matter reaching the sediment is accumulated and slowly remineralized with a time scale of 330 days, to provide a buffer between particle deposition and nutrient release. Nitrogen loss to the sediment is parameterized according to the empirical diagenetic model for sediment denitrification

of Middelburg et al. (1996). Water column denitrification is only active when oxygen concentrations fall below 5 mmol m^{-3}. Sedimentary release of Fe is based on the benthic chamber measurements of Severmann et al. (2010) for the California–Oregon coast and increases as bottom water oxygen concentrations decrease. Atmospheric dust deposition follows the parameterization by Mahowald et al. (2006) and provides an additional source of iron at the surface, although of minor importance compared to sedimentary iron release in the region (Deutsch et al., 2021). The ecosystem is linked to a carbon system module that tracks dissolved inorganic carbon and alkalinity, and an air–sea gas exchange module that allows realistic representation of dissolved gases (e.g., O_2, CO_2, and nitrous oxide), based on the formulation of Wanninkhof (1992).

2.1.3. SCB Model Configuration

The SCB model domain extends along a 450-km stretch of the coast, from Tijuana to Pismo Beach, and about 200 km offshore. This grid, shown in Figure 1a, is composed of 1,400 × 600 grid points, with a nominal resolution of $dx = 0.3$ km. The grid has 60 σ-coordinate vertical levels using the stretching function described in Shchepetkin and McWilliams (2009), with the following stretching parameters: $\theta_s = 6$, $\theta_b = 3$, and $h_c = 250$ m. The model is run with a time step of 30 s, and outputs are saved as 1-day averages.

The oceanic forcing of the 0.3-km domain originates from multilevel offline downscaling. A 4-km simulation is initialized and forced at the open boundaries by a preexisting North-east Pacific-wide ROMS solution at 12-km resolution (Renault et al., 2021), initialized and forced on the boundaries by the global model Mercator Glorys2V3 (http://www.myocean.eu) for the physics, and with reconstruction of biogeochemical fields using world ocean database. We used climatological fields for organic material and relationships with density for nutrients. Full description of the boundary conditions and initialization of the parent configuration at 12-km can be found in Deutsch et al. (2021). The 4-km configuration is run for the period 1995–2017, after a spin-up of 2 years. A 1-km simulation is initialized and forced from the 4-km model, including initial conditions and open boundary conditions, starting in October 1996 and ending in December 2007. The 0.3-km simulation is initialized and forced at its boundaries by the 1-km simulation starting from January 1997 and ending in December 2000. The bathymetry used in this configuration comes from the Southern California Coastal Oceanic Observation System (SCCOOS) 3 Arc-Second Coastal Relief Model Development (90-m horizontal resolution).

The oceanic model is forced by hourly outputs from the atmospheric uncoupled Weather Research and Forecast model (WRF06; Skamarock & Klemp, 2008). Using bulk formulae (W. B. Large, 2006), WRF06 provides heat, surface evaporation, momentum, and atmospheric data and is run at 6-km resolution over a domain similar to the 4-km (Figure 1 and used for Renault, Hall, et al. [2016]) and includes a wind–current coupling parameterization necessary to attain more realistic simulations of the oceanic eddy kinetic energy and circulation (Renault et al., 2020; Renault, Molemaker, McWilliams, et al., 2016).

Model simulations were conducted from 1997 to 2000, a period chosen to capture the effects of all three phases of the El Niño–Southern Oscillation (ENSO); it also captures the beginning of the "modern" state of point source management in the SCB, where several large Publicly Owned Treatment Plants (POTW) were in transition from primary to secondary treatment. (We will refer to submarine point sources outfalls from the treatment plants as "POTW" hereafter.)

2.2. Importance of Submesoscale Circulation

Downscaling to $dx = 0.3$ km allows the model to represent ocean circulation that includes baroclinic and barotropic eddies and turbulence generated at the submesoscale (Capet, Campos, et al., 2008). Submesoscale dynamics energizes frontogenesis by mesoscale straining and mixed layer instabilities (Capet, Campos, et al., 2008; Capet, Klein, et al., 2008; Capet, McWilliams, et al., 2008). Oceanic fronts are a driver of significant nutrient supply to the upper ocean. They have also been recognized as areas of enhanced biomass in many regions of the global ocean (Woodson & Litvin, 2015), as well as important locations for fisheries (e.g., Galarza et al., 2009). Resolving submesoscale eddies dramatically increases the variability of vertical fluxes of biogeochemical tracers and other material properties, eventually allowing a more accurate representation of chemical and biological constituents. Figure 2 (upper panel) shows the temporal variability and horizontal distribution of vertical eddy fluxes of nitrate at 40-m from three different resolutions with the ROMS-BEC model (see Section 2.1.3). Submesoscale dynamics increase instantaneous fluxes by more than 1 order of magnitude, with more frequent and vigorous fine-scale structures (Figure 2, bottom panels) when increasing the resolution from 4 to 1 km, and similarly

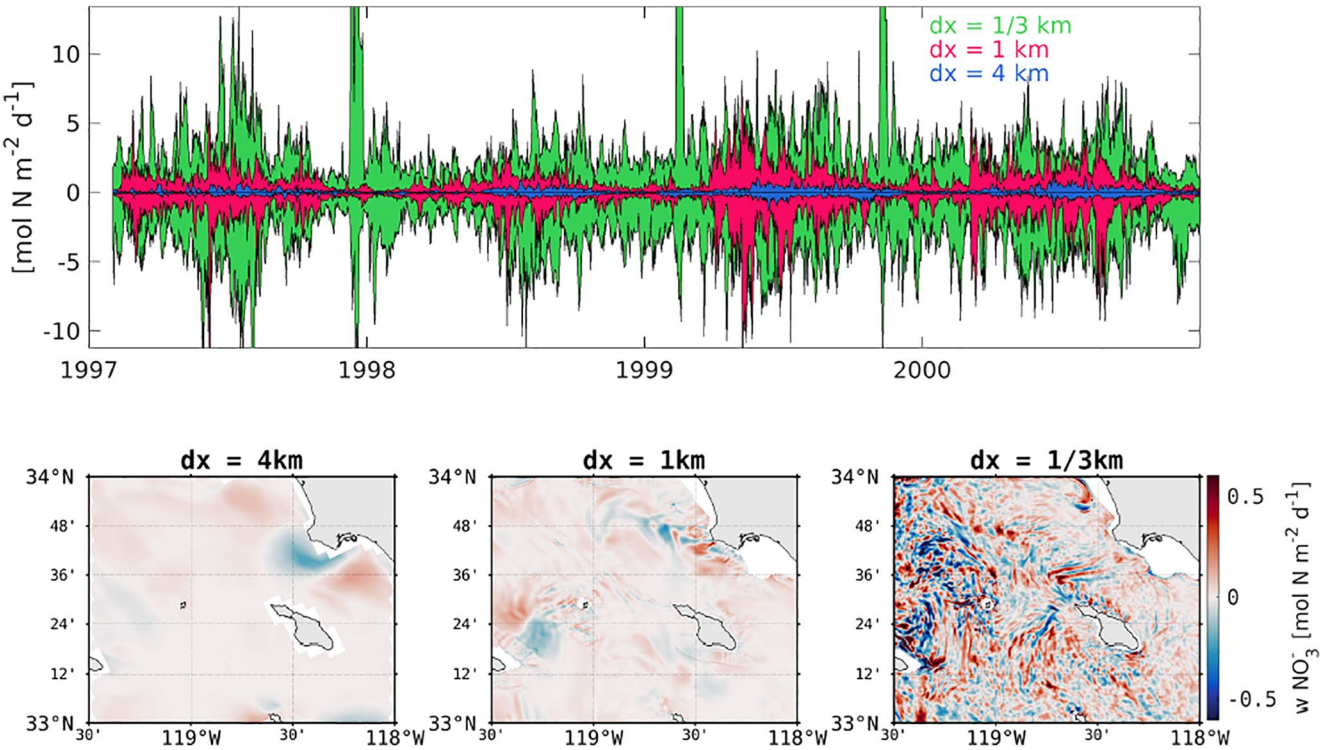

Figure 2. (Upper panel) Time series (1997–2001) of the vertical eddy flux of nitrate at 40-m depth calculated as follows: $\overline{wN} = \overline{w}\,\overline{N} + \overline{w'N'}$, where the overbar represents a monthly average, and the prime the deviation from this average, for region covering the entire Southern California Bight (31.4°–35.3°N and 116.5°–121.8°W). The minimum and maximum values (i.e., the envelope) of the flux are shown in blue for the 4-km solution, in red for the 1-km solution, and in green for the 1/3 km. (Lower panel) Snapshot of the vertical flux of nitrate in spring at 40-m off the coast of Palos Verdes that shows higher magnitudes and enhanced variability as resolution increases.

another order of magnitude when increasing resolution from 1 to 0.3 km. Intensification of vertical flux of nitrate at the euphotic depth has previously been shown in idealized models (Lévy et al., 2012; Mahadevan, 2016) and in realistic simulations in the central California upwelling system (Kessouri, Bianchi, et al., 2020) but has never been modeled in the SCB at this resolution. The impact of the submesoscale on nutrients fluxes is more apparent during winter, when the mixed layer is deeper, wind forcing more intense, and submesoscale circulation more energetic. During this period, nutrients are transported from the nutricline by transient fronts that can last only few days and are only properly resolved in the 0.3-km configuration. Submesoscale eddies have been associated with increased productivity in the oligotrophic ocean (Mahadevan, 2016) and decreased productivity in the upwelling region (Kessouri, Bianchi, et al., 2020). Our submesoscale-resolving simulation at dx = 0.3 km is an opportunity to quantify the balances of nitrogen, dissolved oxygen, carbon, and productivity using a more realistic representation of the physical circulation, as well as a representation of urban anthropogenic inputs to the ocean.

Our simulations show that the increased number of fronts and submesoscale instabilities promote intense variability of nitrate transport, as shown in Figure 2, as well as increased heterogeneity at the subsurface chlorophyll *a* maximum. However, surface phytoplankton biomass is only intensified if the time scale of the enrichment is sufficiently long and maintained in these small-scale features. We argue that modeling at this scale allows for a more accurate simulation of biogeochemical tracers and rates, as described in subsequent sections. However, we also note that comparing the model and observations to highlight the realism of submesoscale processes is challenging, mostly because of the lack of observations of biogeochemical variables at high enough spatial and temporal resolution. Furthermore, changes in the distribution of biogeochemical tracers as the model resolution increases are relatively subtle (Kessouri, Bianchi, et al., 2020) and in general within the range of variability of observations and simulations. By construction, the submesoscale-resolving model better represents scales relevant to coastal circulation and anthropogenic nutrient emission and dispersal, and the underlying dynamics (Capet, Campos, et al., 2008). Showing that this configuration indeed compares realistically with observations, in an average and

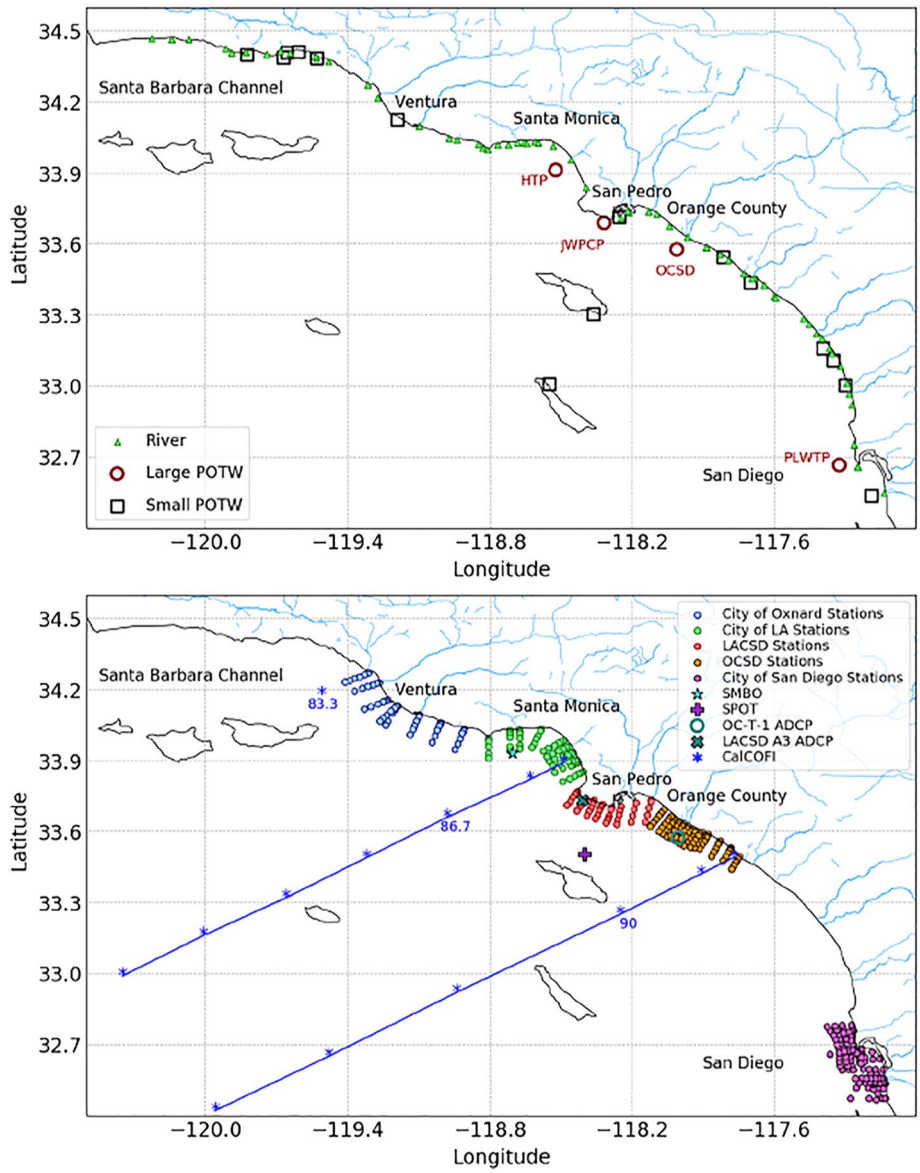

Figure 3. (Upper panel) Location of rivers and Publicly Owned Treatment Plants (POTW) outfalls along the Publicly Owned Treatment Plants (SCB). (Lower panel) Location of monitoring stations used for the validation, including POTW quarterly monitoring surveys, California Cooperative Oceanic Fisheries Investigations (CalCOFI) seasonal observations, showing the line numbers, Santa Monica Bay Observatory (SMBO), and San Pedro Oceanographic Time Series (SPOT), mooring.

statistical sense, strongly supports the validity of the model for coastal biogeochemical applications, even though aspects of the simulations such as submesoscale processes remain challenging to directly assess.

2.3. Terrestrial and Atmospheric Forcing of Freshwater, Nutrients, and Carbon

Model simulations were forced with a monthly time series of spatially explicit inputs (Figure 3, upper), including freshwater flow, nitrogen, phosphorus, silica, iron, and organic carbon representing natural and anthropogenic sources (Sutula et al., 2021b). These data include POTW ocean outfalls and riverine discharges (1997–2017) and spatially explicit modeled estimates of atmospheric deposition. POTW effluent data were compiled from permit monitoring databases and communication with sanitary agencies. Monthly time series of surface water runoff from 75 rivers are derived from model simulations and monitoring data (Sutula et al., 2021b). Direct atmospheric deposition is derived from the Community Multiscale Air Quality (CMAQ) model (Byun et al., 2006) and follows

the implementation of Deutsch et al. (2021). In this paper, we discuss in detail the formulation of the river and wastewater outfall inputs.

2.4. Configuration of River and Wastewater Outfall Forcing in the Model

Ocean outfalls and coastal rivers are modeled as mass sources into the ocean (Figure 3, upper). To accomplish this, we add explicit volume fluxes to the otherwise divergence-free flow in the ocean. The inclusion of these fluxes makes it possible to account for associated sources of tracers, while satisfying conservation laws. Specifically, our approach allows for the proper influx of fresh water in the ocean, without resorting to a "virtual salt" flux, which is a common approach in larger scale ocean models (Kang et al., 2017). Since we explicitly include known volume fluxes for both rivers and outfall pipes, specification of tracer concentration is sufficient to correctly model the source terms. The tracer evolution equations that are used in ROMS are implemented by using control volumes (Shchepetkin & McWilliams, 2005) where for each tracer concentration $C = C(x, y, z, t)$,

$$\frac{\partial \iiint C\, dV}{\partial \mathrm{t}} = \iint u_n C\, dA, \tag{1}$$

where $V = V(x, y, z, t)$ is the volume of the entire domain, u_n is the normal velocity into the volume, and $A = A(x, y)$ is the total area of grid cells source is being input. Additionally, we enforce mass conservation which implies

$$\frac{\partial \mathrm{V}}{\partial \mathrm{t}} = \iint u_n\, dA. \tag{2}$$

In absence of rivers and outfalls, the flow is volume conservative, and the integral on the right-hand side of Equation 2 is zero. Using Equations 1 and 2, it is easy to see that the mean concentration of a tracer can be lowered if the average concentration of the flux entering the control volume is less than the mean concentration in that volume. In this manner, fresh water rivers will lower the salinity of the water in which they enter. All 75 rivers and 23 POTW pipes that are considered in this study are implemented in this manner.

Each individual source is based on the following equation:

$$S(x, y, z, t) = \frac{W(x, y, z)\, Q_s(t)\, C_s(t)}{V_s}, \tag{3}$$

with $S(x, y, z, t)$ is the volume source of contaminant (mmol m^{-3} s^{-1}), $W(x, y, z)$ the nondimensional shape function (with values between 0 and 1), $Q_s(t)$ the water volume flux from the source (m^3 s^{-1}), $C_s(t)$ the concentration of the tracer C in the source water (mmol m^{-3}), and V_s is the effective volume of the source (m^3).

For each source, $Q_s(t)$ and $C_s(t)$ are prescribed as time series. The shape function $W(x, y, z)$ distributes the tracer spatially and in the water column, representing nonresolved mixing and dilution effects. Its values represent the relative intensity of the in situ tracer injection, with values between 0 and 1. Tracer concentration C is distributed in the water column as $C(x, y, z, t) = W(x, y, z)C_s(t)$ The effective 3-D volume of the source is calculated from the shape function $W(x, y, z)$ as

$$V_s = \iiint W(x, y, z)\, dV, \tag{4}$$

where the integral is over the model domain. For convenience, we assume that $W(x, y, z)$ can be separated into a horizontal shape function $A(x, y)$, multiplied by a vertical shape function $H(z)$ (both nondimensional and with values between 0 and 1), such that

$$V_s = \iint A(x, y)\, dx\, dy \int H(z) dz = A_s\, H_s. \tag{5}$$

Here, A_s represents the effective source surface area (m^2), and H_s the effective source thickness (m). The functions $A(x, y)$ and $H(z)$ are defined differently for POTW and rivers. They are assumed to be fixed in time; a time-dependent generalization (e.g., to mimic variations in the depth of the POTW buoyant plume) is straightforward. For POTW inputs, at each main diffuser, the horizontal distribution $A(x, y)$ of the source is shown in Figure S1

of Supporting Information S1. This method of weighting the plume in different cells allows the effluent to be properly diluted vertically and horizontally at this resolution and prevents the model from developing numerical instabilities.

Each large treatment plant has specialized outfall configurations that are taken into account for representation in the model (Figure S1 in Supporting Information S1). The flow is divided in two at Hyperion Treatment Plant (HTP) located 6 km off Marina Del Rey (Santa Monica Bay; Figure S1a in Supporting Information S1) and Point Loma Wastewater Treatment Plant (PLWTP) in San Diego coast (Figure S1d in Supporting Information S1) to account for their Y-shaped diffuser, partitioning 50% of the flow to each diffuser. Orange County Sanitation District (OCSD) located 6 km off Huntington Beach (Figure S1c in Supporting Information S1) has one flow through its L-shaped diffuser. Joint Water Pollution Control Plant (JWPCP) in Palos Verdes shelf (Los Angeles; Figure S1b in Supporting Information S1) has three diffusers, the Y-shape northern typically discharges 17.5% of the flow for each leg of the Y-diffuser, and the southern L-shape diffuser discharges 65% of the flow. The vertical profile of the POTW sources is defined by a Gaussian function centered at a height z above the bottom (h_b), to mimic a buoyant plume, so that $H(z)$ is given by

$$H(z) = e^{-z^2/d_s^2}, \tag{6}$$

where $z = -h_b + h_s$, with h_b is the bottom depth (m), h_s the depth of the buoyant plume above the bottom (m), and d_s is the vertical scale of the POTW plume (m). We further assume $h_s = 20$ m and $d_s = 10$ m, as in Uchiyama et al. (2014).

We distribute the SCB rivers on one horizontal grid point (0.3 km wide), where we assume $A(x, y) = 1$, and similarly distribute the source vertically, with the Gaussian function centered at the surface. h_s here is simply the water column depth to put the maximum input at the surface. Because in ROMS the thickness of vertical grid cells varies in time, to ensure tracer conservation the calculation of the input source volume V_s must be done at each time step, even in the case of a time-independent source shape function $W(x, y, z)$. Effectively, only $H_s = H(z)$ needs to be recalculated at each time step.

3. Model Performance Assessment Approach

The conceptual approach for model performance assessment comprises three components, addressing different aspects of skill assessment: (a) statistical comparison of model output to observational data for state variables by region and season, (b) comparison of model output to observational data for biogeochemical rates, and (c) evaluation of model behavior compared to expected biogeochemical dynamics for coastal zones. Comparison of model output to observational data by region and season is designed to document model skill at reproducing the statistics (e.g., mean values and variability) of ocean physical and biogeochemical parameters at the spatiotemporal scales more relevant for evaluating human impacts on the coastal environment. Comparison of model output to observational data for biogeochemical rates assures that model is capturing the appropriate transformations in nutrients and carbon that structure the ecosystem response to eutrophication. Finally, the evaluation of model behavior compared to the expected physical and biogeochemical dynamics for coastal zones is a more qualitative evaluation of model performance to document that the model broadly reproduces oceanographic phenomena in a way that reflects our understanding of nearshore ocean environments.

3.1. Description of Observational Data Sets

3.1.1. Ship-Based Ocean Monitoring

The SCB is home to a suite of long-running monitoring programs that make it one of the best observed coastal ecosystems in the world (3, lower). Among them, the California Cooperative Oceanic Fisheries Investigations (CalCOFI) program (McClatchie, 2016), initiated in the 1950s, samples the SCB quarterly each year, collecting hydrographic and biogeochemical measurements in coordination with the SCCOOS. These observations are augmented nearshore by quarterly surveys of nearshore water column and benthic parameters conducted collaboratively since 1990 by POTW agencies as a part of their regulatory monitoring requirements (Booth et al., 2014; Howard et al., 2014; McLaughlin et al., 2018; Nezlin et al., 2018). These programs provide good temporal and geographical coverage of both the offshore (CalCOFI) and nearshore (POTW) areas, coinciding with the model

period, and include publicly available water quality data for targeted sites measured quarterly. We validated model output against observed temperature, dissolved oxygen, nitrate, ammonium, chlorophyll, carbon system parameters (pH and aragonite saturation state), primary production, and nitrification.

In situ measurements have inherent uncertainty, due to a combination of measurement sensitivity and sampling frequency and intensity, making them an imperfect "truth" with which to compare to model output. However, this uncertainty is not the same for all parameters. Both temperature and dissolved oxygen are collected using high-resolution probes, though the two programs used in this study incorporate slightly different calibration protocols for dissolved oxygen, lending greater confidence to data–model comparisons for these data sets. Chlorophyll is measured on discrete bottle samples in the CalCOFI program, a high-quality measurement, but inferred from in situ fluorescence measurements in the POTW monitoring program, adding uncertainty to these measurements. Nitrate and ammonium concentrations are measured on discrete bottle samples for both programs, but the detection limits are more sensitive in the CalCOFI program. Furthermore, nutrients are not measured with the same sampling density in POTW monitoring programs as sensor data. Similarly, primary production is measured at a subset of locations in the CalCOFI program and as a short-term special study in Southern California Bight Regional Marine Monitoring Program (Bight Program). Details on measurements and sample collection protocols for the CalCOFI program can be found on their website (https://calcofi.org; McClatchie, 2016) and for the POTW monitoring programs in Howard et al. (2014). We also use selected nutrient observations from the Santa Monica Bay Observatory (SMBO) mooring located in the Santa Monica Bay (Leinweber et al., 2009). Figure 3 shows a map of all monitoring stations used in this study. The repository of data can be found in Kessouri, McLaughlin, et al. (2020).

3.1.2. High-Frequency Radar and Acoustic Doppler Current Profilers

HF radar data from the database of the University of California, San Diego (https://hfrnet-tds.ucsd.edu/thredds/catalog.html) provide surface currents along the west coast of the United States, including the SCB. Seasonally averaged data from 2012 to 2020 were used to analyze trends of surface currents in the Bight compared to the model. Acoustic Doppler Current Profilers (ADCP) provide current data in the water column. ADCP measurement data from OCSD for the period June 1999 to June 2000 and Los Angeles County Sanitation District (LACSD) during November 2000 to June 2007 were used to validate vertical profiles of currents.

3.1.3. Remote Sensing Observations

Satellite ocean color measurements for chlorophyll were used to characterize horizontal gradients at finer scales and higher density than possible with the ship-based monitoring. We use monthly averaged surface chlorophyll concentration from the period 1997 to 2000 derived from the SeaWiFS sensor at 4-km spatial resolution. Large gaps in the data set can occur because of dense cloud cover that occurs in late spring and early summer. The products of the Vertically Generalized Production Model (VGPM) net primary production algorithm (Behrenfeld & Falkowski, 1997) were also considered for this validation. Despite limitations, satellite data provide a valuable representation of the spatial distribution of chlorophyll, temperature, and primary production at seasonal scales over the region.

3.2. Performance Statistics

Our approach to a statistical assessment of agreement between model predictions versus observations reflects the fact that the hydrodynamic model, under the influence of realistic forcings (e.g., wind fields) and without data assimilation, develops its own intrinsic variability in circulation, for example, submesoscale eddies (McWilliams, 2007). The resulting modeled state variables would not necessarily overlap with observations on a point-by-point basis but would be comparable to observations when averaged over appropriate spatiotemporal scales. We assessed a suite of statistics and metrics, following the methodology of Allen et al. (2007), to assess how well the model reproduces the magnitude and gradients of selected state variables, whether the model agreement has an apparent bias, and how well the model reproduces natural variability. We calculated six metrics, defined in the following, where N is the total number of appropriate observational data, D represents each individual observational datum, $\bar{D}$ is the mean of the observational data, M is the model estimate representing an observation, and $\bar{M}$ is the mean of the model estimate. The metrics considered include the following.

Table 1
Summary of Statistical Tests of Model Performance and Their Interpretation Used in This Paper

Statistic	Excellent	Good	Reasonable	Poor
Cost function (Moll & Radach, 2003)	<1	1–2	2–3	>3
Nash–Sutcliff model efficiency (Nash & Sutcliffe, 1970)	>0.65	0.65–0.5	0.5–0.2	<0.2
Bias (Maréchal, 2004)	<|0.1|	|0.1–0.2|	|0.2–0.4|	>|0.4|
H (Welch, 1947)	0			1
Correlation coefficient	1–0.9	0.9–0.8	0.8–0.6	<0.6
p-Value	<0.05			>0.05
Ratio of standard deviations	1–0.9, 1–1.1	0.9–0.8, 1.1–1.2	0.8–0.6, 1.2–1.4	<0.6, >1.4

The Pearson correlation coefficient, reflecting the degree of linear correlation between the observed and model variable, and the statistical significance (p-value) of this correlation:

$$r_{xy} = \frac{\sum_{n=1}^{N}(D_n - \overline{D})(M_n - \overline{M})}{\sqrt{\sum_{n=1}^{N}(D_n - \overline{D})^2}\sqrt{\sum_{n=1}^{N}(M_n - \overline{M})^2}}. \tag{7}$$

The cost function (CF), which gives a nondimensional value indicative of the "goodness of fit" between two sets of data, quantifying the difference between model results and measurement data:

$$\mathrm{CF} = \frac{1}{N}\sum_{n=1}^{N}\frac{|D_n - M_n|}{\sigma_D}, \tag{8}$$

where σ_D is the standard deviation of the observations.

The bias (B; the sum of model error normalized by the data):

$$\mathrm{B} = \frac{\sum(D - M)}{\sum D}. \tag{9}$$

The ratio of the standard deviations (RSD):

$$\mathrm{RSD} = \frac{\sigma_D}{\sigma_M}, \tag{10}$$

where σ_M is the standard deviation of model outputs.

The Nash–Sutcliffe model efficiency (ME; Nash & Sutcliffe, 1970), a measure of the ratio of the model error to the variability of the data:

$$\mathrm{ME} = 1 - \frac{\sum(D_n - M_n)^2}{\sum(D - \overline{D})}. \tag{11}$$

And the two-sample t test, or Welch's t test (Derrick et al., 2016; Welch, 1947):

$$H = (\overline{D} - \overline{M})/\sqrt{\frac{\sigma_D^2}{N} + \frac{\sigma_M^2}{N}}. \tag{12}$$

We score the model performance following Table 1 per the methodology of Allen et al. (2007).

4. Model Performance Assessment Findings

4.1. Ocean Circulation

The SCB is situated at the confluence of water masses from the subarctic Pacific via the California Current, and from the eastern tropical North Pacific via the California Undercurrent, which all interact with the local topography, the coast, and the atmosphere to sustain variability in circulation on interannual, seasonal, and intraseasonal time scales (Bograd et al., 2015; Dong et al., 2009). The effects of this variability in circulation have profound consequences for coastal ocean biogeochemistry (Bograd et al., 2015; Gruber et al., 2011; Nagai et al., 2015; Nezlin et al., 2018) and are therefore critical that the model accurately simulates spatial and temporal variability in circulation patterns.

Figure 4 shows the hydrodynamic characteristics of the SCB in the model compared to data. In the northern SCB, the model shows similar qualitative and quantitative patterns for the horizontal circulation compared to HF data (Figures 4a and 4b) and as seen in Dong et al. (2009). The circulation in the SCB is characterized by northward currents in the first 20 km of the coast and cyclonic circulation in the middle of the SCB that is stronger in summer and weaker in winter. The model successfully reproduces observed current patterns, with similar current magnitudes. The intensity of the northward coastal branch of the current is on average about 0.15–0.3 m s^{-1} in summer versus 0.05–0.15 m s^{-1} in winter. The offshore southward branch is generally about 0.3 m s^{-1} all year-round (Figures 4a and 4b). The dominant current in the coastal band (15 km from coast) of the SCB flows northward and follows the topography along isobaths on the shelf (Figures 4g and 4h).

The simulated June 1999 to June 2000 variability of the current in depth is shown in the vertical profiles extracted off the coast of Palos Verdes and Orange County compared to the ADCP data at the same locations (Figures 4c–4f). The location of both of these profiles is a few kilometers from the continental slope and therefore captures a suite of physical processes, including mesoscale and submesoscale eddies, fronts, jets, and internal tides (Capet, McWilliams, et al., 2008; Dong et al., 2009; Kim et al., 2011). The model generally reproduces the means and range of the variability shown in these close to shore horizontal currents, which demonstrates that ROMS at $dx = 0.3$-km resolution captures the submesoscale variability described in Section 2.2.

In the northern SCB, cyclonic vortices are generated inside the Santa Barbara Channel (Figure 4i) when the northward current that flows along the Ventura coast meets the eastern side of the Channel Islands, with higher intensity in summer (Figure 4a vs. Figure 4b; Winant et al., 2003). Submesoscale eddies are particularly prominent in this region, in particular persistent cyclonic eddies that drive an upward doming of isopycnals (Figure 4j; McGillicuddy, 2016), which supplies nutrients to the euphotic layer. The model correctly reproduces this vertical transport, described in Brzezinski and Washburn (2011), and the high concentrations of nitrate and other nutrients in the upper layers of the Santa Barbara Channel, as further detailed in Section 4.3.1.

In the central and southern SCB (latitude <34.7°N), the model successfully captures flow regimes around the large POTW outfalls, indicating that it can appropriately represent the dispersal of wastewater plumes in these regions. In the Santa Monica and San Pedro Bays, topography drives the circulation of currents inside the Bays, converging back to the main current offshore (Figures 4g and 4h). On top of the Hyperion and JWPCP outfalls (in the Santa Monica Bay and offshore of the Palos Verdes peninsula, respectively), the current is mostly south-eastward. Near the OCSD outfall, the current direction varies in winter between south-eastward and north-westward but is primarily southward in summer (Figures 4a, 4b, 4e, and 4f). At the PLWTP outfall, the current is narrow, with a dominant south-eastern direction, parallel to the coast, demonstrated by both model and HF radar data.

4.2. Vertical Gradients and Seasonal Variability of Temperature and Mixed Layer Depth

The model successfully reproduces the three-dimensional and seasonal variability of physical tracers, here exemplified by temperature. Temperature is the parameter in which we have the highest confidence in the observational record, because observations are abundant, and sensors are accurate and precise, regularly calibrated, and with negligible drifts. The greatest source of observational uncertainty is temporal undersampling, but some sources of model bias may also be important (e.g., from atmospheric forcing, wind, or shortwave detailed in Renault et al. [2021]). Quantitative statistical analysis indicates that model performance is "*excellent*" or "*good*" for nearly all metrics for all regions and seasons (see Table 2). The lowest performance of the model is characterized as "*reasonable*" for certain subregions (Palos Verdes, Orange County, and San Diego) in spring and fall (Palos

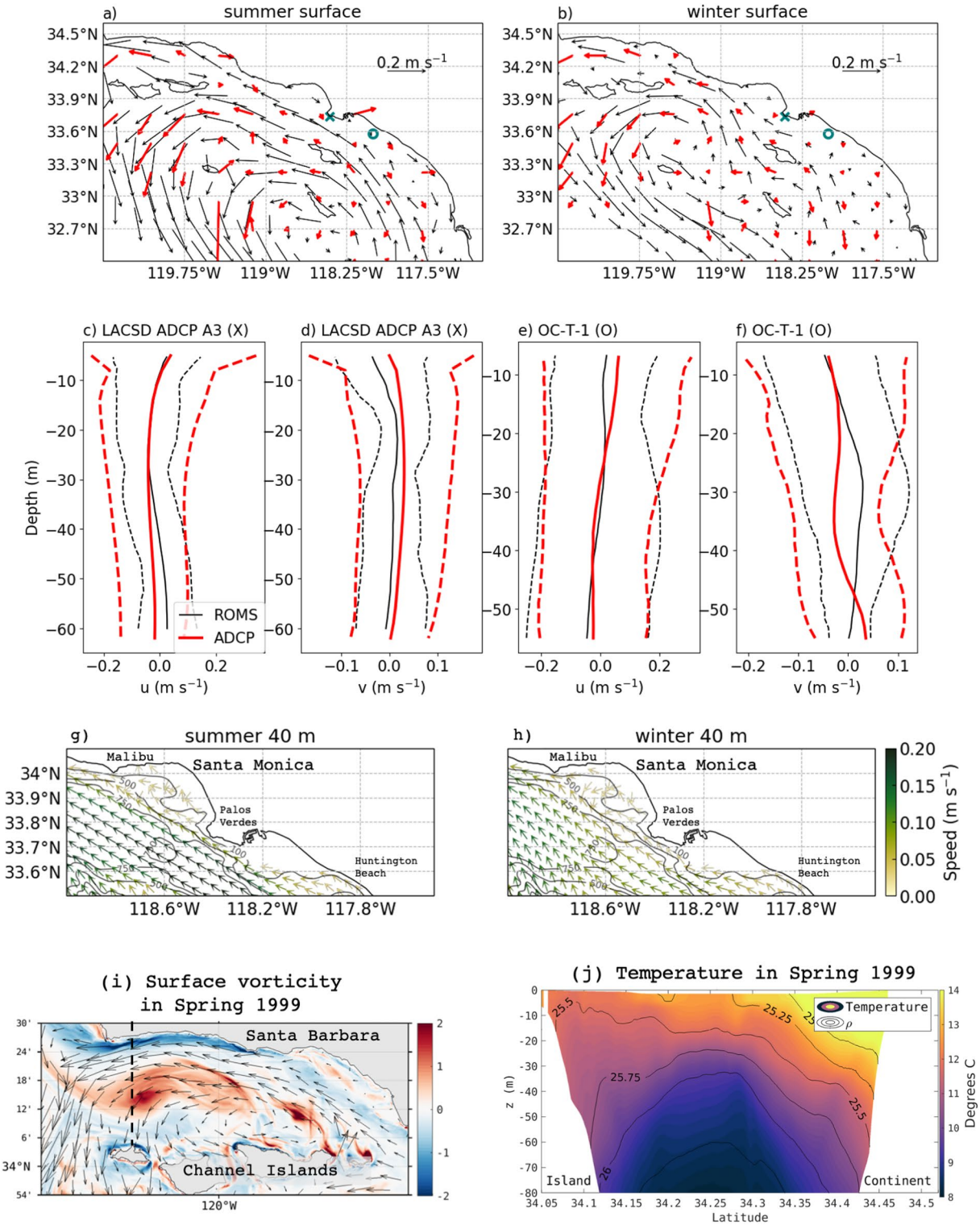

Figure 4. (a) Mean surface currents in the Southern California Bight from high-frequency (HF) data during 2012–2020 (thick red arrows) and model during 1999–2000 (black arrows) in summer and (b) winter. (c–f) Vertical profiles of horizontal velocity components from Acoustic Doppler Current Profilers (ADCP) instruments (thick red lines) and model (thinner black lines). The two dashed lines indicate the 5th and 95th percentile current values. (c, d) ADCP data come from the Los Angeles County Sanitation District (LACSD) mooring A3 stationed at the teal "X" in (a) and (b) and (e, f) come from the Orange County Sanitation District (OCSD) mooring OC-T-1 located at the teal "O." (g) Mean model current direction and speed (colored) at 40 m depth with bathymetry contoured in summer and (h) winter. (i) Surface model vorticity normalized by *f* in spring in Santa Barbara Channel showing cyclonic eddies. (j) Cross section of temperature and density isopycnals as drawn by the dashed line in (i) from model to show eddy-driven uplifting of the isopycnals in the center of Santa Barbara Channel.

Table 2
Statistical Comparison Between In Situ Data and Model Outputs for Temperature Profile in Santa Monica Bay (City of LA Stations)

Santa Monica: temperature								
	H	Correlation coefficient	*p*-value	Cost function	Bias	Ratio of standard deviations	Nash–Sutcliffe model efficiency	Number of observations
Winter	0 E	0.96 E	7E−06 E	0.05 E	−0.04 E	1.10 G	0.81 E	716
Spring	0 E	0.98 E	8E−07 E	0.10 E	−0.10 G	0.78 R	0.51 G	716
Summer	0 E	0.97 E	9E−06 E	0.04 E	−0.02 E	1.07 E	0.93 E	712
Fall	0 E	0.89 G	3E−06 E	0.09 E	−0.08 E	0.98 E	0.51 G	718
All seasons	0 E	0.95 E	3E−05 E	0.08 E	−0.07 E	1.02 E	0.73 E	2,862

Note. Letters next to numbers indicate model performance: E, excellent; G, good; R, reasonable; P, poor.

Verdes only; see Table S2 in Supporting Information S1). As noted above, this may be due to undersampling during these months, which can be highly variable because the region is shifting between a well-mixed to a more stratified ocean regime. Detailed information on the other subregions and their statistical comparison can be found in Tables S1–S4 in Supporting Information S1.

Following common practices (de Boyer Montégut et al., 2004), we define the mixed layer depth (MLD) as the depth at which temperature decreases from its surface value by more than 0.2°C. On average, the MLD deepens from the coast to offshore and varies with season (e.g., in Santa Monica Bay in Figure 5b). The model successfully simulates the seasonal cycle of MLD along the coast. For example, the model recreates the observed seasonal

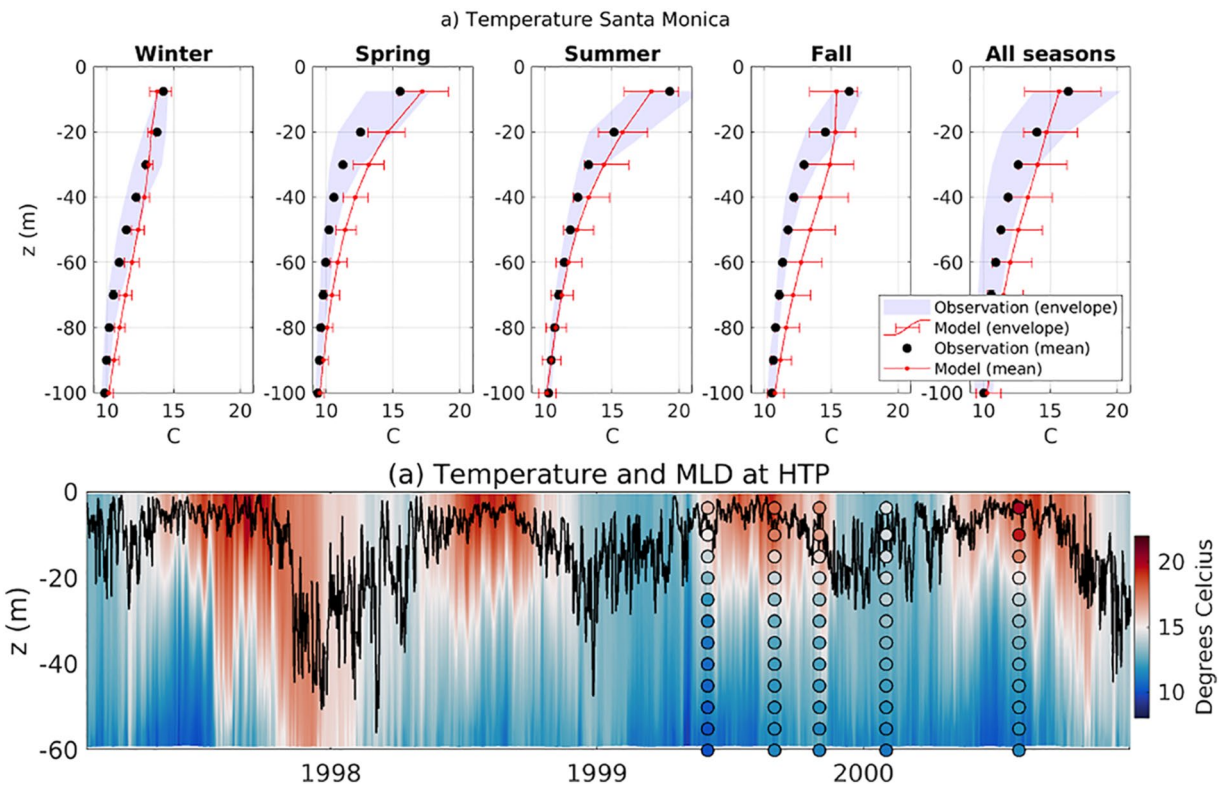

Figure 5. (a) Average seasonal profiles of temperature in the Santa Monica Bay. The red lines and red bars show the spatiotemporal mean and the variability from the model, respectively. The black dots and the gray shading show the spatiotemporal mean and the variability from in situ data (City of LA stations), respectively. (b) Hovmöller diagram of temperature at the location of the Hyperion POTW outfall (Hyperion Treatment Plant [HTP]) in the Santa Monica Bay issued from the model. The black line shows the simulated time series of mixed layer depth. The deepest mixing occurs during El Niño 1998 (>40 m). Colored dots are average concentrations from in situ measurements.

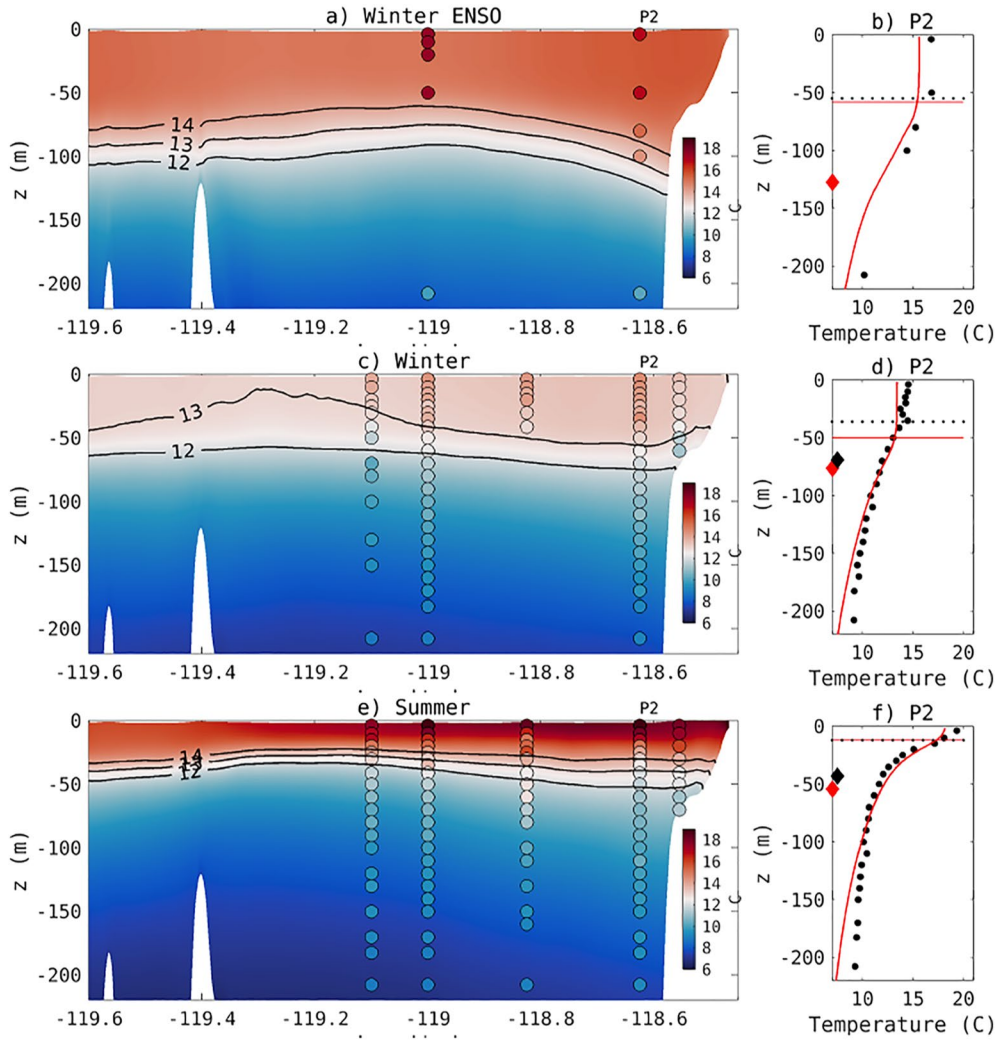

Figure 6. (a) Cross section of average temperature following line 86.7 from CalCOFI monitoring stations during an El Niño winter (December 1997 to February 1998). (b) Profile at station P2. Black dots are CalCOFI in situ data; red line is the simulated profile. The horizontal line is the MLD (black is CalCOFI; red is simulated). Diamonds (black is CalCOFI; red is simulated) is the depth of the maximum gradient to estimate the depth of the seasonal thermocline at 12°C. (c, d) Similar to (a) and (b) for average winter, and (e, f) for average summer.

deepening of the mixed layer in the Santa Monica Bay to depths greater than 16–20 m (the typical depth of the upper signature of the POTW plumes, see Section 4.3.2) nearly every winter (black line in Figure 5b).

Regular winter shows a homogeneous upper layer of <14°C temperature, and a mixed layer located at 18–20 m in the coastal region and 40–60 m offshore. The surface ocean is colder around the Channel Islands (SST <12°C; see Figure 19). In the open ocean, the model reproduces the destratification with deepening of the thermocline to about 70 m and a MLD at about 40 m (Figures 6c and 6d). In summer, stratification is the strongest, reflecting an intense vertical temperature gradient, and the MLD (both in the model and in the observations) is found few meters below the surface (approximately 10 m). Temperature varies rapidly from more than 20°C at the surface in the southern domain (16°C–17°C in the northern domain) to less than 12°C at 50 m depth over the entire SCB (see also Figure 19). In the open ocean, the model succeeds in reproducing the stratification that brings the seasonal thermocline to 50 m and the MLD to 15 m (Figures 6c and 6e).

The model reproduces interannual variability in MLD under the influence of ENSO (hereafter referred to as El Niño, i.e., the period from fall 1997 to spring 1998 in Figure 5b), when the MLD reached 40 m. We show that during winter of El Niño year, the entire water column of the SCB is warmer than on average, and surface

temperature is more homogeneous, varying between 15.5°C and 17°C (Figure 6a). In the open ocean, during El Niño, with warmer upper layer than regular winters, the model shows good performance in reproducing the deepening of the seasonal thermocline (>120 m) and of the MLD (>50 m; e.g., offshore Santa Monica Bay in Figures 6a and 6b). These patterns of variability in temperature are consistent with regional observations of El Niño in the SCB (Todd et al., 2011).

4.3. Dissolved Inorganic Nitrogen

4.3.1. Spatial Patterns and Seasonality of Nitrate

Nitrate observations are only broadly available in the offshore CalCOFI data set, so only large-scale regional patterns in nitrate concentration can be validated. There is a clear seasonality of nitrate, where surface concentrations are higher in spring and summer and decrease in fall and winter (Figure 7). The model reproduces the average seasonal patterns observed in the in situ nitrate data across multiple regions. The model also captures along-shore variability in coastal nitrate concentrations, reproducing values greater than 25 mmol N m^{-3} off Santa Barbara, 20 mmol N m^{-3} off Los Angeles, and 15 mmol N m^{-3} off San Diego.

The model also reproduces observed patterns in the depth of the nitracline (Mantyla et al., 2008; Nezlin et al., 2018), which tends to follow sloping density surfaces in the region. These patterns include the high values at the euphotic depth limit (~50 m below the surface) along the Santa Barbara coast in spring; the doming of the nitracline in the center of the Santa Barbara Channel (Figure 7b); the 20–30 m deep nitracline along the Los Angeles coast; and the deepening of the nitracline from about 30 m at the coast to more than 60 m offshore in San Diego. In the offshore region of the SCB, the model is consistent with observations showing high nitrate (>20 mmol N m^{-3}) around the Channel Islands (not shown) as compared to less than 5 mmol N m^{-3} farther offshore. This pattern is strongest in winter and summer, when the offshore regions are particularly oligotrophic (surface NO_3^- < 1 mmol m^{-3}) throughout the SCB.

4.3.2. Vertical Gradients and Seasonal Variability of Ammonium

Ammonium concentrations above a natural background concentration of 1 mmol N m^{-3} are indicative of POTW wastewater plumes. The model reproduces the observed average vertical profile of ammonium in the Santa Monica Bay, falling within the range of observed variability (Figure S8a in Supporting Information S1). Similar figures for other regions are shown in Figures S6–S9 of Supporting Information S1. All regions show a similar maximum concentration between 30 and 45 m below the surface, in all seasons. The highest concentrations are seen in summer, when stratification is stronger, while lower concentrations in winter likely reflect increased dilution by seasonal mixing from the deepening of the mixed layer (Figure 8b). Near ocean outfalls, both model and observations show middepth peaks of ammonium concentration, occasionally exceeding 10 mmol m^{-3}, which considerably overshadow values observed away from outfalls. In the model, these high ammonium concentrations are caused by wastewater plumes.

The main source of uncertainty in data–model comparisons is the limited spatial and temporal coverage of measurements. Ammonium is typically measured near ocean outfalls and is therefore biased toward high concentrations, but the data set is highly variable. Methodological difficulties exist with the measurement of ammonium in seawater, and as such, we excluded nondetectable ammonium values in our analyses. Near the submarine outfalls, ammonium concentrations are likely extremely heterogeneous due to buoyant plume filaments, as observed in DiGiacomo et al. (2004) and in Warrick et al. (2007) in the Santa Monica Bay, as well as in other regions (e.g., Florida, in Marmorino et al. [2010]) and in idealized case studies (e.g., Ho et al., 2021). These plume filaments are caused by horizontal advection and straining of the discharged effluent by currents. As a result, the undersampling of ammonium may have led to poor statistical agreement between observations and model output. The model shows high to moderate agreement for the shape of the profile and the mean concentration (Table 3). However, *p*-values for the correlations were not always significant. Similarly, there were often biases and low performance regarding variability statistics. This low model performance can be explained by the following two reasons: (a) spatial sampling is likely missing plume filaments, for example observational data points with high ammonium values that are capturing the plume are recorded next to very low or nondetectable values and (b) the resolution of the model (0.3 km), as well as model averaging over the day, season, and depth range causes plume filaments to appear more uniformly spread near the outfalls. Because plume filaments are lost in this averaging, the model represents plumes as cloud-like distributions around outfalls; nevertheless, the average ammonium

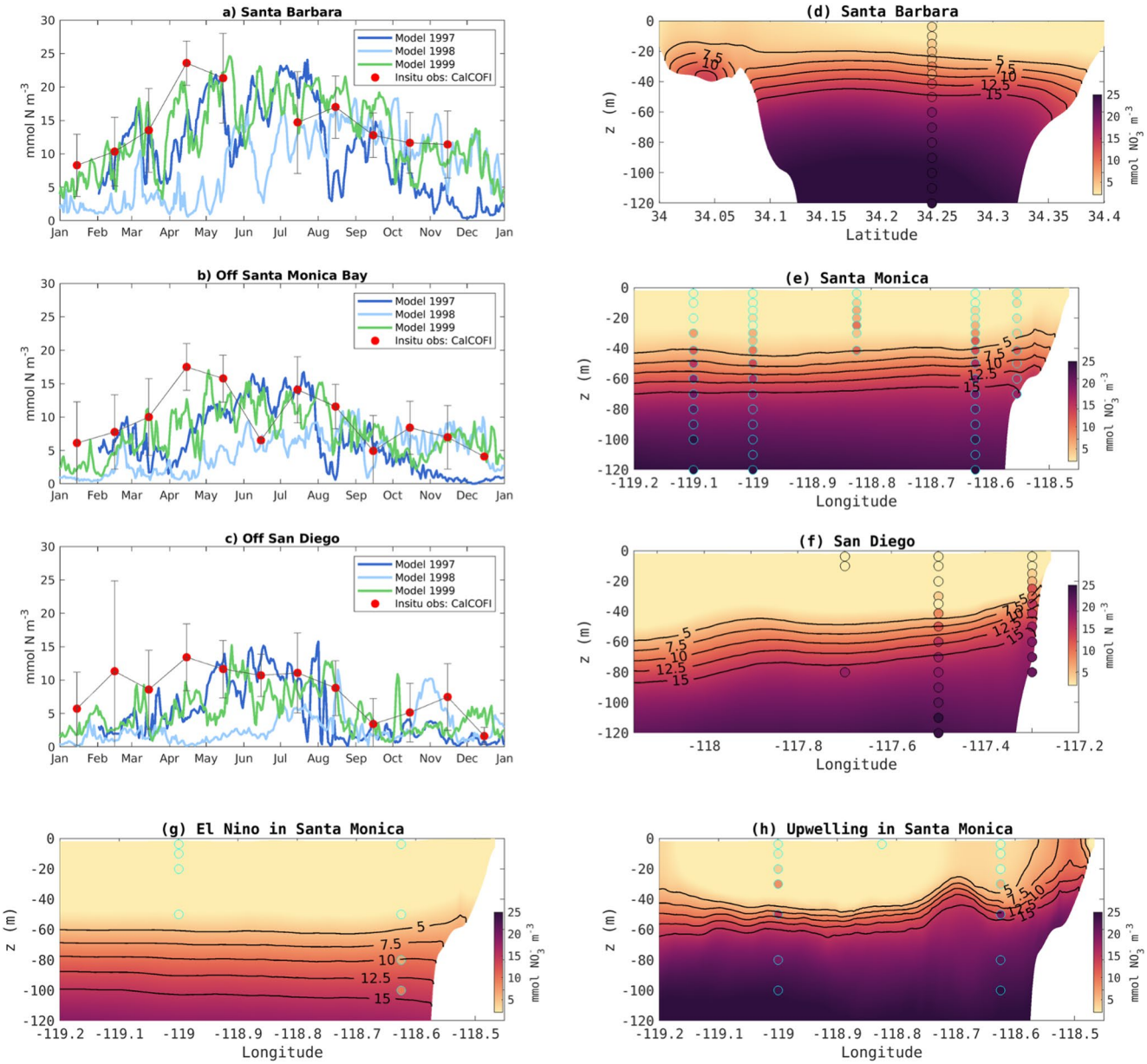

Figure 7. (a–c) Time series of nitrate concentration at 50 m depth in three different locations of the SCB: (a) is near the center of Santa Barbara Channel, (b) is offshore the Santa Monica Bay, and (c) is offshore San Diego. Model outputs are represented by the lines for three different years, with the dots showing mean values from in situ measurement from CalCOFI, and gray bars the standard deviation from the mean. The time series show prominent interannual variability in addition to seasonal variability. While the years 1997 and 1999 show similar nitrate distributions, the El Niño period between the end of 1997 and 1998 is significantly different, showing nearly uniform concentrations between November 1997 and May 1998. This is caused by the deepening of the thermocline during El Niño, which depresses the nutricline. (d–f) Cross sections showing the average springtime nitrate concentration in (d) the Santa Barbara region, (e) the SM region, and (f) the SD region. Background are model outputs and dots are CalCOFI in situ measurements. Model and in situ data agree on the vertical and seasonal patterns in the three regions. They highlight the main differences in these three regimes, consisting of a shallower nitracline in the Santa Barbara Channel, and a deeper nitracline in southern waters. (g, h) Comparison of nitrate concentrations during (g) winter El Niño (January–March 1998) and (h) during an upwelling event (the first week of May 1999) to illustrate the ability of the model (vs. in situ CalCOFI data) to simulate the vertical displacement of the nitracline during these specific events.

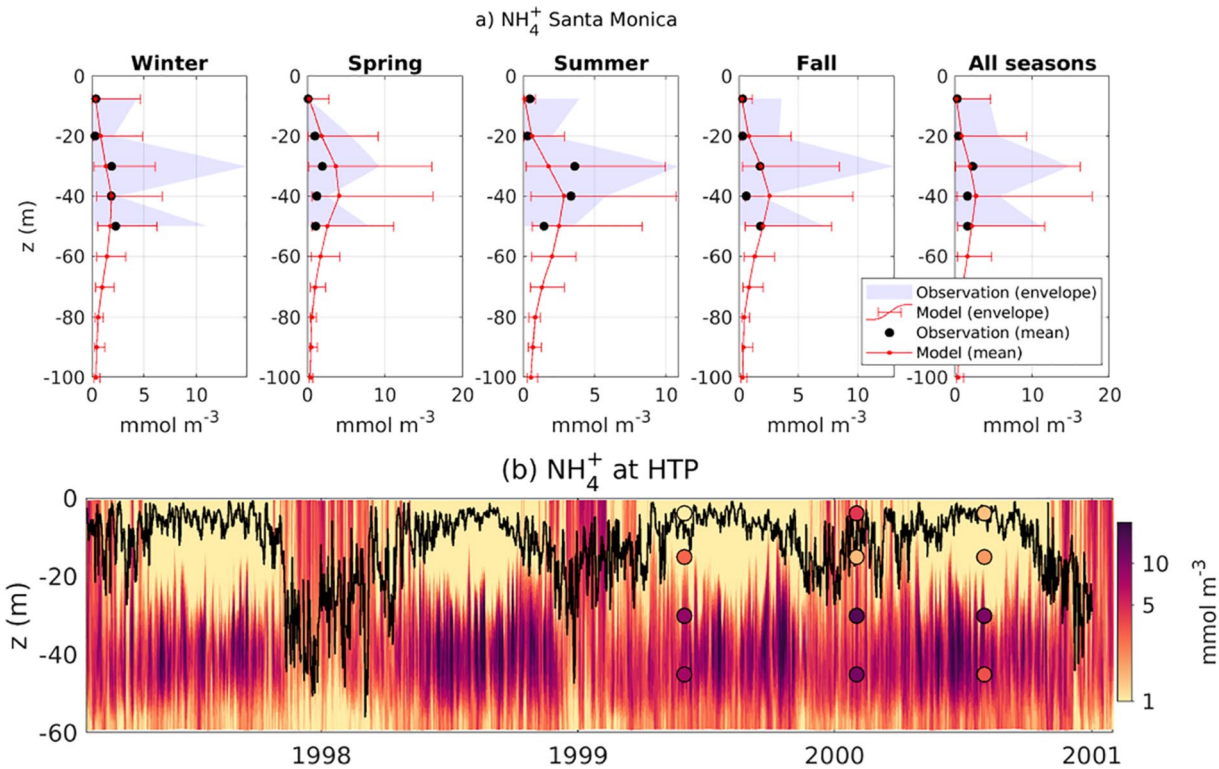

Figure 8. As for Figure 5, but for ammonium concentration. These profiles are showing agreement on intensity, seasonality, and shape of the vertical profile with exceptionally high concentrations at middepth.

concentration of wastewater plumes is reasonably well represented. Detailed information on the other subregions and their statistical comparison can be found in Tables S1–S4 of Supporting Information S1.

4.3.3. Horizontal Gradients of Ammonium

Both in situ observations (dots in Figures 9 and 8a) and model output (background colors in Figure 9 and red line in Figure 8a) show high concentrations of ammonium in the subsurface layer below the thermocline (Figure 9c), which we refer to as "high-ammonium plume." This high-ammonium plume can extend from Huntington Beach to South Ventura, encompassing three of the four major wastewater treatment plant outfalls in the SCB (see Section 2.4). Both model and observations show that the width and strength of the high-ammonium plume are greatest in summer compared to other seasons. The SMBO mooring (Leinweber et al., 2009) located 17 km northwest of the submarine pipe Hyperion in Santa Monica Bay (Figure 9g) frequently recorded concentrations higher than 2 mmol m^{-3}, and up to 4 mmol m^{-3} at middepth (Figure 9e), consistent with the model (Figure 9f). The depth of the maximum variability is at 40 m in the model, and slightly shallower in the SMBO data, possibly because

Table 3
Statistical Comparison Between In Situ Data and Model Outputs for Ammonium Profile in Santa Monica Bay

Santa Monica: ammonium								
	H	Correlation coefficient	*p*-Value	Cost function	Bias	Ratio of standard deviations	Nash–Sutcliffe model efficiency	Number of observations
Winter	0 E	0.94 E	0.06 P	0.54 E	0.24 R	1.86 P	0.68 E	20
Spring	0 E	0.85 G	0.14 P	0.58 E	−0.57 P	0.69 R	−0.61 P	21
Summer	0 E	0.58 P	0.42 P	0.72 E	0.19 G	1.76 P	0.29 R	21
Fall	0 E	0.91 E	0.09 P	0.42 E	0.07 E	1.47 P	0.80 E	21
All seasons	0 E	0.81 G	0.10 P	0.36 E	−0.03 E	1.23 R	0.60 G	83

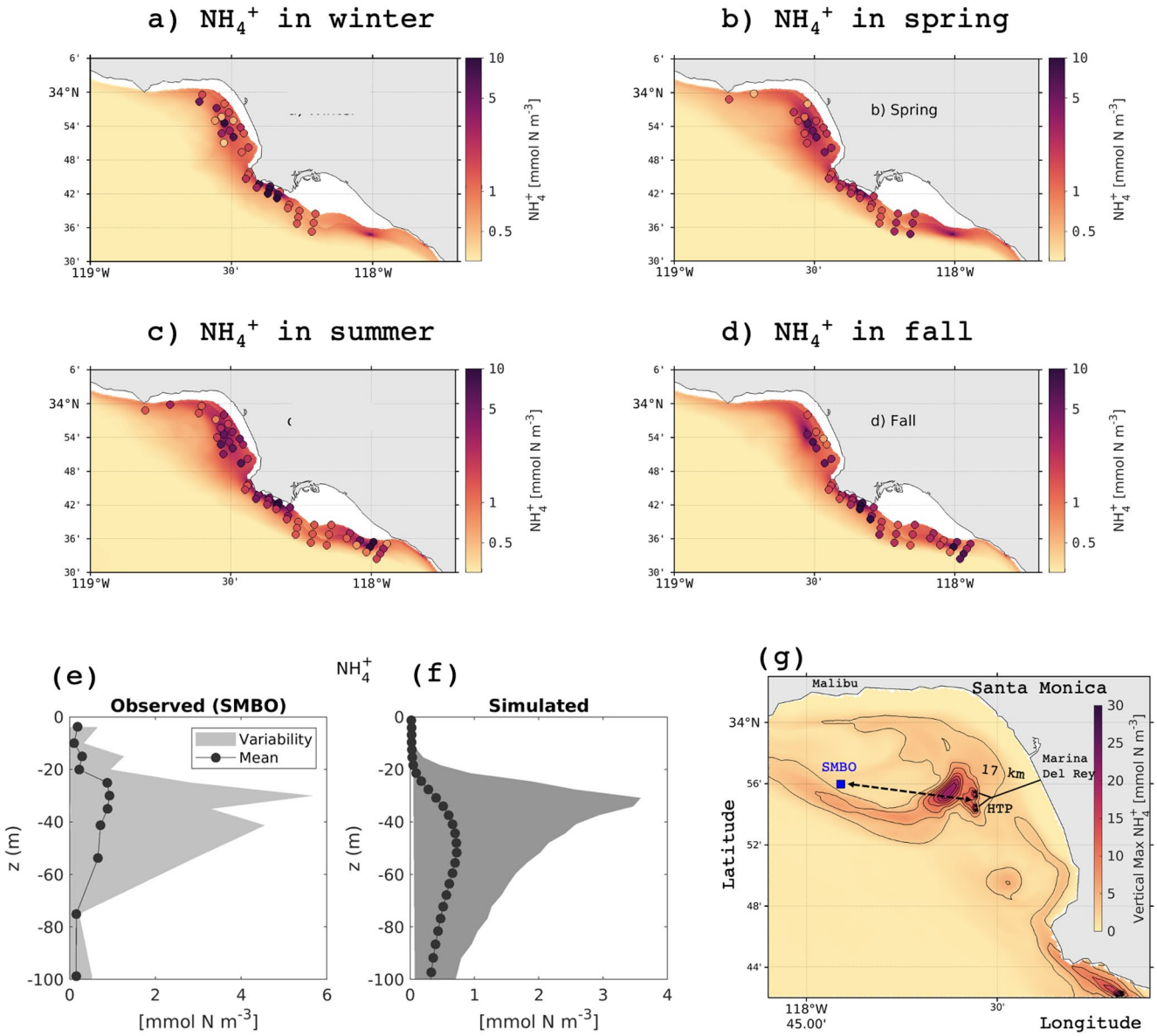

Figure 9. (a–d) Seasonal average ammonium concentration between 30 and 45 m depth from the model and dots from observations. High values highlight the movement and dispersion of subsurface wastewater plumes along the Orange and Los Angeles Counties. The highest concentrations are located within a narrow coastal band of about 10–15 km width and are carried along the topography by the mean currents. (e, f) A statistical comparison of the vertical profiles of ammonium at the SMBO mooring and the same location in the model. The anthropogenic ammonium plume signature is apparent, albeit intermittently, 17 km away from the Hyperion outfall. (g) The simulated vertical maximum concentration of NH_4^+ averaged during a representative day to illustrate the dispersal of the effluent toward SMBO originating from the two diffusers of Hyperion Treatment Plant (HTP).

of a mismatch in the time period (1997–2000 for the model, and 2004–2010 for the SMBO). During winter, the model indicates vertical mixing and dilution of the plume at the surface. Accordingly, ammonium concentrations decrease slightly at depth (Figure 9a) and increase at the surface, reaching values up to 2–6 mmol m^{-3}, also consistent with observations around the outfall pipes (Figure 8a).

4.3.4. Spatial Patterns in Rates of Nitrogen Transformation

Although we had no in situ nitrogen transformation rates with which to compare model output during the simulation period, several data sets exist for the region that can serve as a test for whether the model is simulating reasonable patterns in rates via the right mechanisms. We found that modeled rates do agree with observed nitrogen transformation rates. Nitrification rates, the sequential oxidation of NH_4^+ to NO_3^- via NO_2^-, have been

observed to be higher within wastewater plumes in the SCB (McLaughlin et al., 2021), a pattern driven by high ammonium concentrations in the discharges (McLaughlin, Nezlin, et al., 2017). In both observations and the model, nitrification predominately occurs below the euphotic layer. Modeled vertically integrated nitrification rates vary between 0.15 and 1.5 mmol N m^{-2} day^{-1}, consistent with observations within the SCB and in the California Current (Table 5). The model also reproduces higher nitrification rates within wastewater plumes (see Figure S22 in Supporting Information S1). There is also good agreement between observed and modeled rates of nitrate and ammonium uptake by phytoplankton communities (Kudela et al., 2017; McLaughlin et al., 2021). Modeled nitrate uptake rates vary between 2 and 11 mmol N m^{-2} day^{-1} and ammonium uptake rates vary between 6 and 51 mmol N m^{-2} day^{-1} in the Los Angeles and Orange County coasts, consistent with observations in the SCB (Table 5).

4.4. Chlorophyll Concentrations

In general, the model was found to reproduce vertical and horizontal gradients in chlorophyll concentration in different subregions (Figure 12). The timing of blooms was consistent with changes in mixing and nutrient delivery in the SCB. We present three different subregions characterized by distinct hydrodynamic regimes: the Santa Barbara Channel, the Los Angeles coast, and San Diego coast.

There are several sources of uncertainty in the chlorophyll, primary production, phytoplankton growth, and grazing rates observational records. For chlorophyll, bottle measurements are accurate and precise but measure a limited portion of the water column. Sensors are accurate and precise in their measurement of fluorescence and have a rapid response time, providing vertically resolved profiles; however, the algorithm to convert fluorescence to chlorophyll concentration is inaccurate for the SCB. As a result, a correction factor has been applied to Bight data which adds uncertainty to the observational data set (Nezlin et al., 2018). Satellite measurements of chlorophyll are inferred from ocean color (Kahru et al., 2009). This method works well offshore but breaks down nearshore where terrestrially derived colored dissolved organic matter creates uncertainty in reported satellite chlorophyll estimates on the order of 100% or greater (Zheng & DiGiacomo, 2017). For primary production, the incubation method to derive the rates is sensitive and precise (Cullen, 2001), though measured rates are subject to bottle effects and there is some ambiguity as to whether the experiments measure net primary production or gross primary production (Regaudie-de Gioux et al., 2014). Phytoplankton growth and zooplankton grazing are also determined experimentally, and duplicate measurements indicate that these methods are not very precise, with differences between duplicates ranging from 80% to 200% (Landry et al., 2009; Li et al., 2011). For all three measurements, spatial and temporal undersampling, particularly during seasons with high variability, adds uncertainty to the data–model comparison.

4.4.1. Horizontal Gradients in Chlorophyll

Despite the uncertainties outlined above, the model successfully simulates horizontal gradients in chlorophyll in the three subregions (Santa Barbara, Los Angeles, and San Diego). The model captures the early, wide-spread spring bloom in the Santa Barbara Channel, which occurs as a combination of a coastal bloom driven by spring upwelling, followed by a bloom in the central and southwestern regions of the Channel (near the islands) in spring and summer (Figure 10). The latter is driven by the strengthening of the cyclonic circulation in the Channel, which transports nutrients to the upper layers, and is regularly observed in the region (Brzezinski & Washburn, 2011). The model captures the strong seasonality in chlorophyll, wherein concentrations change from near zero in winter to up to 8 mg Chl m^{-3} in spring. Of the three regions, the blooms off Santa Barbara extend further into late summer and fall, where the average concentration is approximately 1–2 mg Chl m^{-3}, a pattern replicated in both model and observations. Spatially, the model correctly reproduces the main patterns observed in satellite-based reconstructions, with spatial correlation coefficient varying between 0.5 and 0.9, and a cost function demonstrating excellent scores. The bias is also excellent (<0.05) in all seasons except winter, when it is reasonable, potentially because of weaker spatial gradients than observed.

In the Los Angeles subregion, the model predicts broad patterns in chlorophyll concentrations with good spatial correlation coefficients across the seasons (0.75–0.89), including a persistent bloom in the San Pedro Bay, consistent with in situ observations (Nezlin et al., 2012), and remote sensing (Figure 11). Spatially, the cost function shows excellent scores across all seasons, and the bias is good to reasonable. Both satellite-derived and modeled chlorophyll show concentrations in the San Pedro Bay consistently higher than 3 mg Chl m^{-3} year-round, often

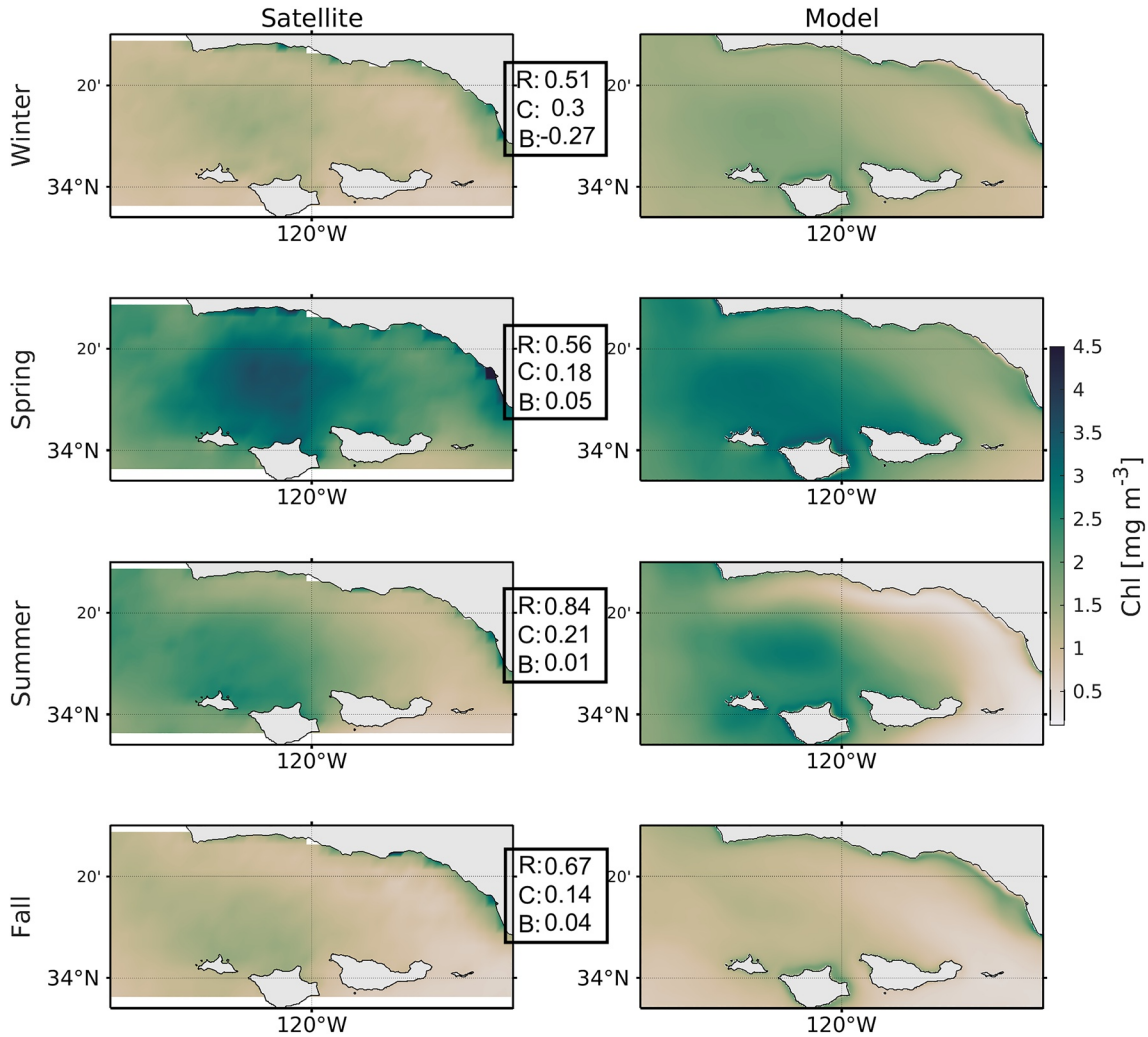

Figure 10. Comparison of seasonally averaged surface chlorophyll between SeaWiFS remote sensing data (left panels) and the model (right panels) in the Santa Barbara Channel, where an important seasonal bloom is observed. The three numbers for each season represent statistics of spatial comparison between the observed and simulated chlorophyll: Pearson's correlation (*R*), cost function (C), and bias (B).

extending into the Santa Monica Bay. The model also reproduces the strong offshore gradients in chlorophyll, where across less than 15 km offshore surface concentrations are reduced 3–4-fold (<1 mg Chl m^{-3}) further decreasing toward the open ocean. The model also reproduces the timing and magnitude of the blooms in the Santa Monica and San Pedro Bays. The difference in timing of maximum chlorophyll concentrations between the Santa Monica and San Pedro Bays likely reflects differences in nutrient supply. Nutrients, in particular ammonium, are available near the surface during winter (see Section 4.3.2), reflecting more vigorous mixing of the wastewater plume and land-based nutrient supply by rivers (in particular in the San Pedro Bay) during winter storms (Lyon & Stein, 2009). Storms and winter mixing events have been connected to phytoplankton blooms in the region (Mantyla et al., 2008; Nezlin et al., 2012). Further offshore in the Los Angeles region, the model recreates the weak seasonality of surface chlorophyll, with higher concentrations during winter and spring, and lower concentrations in summer and fall. In the offshore region of the Santa Monica Bay, the seasonal cycle is marked by the increase of surface phytoplankton between March and May as shown in Figure 12b. Mean chlorophyll values reach up to 3–4 mg Chl m^{-3} in April and May, although concentrations below 2 mg Chl m^{-3} are more common, consistent with observations over the same period.

Offshore of the San Diego coast, the model recreates a slight increase in surface chlorophyll in March; however, concentrations are generally below 1 mg Chl m^{-3} year-round (Figure 12c). The oligotrophic conditions of the southern Bight (Mantyla et al., 2008; Nezlin et al., 2012) have been attributed to a deeper nitracline, which in

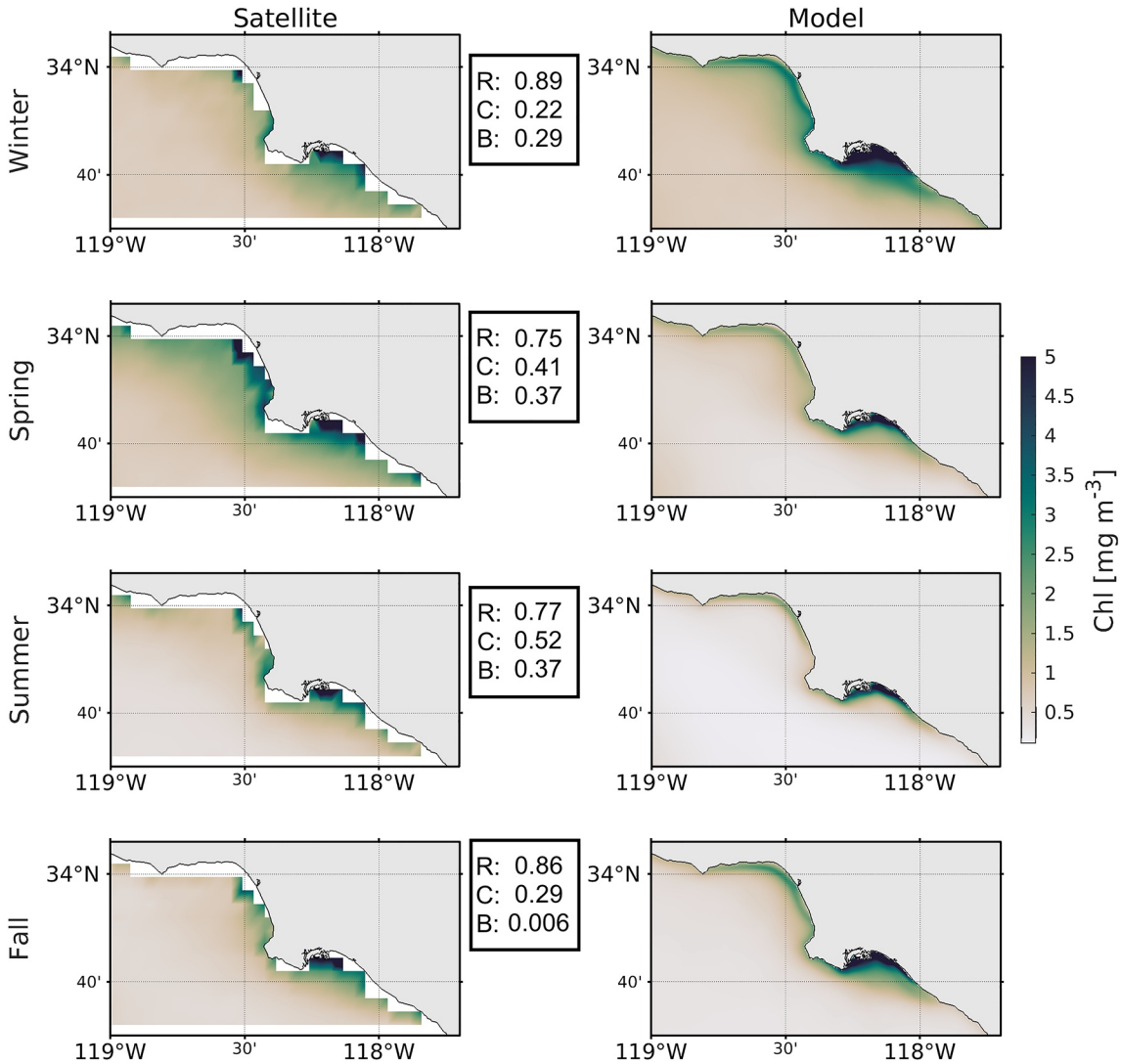

Figure 11. Comparison of seasonally averaged surface chlorophyll between SeaWiFS remote sensing data (left panels) and the model (right panels) for years 1998–2000 in the Santa Monica and San Pedro Bays, where major POTW outfalls are found. The figure highlights the persistent coastal phytoplankton bloom and the sharp inshore–offshore gradients. The three numbers for each season represent statistics of spatial comparison between the observed and simulated chlorophyll: Pearson's correlation (*R*), cost function (C), and bias (B).

turns supports a deep chlorophyll maximum layer (Mantyla et al., 2008). This feature is well represented in the model, which reproduces relatively high concentrations of chlorophyll in subsurface layers (generally between 20 and 90 m depth in the region).

4.4.2. Vertical Gradients and Seasonal Variability of Chlorophyll

The goodness-of-fit statistical metrics (correlation coefficient and cost function) for chlorophyll are generally *excellent* or *good* for most seasons for all subregions (Table 4). We were most concerned with performance for these metrics because the remaining statistics may be affected by the aforementioned uncertainties due to the fluorometry calibration. The observational measurements should be internally consistent (if not accurate), so the shapes of profiles should be "correct" even if the magnitude is off due to poor calibration, and the model was able to replicate these shapes accurately. Despite calibration issues, the model reproduced chlorophyll reasonably well for the northern Bight subregions of Santa Monica Bay (Figure 13) and Ventura/Oxnard (Figure S11 in Supporting Information S1). Similar figures for other regions are shown in Figures S10–S13 of Supporting Information S1. All show that the model is reproducing the magnitude and general shape of observed profiles. However, the model did not capture the variability for most regions (except for Palos Verdes), generally scoring *reasonable*

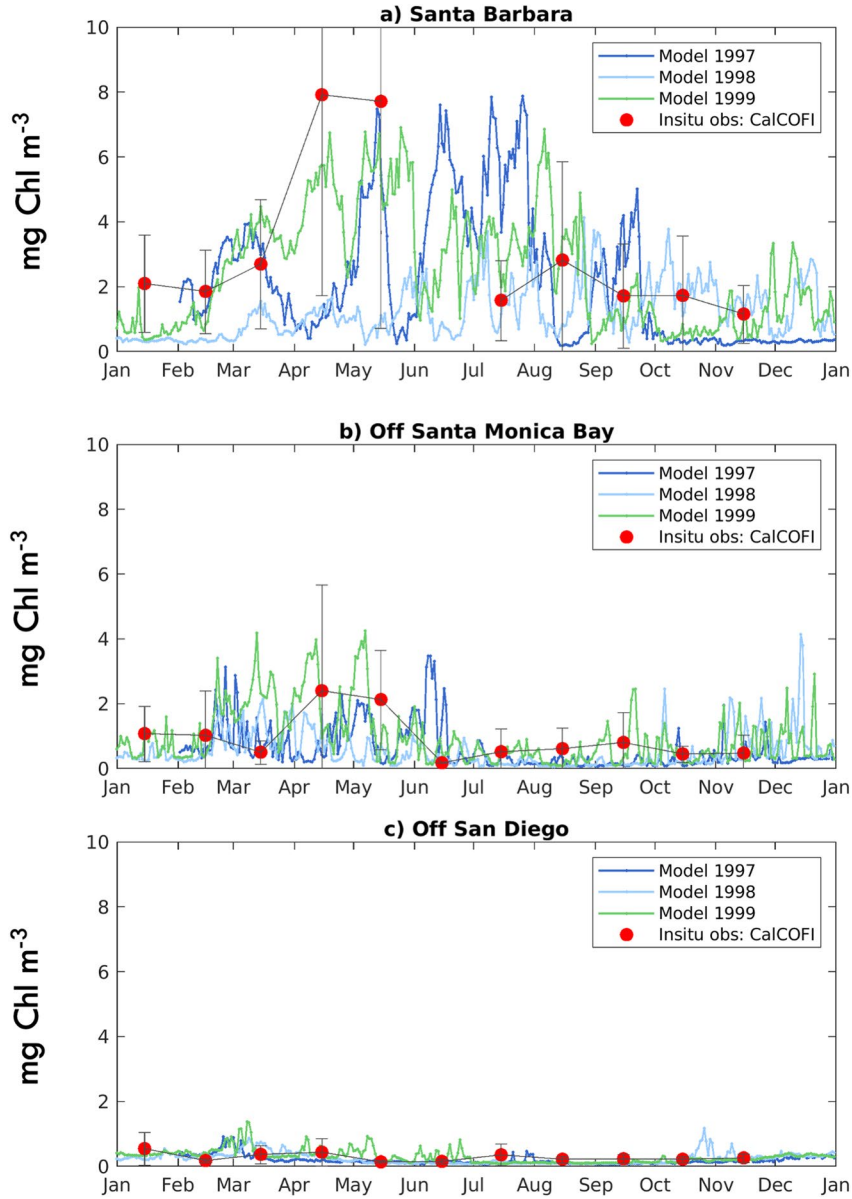

Figure 12. Comparison of surface chlorophyll concentration between different years of model output and a climatology from CalCOFI in situ data. (a) Near the center of Santa Barbara Channel, (b) offshore the Santa Monica Bay, and (c) offshore San Diego. The model reproduces different productivity regimes across the Southern California Bight, with highly productive waters in the northern region, where average concentrations greater than 3 mg m^{-3} are observed for more than half of the year, and oligotrophic southern regions, where average surface concentrations rarely exceed 1 mg m^{-3}.

or *poor* in the ratio of standard deviations for most seasons, particularly spring. This is likely a result of the spatial and temporal averaging. Chlorophyll is highly variable in space and time and undersampling in either of these dimensions will adversely affect variability estimates for a region and season. Therefore, reasonable performance for these metrics was not unexpected. This suggests that the model may provide a conservative estimate of phytoplankton biomass in the southern Bight, while reproducing accurate spatial and temporal patterns in that biomass.

In addition to transporting nutrients from depth, upwelling "seeds" surface waters with subsurface water masses dominated by selected phytoplankton species, stimulating surface blooms near the coast (Seegers et al., 2015). The model successfully reproduces this process, wherein the subsurface chlorophyll maximum shoals and intensifies in spring, forced by the vertical movement of the thermocline driven by upwelling. This seasonal dynamics occurs across the domain in the model.

Table 4
Statistical Comparison Between In Situ Data and Model Outputs for Chlorophyll Profile in Santa Monica Bay

Santa Monica: chlorophyll								
	H	Correlation coefficient	*p*-value	Cost function	Bias	Ratio of standard deviations	Nash–Sutcliffe model efficiency	Number of observations
Winter	0 E	0.99 E	9E−06 E	0.48 E	0.09 E	0.91 E	0.94 E	714
Spring	0 E	0.93 E	9E−05 E	0.90 E	−0.42 P	0.52 P	−0.49 P	716
Summer	0 E	0.99 E	1E−08 E	0.58 E	−0.07 E	0.60 R	0.47 R	712
Fall	0 E	0.99 E	8E−08 E	0.48 E	0.16 G	0.75 R	0.76 E	718
All seasons	0 E	0.99 E	4E−08 E	0.50 E	−0.01 E	0.73 R	0.80 E	2,860

Offshore, in the more oligotrophic portion of the SCB, the model predicts that more than 60% of the maximum concentration of phytoplankton biomass remains below the surface year-round, constantly fed by subsurface nutrients injections. This is consistent with observations of a deep chlorophyll maximum throughout the region (Mantyla et al., 2008; Nezlin et al., 2018; Seegers et al., 2015), and with observations at SPOT located between the Palos Verdes Peninsula and Catalina Island (Figure 3, lower panel). At SPOT, a region weakly influenced by anthropogenic nutrients inputs at the surface, the model realistically simulates the seasonal cycle of chlorophyll. While ammonium does not exceed typical "natural" values of ∼1 mmol m^{-3} below the surface, chlorophyll concentrations regularly reach more than 2 mg m^{-3} between 20 and 40 m in summer, in agreement with in situ measurement (Beman et al., 2011; Teel et al., 2018). (Additional figures to support the analysis are reported in Figure S23 of Supporting Information S1.)

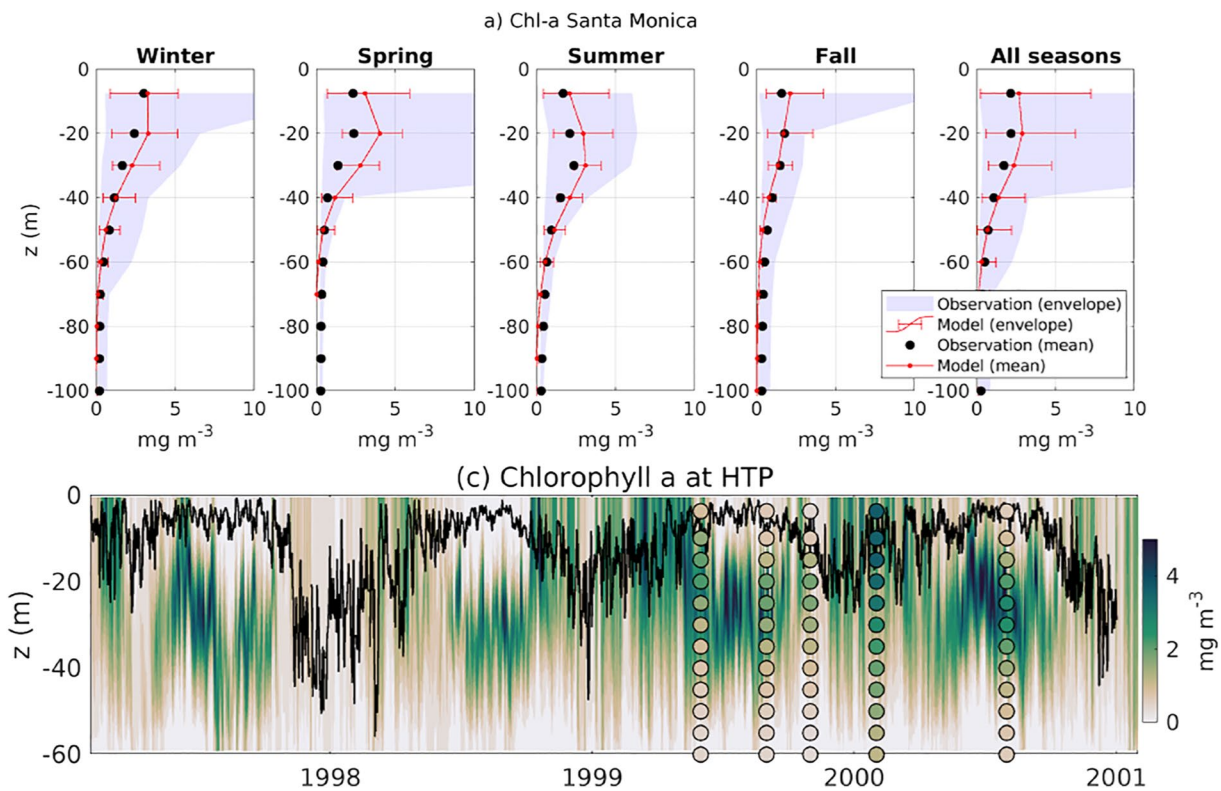

Figure 13. As for Figure 5, but for chlorophyll concentration. Vertical profiles show a good agreement between simulated and in situ data and display the formation of a subsurface chlorophyll maximum in summer, and a surface maximum in winter and spring. Concentrations in winter vary up to +5 mg Chl m^{-3}. Note the very low concentrations during 1998 El Niño in the entire water column.

Table 5
Comparison of Biogeochemical Rates Between Published Literature and Model

	Bight 13	Literature	Model
Primary production (g C m^{-2} $year^{-1}$)	47.4, 1,037.4		250, 1,660
Nitrification (mmol m^{-3} day^{-1})	0, 0.225	0.02, 0.08	0.001, 0.27
NO_3^- uptake rate (mmol N mg Chl^{-1} day^{-1})	0.005, 2.16		0.03, 0.15
NH_4^+ uptake rate (mmol N mg Chl^{-1} day^{-1})	0.10, 8.30		0.08, 0.15
Total phytoplankton growth μ (day^{-1})		0.05, 0.8	0.3, 0.4
Grazing (day^{-1})		0.02, 0.5	0.3, 1.5

Note. Values are minimum and maximum. Literature values come from Landry et al. (2009) and Li et al. (2011). Bight 13 is extracted from McLaughlin et al. (2021) study.

However, in regions more heavily influenced by anthropogenic nutrients, such as the Santa Monica Bay, the chlorophyll maximum progressively deepens from the surface in winter to about 25–30 m depth in spring and summer, with chlorophyll concentrations exceeding 5 mg Chl m^{-3} (Figure 13a). This subsurface chlorophyll maximum is maintained for 4–5 months (Figure 13b) before the stratification is weakened by winter mixing.

4.4.3. Primary Production

Validation of the rates of primary production, phytoplankton growth, and zooplankton grazing (Table 5) provides an independent check on mechanisms responsible for chlorophyll as a state variable. The spatial and temporal frequency of these data, garnered from CalCOFI observations and literature values, is low. The most data as well as the most standardized methodologies are available for primary production. However, many of the primary production measurements used in this validation do not temporally coincide with the model period. Despite these uncertainties, the model generally reproduces expected large-scale patterns and seasonal variability in primary production.

This large-scale variability was also mentioned in Deutsch et al. (2021). Model and data both show lower productivity in winter (Figures 14a and 14c) and higher in spring (Figures 14b and 14d), when the primary production is high along the coastal band, in the northern Bight around the Channel Islands (Figure 14d), consistent with observations (Figure 14b). This is also consistent with the so-called "green ribbon" of high chlorophyll observed along the coast throughout the SCB (Lucas et al., 2011). The model reasonably reproduces the seasonal cycle of primary production in each of the subregions.

Phytoplankton are generally limited by a combination of nutrients and light, the latter of which is only limiting at depth in the SCB (Deutsch et al., 2021).

In winter, nitrogen is high at the surface in the northern SCB, and thus is not limiting. In the southern SCB, light and nitrogen are colimiting due to stronger stratification, leading to oligotrophic conditions. In spring and through the summer, nitrogen is limiting nearly everywhere except in the Santa Barbara Channel and near the Channel Islands, where upwelling and submesoscale eddies maintain high nutrients at the surface.

The scatter plots in Figures 14e and 14f show a comparison of the simulated primary production between the in situ CalCOFI data and that derived from remote sensing (empirically adjusting the Behrenfeld–Falkowski VGPM). The model shows a correlation coefficient of about 0.6 with CalCOFI, similarly to that reported by Kahru et al. (2009) when comparing the VGPM product with CalCOFI. The model shows a stronger correlation with VGPM data, with a correlation coefficient of the order of 0.8.

Finally, while slightly outside our model domain and simulation period, the modeled phytoplankton growth and zooplankton grazing rates were within the same order of magnitude as the measured rates from the California Current Long Term Ecological Research (CC-LTER) project (see Landry et al., 2009; Table 5) in the northern portion of the Bight.

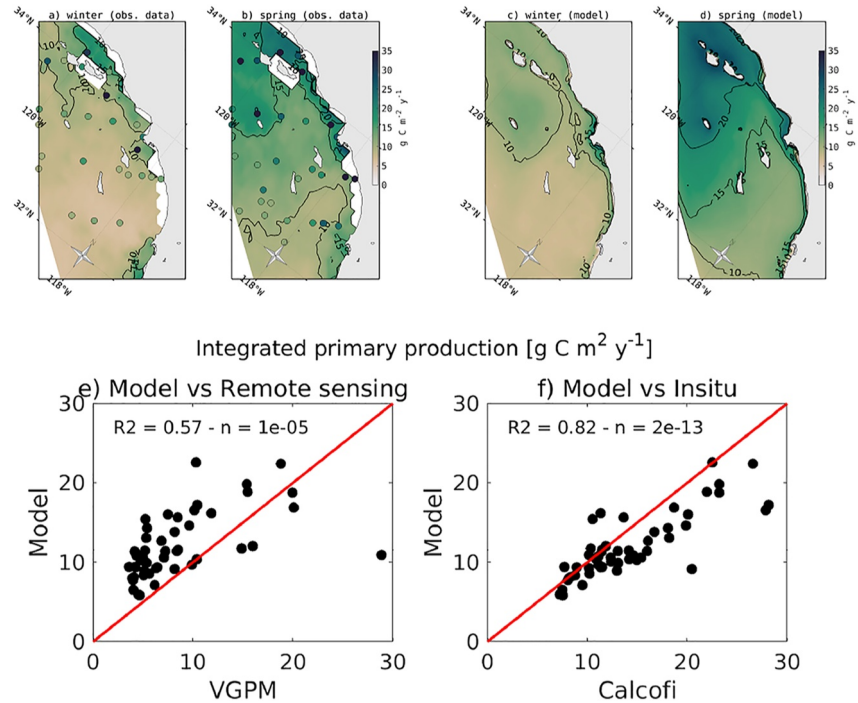

Figure 14. (a, b) Maps of vertically integrated Vertically Generalized Production Model (VGPM) net primary production and CalCOFI in situ measurements plotted as dots for (a) winter (January and February) and (b) spring (April–June). (c, d) Maps of vertically integrated primary production from the model, in (c) winter and (d) spring. Note the higher values for CalCOFI in situ measurements as compared to the satellite estimate, in better agreement with the model.

4.5. Carbonate System and Oxygen Parameters

The model predicts changes in dissolved oxygen and carbon system parameters related to photosynthesis and respiration, as well as horizontal transport and vertical mixing. As described in Section 4.4.1, the coasts of Los Angeles and Santa Barbara are hot-spots of intensified plankton activity, and both systems are impacted by high variability and small-scale eddy circulation. In the upper layers, photosynthesis increases both dissolved oxygen and pH (Figures 16–18), consistent with observations in these regions. The Santa Monica Bay shows the highest oxygen production rates (60 mmol m^{-2} day^{-1}), followed by the Santa Barbara coast (57 mmol m^{-2} day^{-1}), while rates in the Orange County and San Diego coasts are nearly 2 times lower. Oxygen and carbon are further replenished at the surface by air–sea gas exchange with the atmosphere. Export of newly fixed organic carbon leads in both regions to high remineralization rates that consume oxygen and release carbon dioxide at depth. We simulate similar high organic matter export (around 30 mmol m^{-2} day^{-1}) in the Santa Barbara and Los Angeles coasts (see Figure S24 in Supporting Information S1).

The reliability of these predictions can be tested by validation of dissolved oxygen and carbonate system parameters. There are several sources of uncertainty in the dissolved oxygen, pH, and aragonite saturation state observational records, which affect data–model comparisons. For dissolved oxygen, sensors are relatively accurate and precise and have a rapid response time (<1 s) when generating vertically resolved profiles. Repeated field measurement accuracy for CTD dissolved oxygen sensors was reported to be approximately 8 mmol m^{-3} (Coppola et al., 2013). The pH observational record is particularly fraught with uncertainty. An evaluation of pH sensor data in the SCB indicated that, while sensor pH measurements were well correlated with discrete bottle samples collected at the same depth, there was a clear bias in pH, with sensor measurements underpredicting bottle measurements and high variability in the differences between paired bottle and sensor measurements (ΔpH ranging from ±0.5; McLaughlin, Dickson, et al., 2017). The aragonite saturation state is estimated using an algorithm developed for the region (Alin et al., 2012) for both in situ observations and model output, because complete measurements of carbon system parameters required to calculate Ω_{Ar} are missing. For all three variables, spatial and temporal undersampling, particularly during seasons with high variability, adds uncertainty to the data–model comparison.

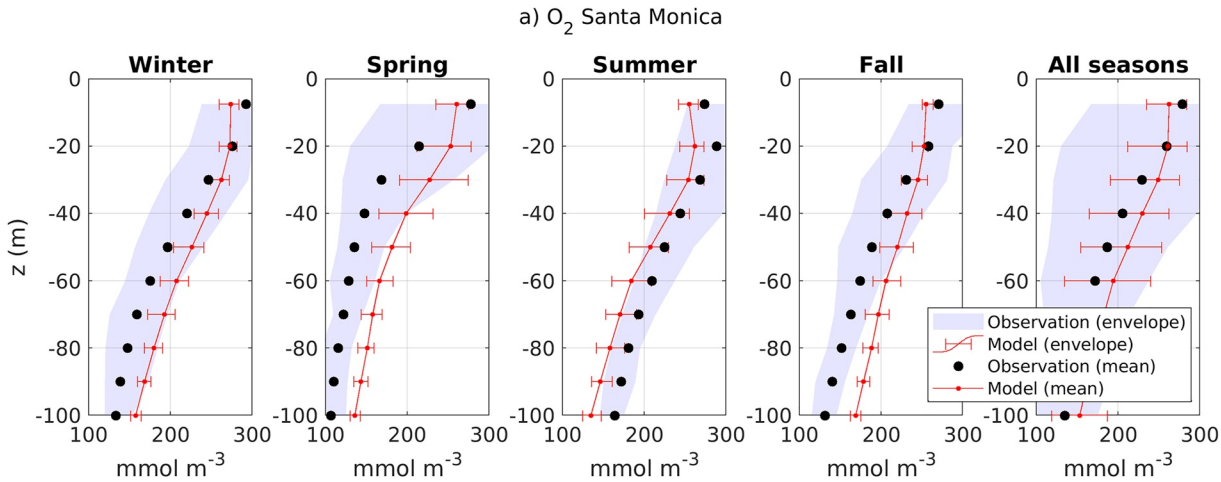

Figure 15. As for Figure 5, but for oxygen concentration.

4.5.1. Vertical Gradients and Seasonal Variability of Dissolved Oxygen

The model reproduces observed seasonal and spatial patterns in dissolved oxygen concentration (Figure 15), accurately simulating magnitude, vertical and horizontal gradients, and variability. Quantitative statistical analysis (see Table 6) indicated that the model performance was "excellent" or "good" for nearly all metrics for all regions and seasons. The lowest performance of the model was characterized as "poor" for two subregions for the Nash–Sutcliff model efficiency during spring and "reasonable" for some metrics in some subregions, which may be related to undersampling during seasons with high variability, as described above. Similar to temperature, we tested whether the variability in spring may be impacting the performance statistics by extracting random profiles for the region (not shown, expressed with large error-bars in the spring season plots in Figure 16), which show how dissolved oxygen on a single arbitrary day can more closely align with the observations. This supports the hypothesis that observational uncertainty is behind the lack of observational agreement with the model. Model performance was lowest in the Orange County and San Diego subregions, where model predictions tended to overestimate dissolved oxygen, consistent with the chlorophyll underprediction, a likely consequence of the lack of cross-border inputs from Mexican waters.

The model also reproduces the seasonality in dissolved oxygen in all subregions (Figure 16), characterized by large meridional and vertical variability. Near the Channel Islands, dissolved oxygen varies at 50 m by up to 140 mmol O_2 m^{-3} between the highest winter values and the lowest summer values, reflecting the dynamics of upwelling, productivity, and air–sea gas exchange. Offshore the coasts of Santa Monica and San Diego, the variability between winter and summer is of the order of 80–90 mmol O_2 m^{-3}. Surface concentrations are everywhere above 240 mmol O_2 m^{-3} year-round, consistent with observations. The highest summer concentrations are observed at the depth of the deep chlorophyll maximum, reflecting photosynthesis, while decreasing at depth to below 150 mmol O_2 m^{-3}. These patterns are generally consistent with observations in the same regions.

During the 1998 El Niño event, the model shows a net decrease of dissolved oxygen near the surface, and a net increase below it. During this period, the entire upper layer (0–80 m) is characterized by a homogeneous oxygen concentration of about 240 mmol O_2 m^{-3} over almost the entire SCB (not shown). Only the San Pedro and Santa Monica Bays show higher concentrations, which we attribute to the local anthropogenic nutrient enrichment and subsequent blooms (see Figure 19). This is consistent with observations of the 1998 El Niño event in California coastal waters (Booth et al., 2014; Chavez et al., 2002).

4.5.2. Vertical Gradients and Seasonal Variability of Carbon System Parameters

Together with pH, the saturation state of aragonite (Ω_{Ar}) is often used as a metric to identify the potential impact of ocean acidification on marine calcifiers, because it is a measure of the availability of carbonate ions for calcium carbonate precipitation (Bednarsek et al., 2019). Ω_{Ar} shows similar vertical variability as dissolved oxygen (Alin et al., 2012; Juranek et al., 2009). Similar to oxygen loss, reduction in pH and Ω_{Ar} in the upper layers can be caused by coastal upwelling and local physical processes (Feely et al., 2018); over longer time scales it also

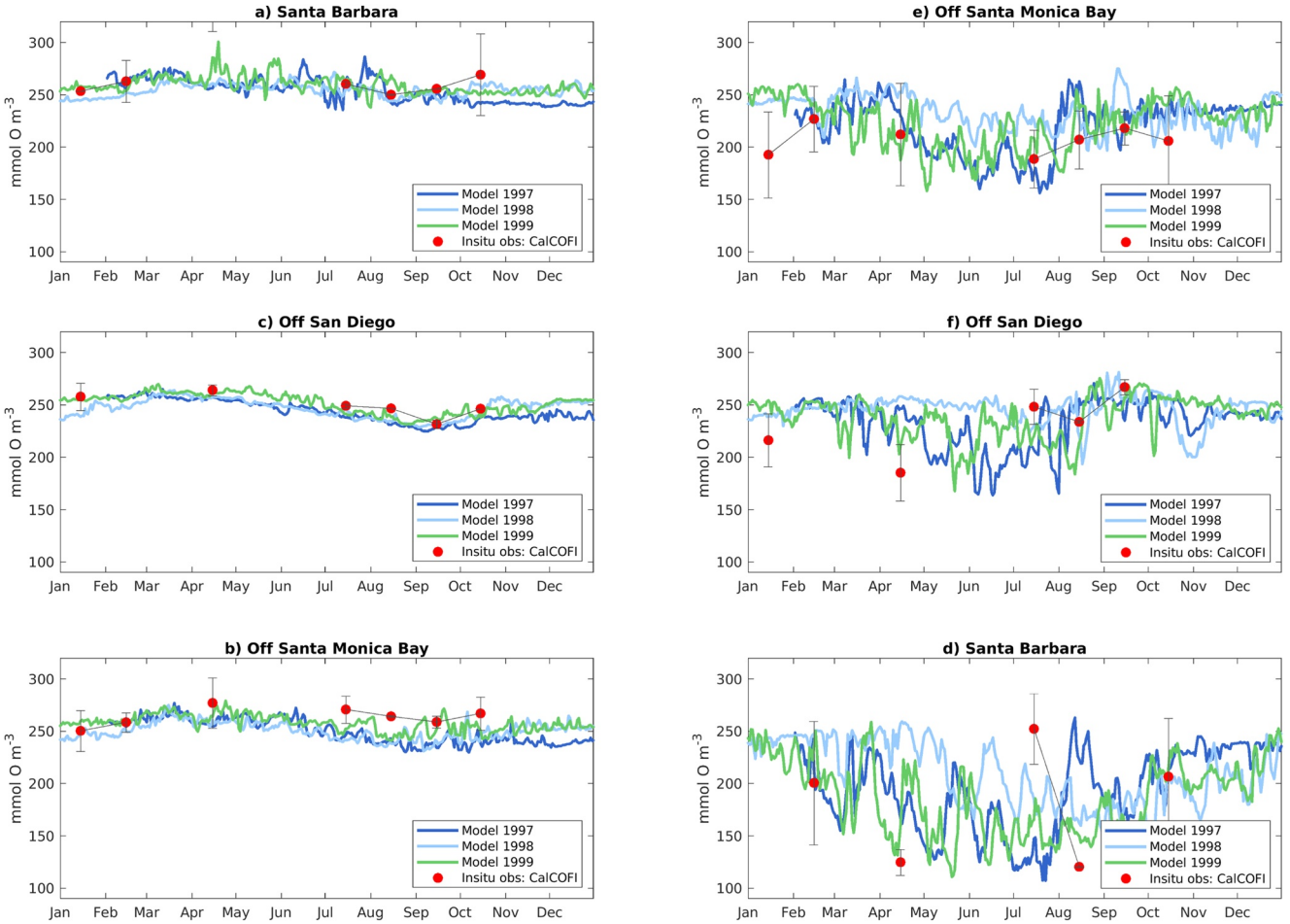

Figure 16. Comparison of dissolved oxygen concentration between different years of model output and a climatology from CalCOFI in situ data. SB is near the center of Santa Barbara Channel, SM is offshore the Santa Monica Bay, and SD is offshore San Diego. Left panels show surface concentrations, right panels concentrations at 50 m depth.

reflects oceanic uptake of anthropogenic carbon from the atmosphere. We utilize sensor pH data sets to evaluate vertical profiles in the carbonate system. Because of the known uncertainty in pH measurements, we are most concerned with how well the model reproduced the shape of the profiles (i.e., goodness-of-fit estimates, as with

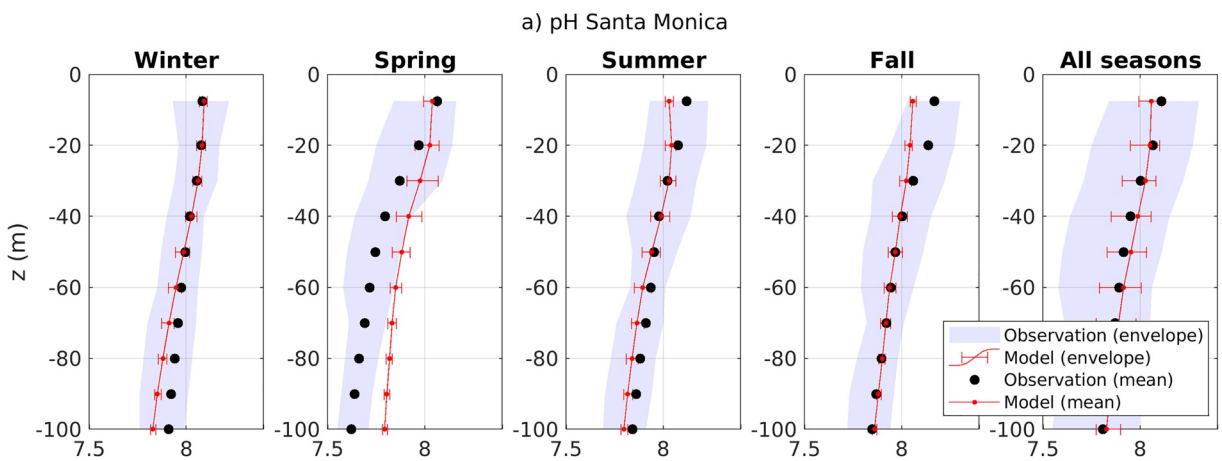

Figure 17. As for Figure 5 but for seawater pH.

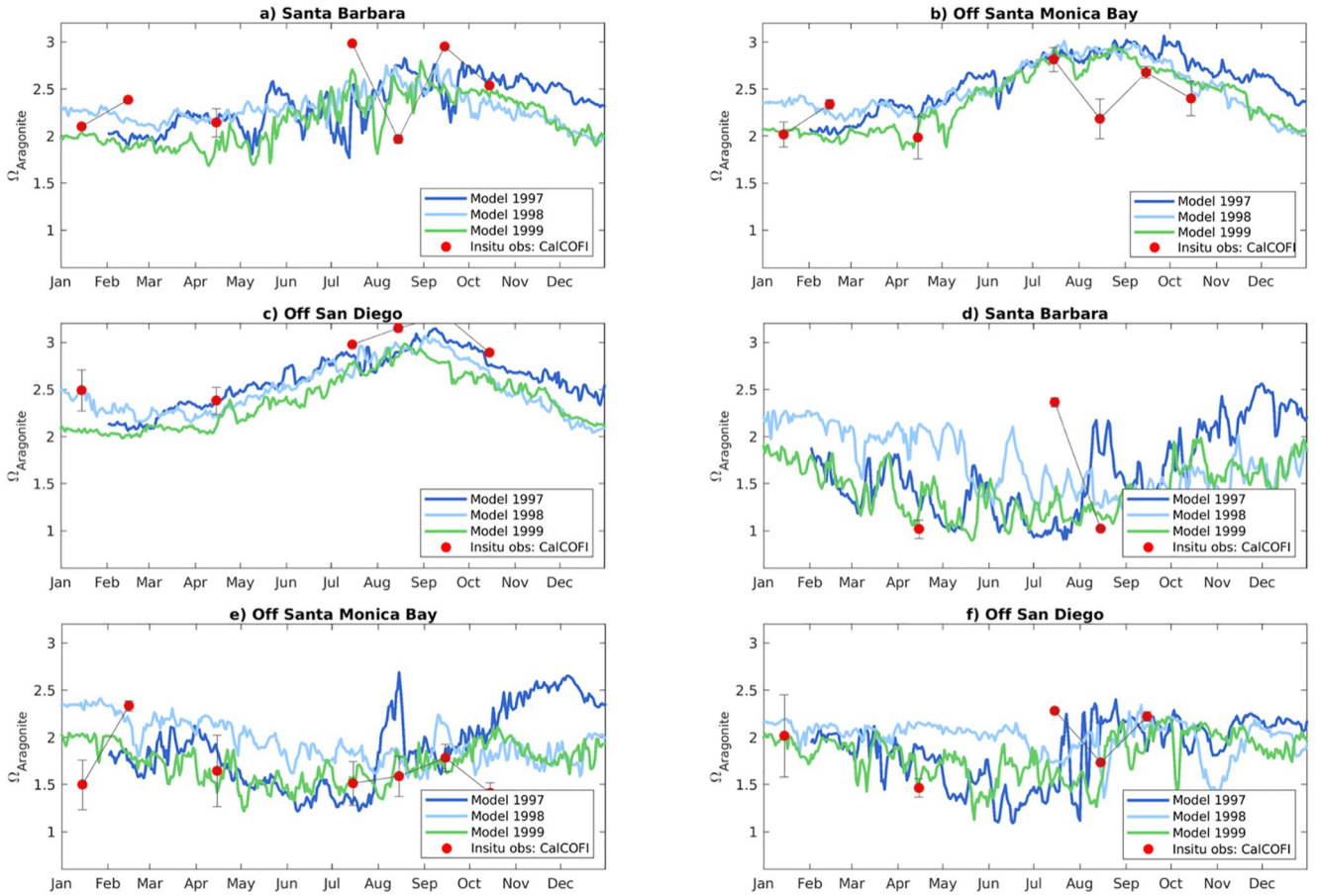

Figure 18. Comparison of the saturation state of aragonite between different years of model output and a climatology from CalCOFI in situ data. SB is near the center of Santa Barbara Channel, SM is offshore the Santa Monica Bay, and SD is offshore San Diego. Left panels show surface values, right panels values at 50 m depth.

Table 6
Statistical Comparison Between In Situ Data and Model Outputs for Dissolved Oxygen and pH Profile in Santa Monica Bay

Santa Monica								
	H	Correlation coefficient	*p*-value	Cost function	Bias	Ratio of standard deviations	Nash–Sutcliffe model efficiency	Number of observations
Oxygen								
Winter	0 E	0.97 E	9E−07 E	0.14 E	−0.09 E	1.20 G	0.77 E	716
Spring	0 E	0.91 E	3E−04 E	0.26 E	−0.23 R	1.03 E	0.37 R	702
Summer	0 E	0.99 E	2E−10 E	0.07 E	0.07 E	0.99 E	0.86 E	712
Fall	0 E	0.97 E	2E−06 E	0.19 E	−0.14 G	1.49 P	0.42 R	718
All seasons	0 E	0.97 E	3E−06 E	0.14 E	−0.11 G	1.18 G	0.69 E	2,848
pH								
Winter	0 E	0.99 E	2E−08 E	0.01 E	0.01 E	0.59 P	0.57 G	632
Spring	0 E	0.97 E	2E−06 E	0.02 E	−0.02 E	1.45 P	0.15 P	702
Summer	0 E	0.96 E	9E−06 E	0.01 E	0.01 E	1.01 E	0.85 E	712
Fall	0 E	0.97 E	3E−06 E	0.01 E	0.01 E	1.49 P	0.78 E	715
All seasons	0 E	0.97 E	5E−06 E	0.01 E	−0.01 E	1.12 G	0.84 E	2,761

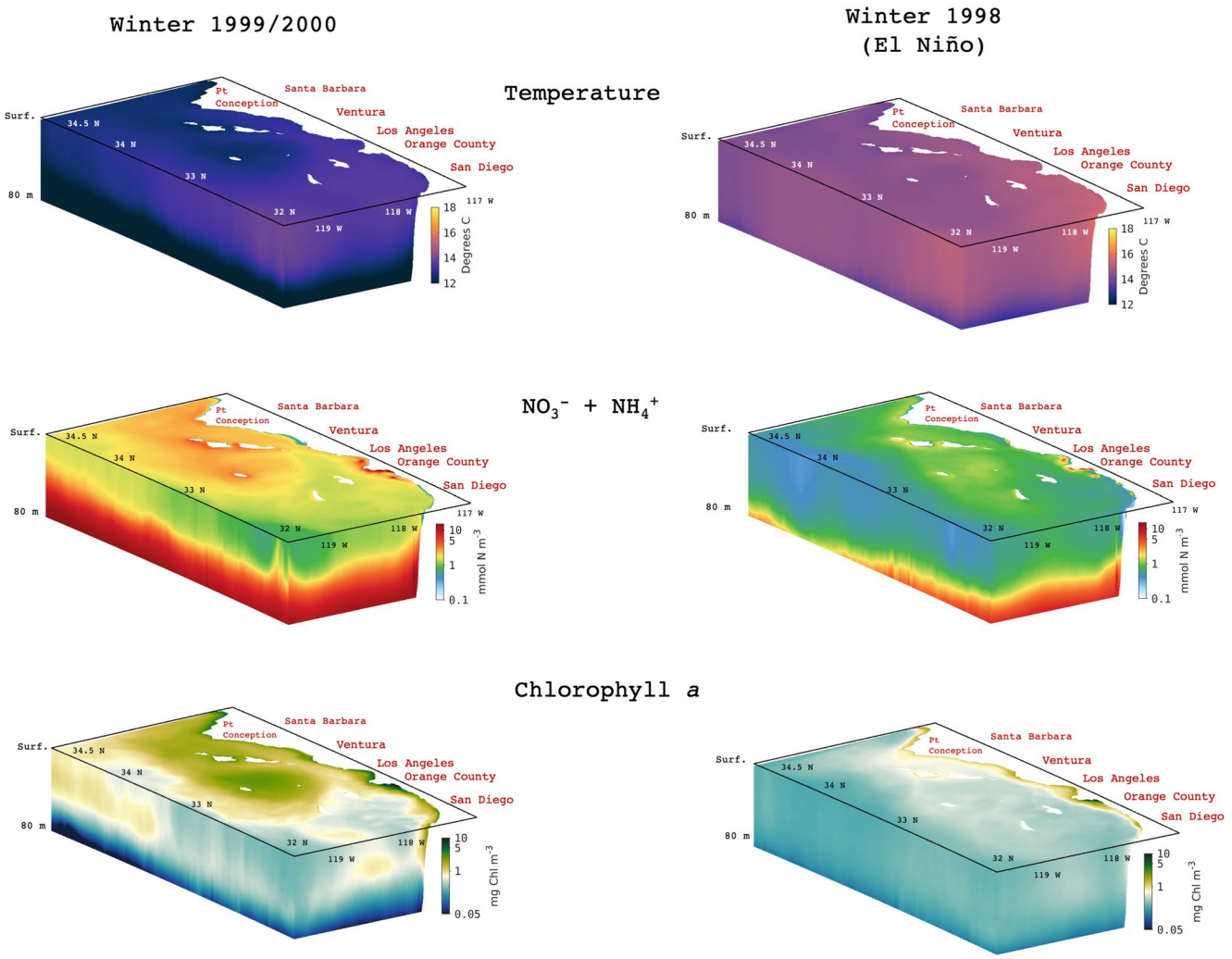

Figure 19. Three-dimensional illustration of temperature, DIN (NO_3^- + NH_4^+) and chlorophyll in the Southern California Bight. Panels show winter 1999 and 2000 (left panels), winter 1998 during El Niño period (right panels).

chlorophyll). Sensor-derived pH profile measurements should be internally consistent within a data set (if the sensor is working properly and if pressure issues are minimal), providing some value to goodness-of-fit assessments. Given these constraints, the data–model comparisons for pH sensor data were generally "*excellent*" or "*good*" for all subregions and all seasons 17. Unsurprisingly, the model performance reproducing observational means and variability was generally "*reasonable*" or "*poor*" for most subregions and seasons, with some, if not most, of this disagreement due to difficulties in conducting a validation of the model with large uncertainties in sensor-derived pH profiles. Recently, the CalCOFI program has incorporated Ω_{Ar} into its sampling design. Although the data do not line up with the model period, they are useful for evaluating seasonal variability in the model. Generally, the model reproduces seasonal and vertical variability in Ω_{Ar}, with higher saturation states in the summer and fall, when waters are generally more stratified, and lower values in winter and spring, when upwelling brings undersaturated waters closer to the surface. Ω_{Ar} is also much lower and more highly variable at depth. These patterns are consistent with observations throughout the SCB (McLaughlin et al., 2018).

5. Discussion and Conclusions

In this study, we demonstrated the readiness of a high-resolution, dynamically downscaled, physical–biogeochemical model to mechanistically investigate links between a comprehensive reconstruction of terrestrial and atmospheric nutrient inputs, coastal eutrophication, and biogeochemical change in the SCB coastal waters. This

modeling platform is an important achievement because it strikes a balance of capturing the forcing of coast-wide basin mesoscale phenomena, while capturing the combined effects of bathymetry and submesoscale eddies that intensify transport of nutrients and biological material. Moreover, this model allows hindcast simulations of primary production, ocean acidification, and oxygen loss at time scales that can approach the multiannual frequencies of intrinsic ocean variability. Future research using this model will make the grand challenge of disentangling natural variability, climate change, and local anthropogenic forcing a tractable task in the near-term. The model evaluation developed in this manuscript discusses sources of uncertainties for the circulation, nitrogen, carbon, and oxygen cycles that serve as a critical element to communicate results to regional stakeholders.

ROMS has a long history of validation and management acceptance through various applications in the CCS (e.g., Capet, Campos, et al., 2008; Capet, Colas, et al., 2008; Capet, McWilliams, et al., 2008; Capet et al., 2004; Marchesiello et al., 2003; Renault, Molemaker, Gula, et al., 2016; Shchepetkin & McWilliams, 2011). In contrast, experience with BEC within the SCB is more limited. Our validation study of coastal eutrophication gradients in the SCB nearshore complements the U.S. West Coast-wide study of Deutsch et al. (2021) and strengthens confidence that the basic CCS BEC model formulation, forcing and parameterization is appropriate not only for coast-wide analyses but also for detailed local studies of coastal eutrophication in the highly urbanized SCB (Kessouri et al., 2021).

In Table 1, we summarize a series of statistical tests of model performance and provide guidelines for their interpretation. These tests are helpful for the validation and interpretation of the model results, because they quantify in an accessible and succinct way information related to magnitude, variability, gradients, and systematic biases of model variables relative to observations. We further apply these tests to the most relevant biogeochemical variables (see Tables 2–4 and 6). In particular, tests that compare variability are essential for a model that resolves submesoscale circulation and the environmental heterogeneity it produces. These tests are used to build confidence in the use of the model for coastal applications, including scenarios and attribution experiments (Kessouri et al., 2021). Likewise, they could be helpful metrics for multimodel ensemble comparisons.

The representation of physical processes such as vertical mixing and horizontal circulation was consistent across the model and measurements. The model reproduces the main structure of the climatological upwelling front and cross-shore isopycnal slopes, and the mean current patterns and associated temperature gradients. We also demonstrate good agreement between model simulations and the mean distributions and variability of key ecosystem metrics, including surface nutrients and productivity, and subsurface O_2 and carbonate saturation. The spatial patterns of primary production, phytoplankton growth rates, and zooplankton grazing are broadly consistent with measured rates. The distribution of primary production is governed by the trade-off between nutrient and light limitation, a balance that reproduces and explains the observed spatial variations in the depth of the deep chlorophyll maximum. Statistical measures of model agreement on biogeochemical state variables were excellent to good and the range of predicted biogeochemical rates on par with observations. Under the realistic flow fields produced by ROMS, the conformity of model predictions with a rich observational data set is a strong demonstration of model validity for coastal eutrophication applications (Kessouri et al., 2021). We also demonstrated that the model responds with confidence to the variability caused by El Niño, modifying the vertical distribution of the physical and biogeochemical properties across the upper ocean of the entire Bight, as illustrated by the three-dimensional change in key ocean variables shown in Figure 19.

While the broad agreement between the model and observations for a range of variables is encouraging, some aspects of the model require further investigation and improvement. Some of the model biases are shared across multiple variables, reflecting common underlying processes, such as stoichiometry of biogeochemical reactions and solubility. We highlighted some of these biases, for example: (a) spring is the season when we see the largest variability, in particular between March and June, when we move from vigorous surface mixing to a strong stratification; (b) along the coasts of Orange County and San Diego, missing nutrients sources from the southern boundary (i.e., from the Mexican coast) likely drive underestimation of phytoplankton concentration, productivity, and carbon export, causing an overestimation of oxygen, pH, and calcium carbonate saturation in the subsurface; (c) overestimation of temperature in the water column impacts oxygen and carbon solubility by decreasing their equilibrium concentrations; this suggests that some biases, such as too high oxygen in spring, hide overcompensation by other biogeochemical processes, for example, high productivity or reduced gas exchange; and (d) underestimation or overestimation of pH and calcium carbonate saturation state is tightly linked by inorganic

carbon chemistry and can in turn reflect biases in circulation, water column structure, temperature, and nutrient cycles of the type discussed above.

The structure of the model is also limited in its representation of ecosystem dynamics. For example, phytoplankton diversity is limited in the model, preventing it from properly simulating events such as dinoflagellate-driven red tides, which occur over short periods on nearshore coastal scales, typically in spring. Despite the good performance of the model in reproducing total primary production and grazing rates, the model does not include multiple zooplankton functional groups, thus providing little information on the dynamics and transfer of energy to higher trophic levels, or the formation of rapidly sinking fecal pellets. From a hydrodynamics point of view, with a horizontal resolution of 300 m, the model does not directly resolve physical processes occurring at sales of tens of meters (Dauhajre et al., 2019), for example the dilution and entrainment of buoyant wastewater plumes, which are parameterized in the model, or the vertical and horizontal transport of tracers in the very nearshore surf zone.

Quantitative and qualitative results of confidence assessments are essential for informing management decisions, evaluating management strategies, and providing a basis for risk analyses. The most successful management approaches are those that explicitly incorporate uncertainty (e.g., Taylor et al., 2000). An assessment of model validation must consider the complex combination of model and observational uncertainties (Allen et al., 2007), including (a) uncertainty/error in the model, with the inclusion of intrinsic variability; (b) uncertainty/error in measured data; (c) uncertainty from the difference in spatial scale of the model output relative to the measured data used in the comparison (specifically, comparing a 0.3-km grid cell to a discrete sampling station); and (d) uncertainty from the difference in temporal averaging of the model output relative to the measured data. For parameters in which we have high confidence in the observational record, that is, temperature and dissolved oxygen, model performance statistics show excellent agreement for mean profiles, vertical and horizontal gradients, as well as seasonal variability. The model reproduces chlorophyll reasonably well, albeit with some biases, which can be in part attributed to a simplified representation of plankton diversity, measurement uncertainty, sparseness of in situ data, cloud cover, and algorithm biases in satellite products. Variables such as pH and ammonium show lower agreement, likely due to measurement uncertainty and sampling bias, but general spatial and temporal patterns are correctly reproduced in the model.

Greater clarity is needed in the requirements for model performance and uncertainty to support decisions on management of SCB coastal water quality and eutrophication (Boesch, 2019). These requirements are likely to be driven largely by the approach that will be used to interpret a "significant impact" (e.g., existing water quality pH and dissolved oxygen criteria, or biologically relevant thresholds; Weisberg et al., 2016), as these have significant implications for required model precision and accuracy on different spatial and temporal scales. Future efforts to constrain uncertainty could include sensitivity analyses and model ensemble comparisons of the BEC biogeochemical component with other biogeochemical models that feature increasingly complex representations of planktonic functional groups, benthic communities, and sediment–pelagic interactions. Finally, long-term investments are needed in coupled chemical–biological observations of phytoplankton and zooplankton diversity and community structure. These observations are critical to provide understanding of the evolution of lower trophic ecosystem structure with climate change, and their relationship with biogeochemical cycles linked to ocean acidification and oxygen loss (Sailley et al., 2013). Ultimately, the need to constrain uncertainty will likely scale with the economic import of management decisions under consideration, which could range from increased monitoring requirements to multibillion dollar nonpoint source controls and wastewater treatment plant upgrades.

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Data Availability Statement

Code is available in Kessouri, McWilliams, et al. (2020). In situ observation data to generate the figures and statistics are available in Kessouri, McLaughlin, et al. (2020). Local land-based and atmospheric data can be found in Sutula et al. (2021a).

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