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RESEARCH ARTICLE

# Estimating Aboveground Carbon Dynamic of China Using Optical and Microwave Remote-Sensing Datasets from 2013 to 2019

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Over the past 2 to 3 decades, Chinese forests are estimated to act as a large carbon sink, yet the magnitude and spatial patterns of this sink differ considerably among studies. Using 3 microwave (L- and X-band vegetation optical depth [VOD]) and 3 optical (normalized difference vegetation index, leaf area index, and tree cover) remote-sensing vegetation products, this study compared the estimated live woody aboveground biomass carbon (AGC) dynamics over China between 2013 and 2019. Our results showed that tree cover has the highest spatial consistency with 3 published AGC maps (mean correlation value  $R = 0.84$ ), followed by L-VOD ( $R = 0.83$ ), which outperform the other VODs. An AGC estimation model was proposed to combine all indices to estimate the annual AGC dynamics in China during 2013 to 2019. The performance of the AGC estimation model was good (root mean square error = 0.05 Pg C and  $R^2 = 0.90$

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with a mean relative uncertainty of 9.8% at pixel scale [0.25°]). Results of the AGC estimation model showed that carbon uptake by the forests in China was about  $+0.17 \text{ Pg C year}^{-1}$  from 2013 to 2019. At the regional level, provinces in southwest China including Guizhou ( $+22.35 \text{ Tg C year}^{-1}$ ), Sichuan ( $+14.49 \text{ Tg C year}^{-1}$ ), and Hunan ( $+11.42 \text{ Tg C year}^{-1}$ ) provinces had the highest carbon sink rates during 2013 to 2019. Most of the carbon-sink regions have been afforested recently, implying that afforestation and ecological engineering projects have been effective means for carbon sequestration in these regions.

## Introduction

China's forested areas represent a carbon sink that contributes substantially to regional and global carbon storage [1,2]. However, both the magnitude and spatial patterns of the carbon changes in China differ across published reports. A precise quantification of the annual aboveground biomass carbon (AGC) dynamics is therefore important for understanding how China's forests contribute to the global carbon balance [3,4]. Satellite records offer the potential to monitor AGC changes by combining diverse remote-sensing sources including optical, synthetic aperture radar (SAR), light detection and ranging (LiDAR), and microwave sensors [5–10]. Yet, current assessments of the AGC dynamics in China using satellite records remain highly uncertain [11,12].

Optical vegetation indices (VIs) are related to the greenness components of the vegetation and thus indirectly to AGC [13]. VIs could provide better spatial coverage with high resolution and are available for long-term time series, which can provide effective global monitoring of vegetation trends [14]. Previous studies have used optical VIs for AGC estimation in different regions, e.g., tropical regions [5,7,15], in the USA [16,17], and in China [9,18–20]. Xu et al. [21] developed globally estimates of carbon stock changes of live woody biomass using measurements from ground, air, and space. These AGC maps were produced by integrating remote-sensing data (e.g., optical, SAR, and LiDAR) and field measurements using the machine learning methods [22,23]. They constituted the best benchmarks to date for carbon stored in living woody vegetation and have been used as benchmark maps for monitoring the AGC dynamics over the global and tropical regions [6,10,24,25]. However, optical VI signals are known to be affected by clouds and aerosols and suffer from the saturation problem in densely vegetated areas [26,27]. Moreover, the retrieved AGC values using these approaches differ in both magnitude and spatial patterns among the different studies and are available for only a single epoch, and therefore cannot be used to monitor the temporal dynamics of carbon stocks [24,28]. Although LiDAR is considered as the most promising method for estimating AGC, there was no dedicated spaceborne LiDAR designed specifically to estimate vegetation AGC until the launch of Global Ecosystem Dynamics Investigation (GEDI) in 2019 [29].

Vegetation optical depth (VOD) retrieved from microwave sensors measures the vegetation water content, thus providing a useful proxy for monitoring AGC dynamics [6,10,30–32]. Compared to optical VIs, VOD products are sensitive to both leafy and woody vegetation components and are mostly insensitive to atmospheric and cloud effects [33,34]. Changes in AGC stocks have been inferred using VOD products from multiple microwave sensors, including VOD products at high frequency (C-, X-, and K-bands, 6.9 to 18.7 GHz) [32,35–40], and at low frequency (L-band, 1.4 GHz) [41,42]. Low-frequency VOD has been shown to have better capabilities to monitor the AGC dynamics due to the higher penetration capacity of the

microwave observations at large wavelengths ( $\sim 20 \text{ cm}$  at L-band) [6,10,24,25,43]. Therefore, L-band VOD (L-VOD) is less sensitive to saturation effects and more sensitive to AGC held in stems and branches than high-frequency VOD products [44,45]. Yet, radio frequency interference (RFI) at L-band degrades locally the performance of L-VOD for monitoring of AGC [46,47], especially over some regions in China. High-frequency VOD products are less affected by RFI but can suffer from saturation in dense forests [48].

Given the respective weaknesses and advantages of VODs and VIs, combining the 2 types of data could be a potential way to improve the quantification of the AGC dynamics [45,49–51]. Assessing the sensitivity of VODs and VIs to AGC is, however, an essential preliminary step in any attempt to monitor AGC dynamics from remote-sensing. The performance of VODs for AGC monitoring has been evaluated at global [48,52,53] or regional scales, for example, in tropical regions [54–56], Australia [57,58], and the USA [59]. However, there is a lack of studies in China, especially considering the strong influence of RFI on L-VOD over China [47].

Several empirical functions have been proposed to spatially regress AGC against VOD at large scales, e.g., linear regressions [24], arctangent regressions [6,32,43], and logistic regressions [48,55]. However, the relationships between AGC and VODs/VIs are quite complex and may not be well captured by these regression functions [48,54]. As an effective tool to fit complex data with nonlinear partial relations, machine learning approaches have been increasingly used for forest structure and AGC mapping [21,60]. Among all machine learning approaches, random forest (RF) models [61] show excellent performance in modeling the complex relationships between different land surface variables and AGC [62]. An advantage of RF is that multiple predictors can be incorporated without making further assumptions about their covariance structure, which makes RF models suitable for describing nonlinear relationships between AGC and VODs/VIs.

The objectives of this study are as follows: (a) to assess the performances of 3 VODs (e.g., L-VOD, IB-VOD, and LPDR-VOD) and 3 optical VIs (e.g., normalized difference vegetation index [NDVI], leaf area index [LAI], and tree cover [TC]) in predicting AGC across China using 3 AGC benchmark maps; (b) to produce AGC products from the combination of VODs and VIs using a RF model; and (c) to quantify the interannual dynamics of AGC stocks across China for the period 2013 to 2019.

## Materials

Forests, which cover more than 20% of mainland China [3], are mainly located in the northeastern, southern, and southwestern regions. Over the past 2 decades, ecological restoration projects were implemented in several regions (Fig. 1), including: Three-North Shelter Forest Program, Tibetan Plateau, Loess Plateau, and Southern China [63].

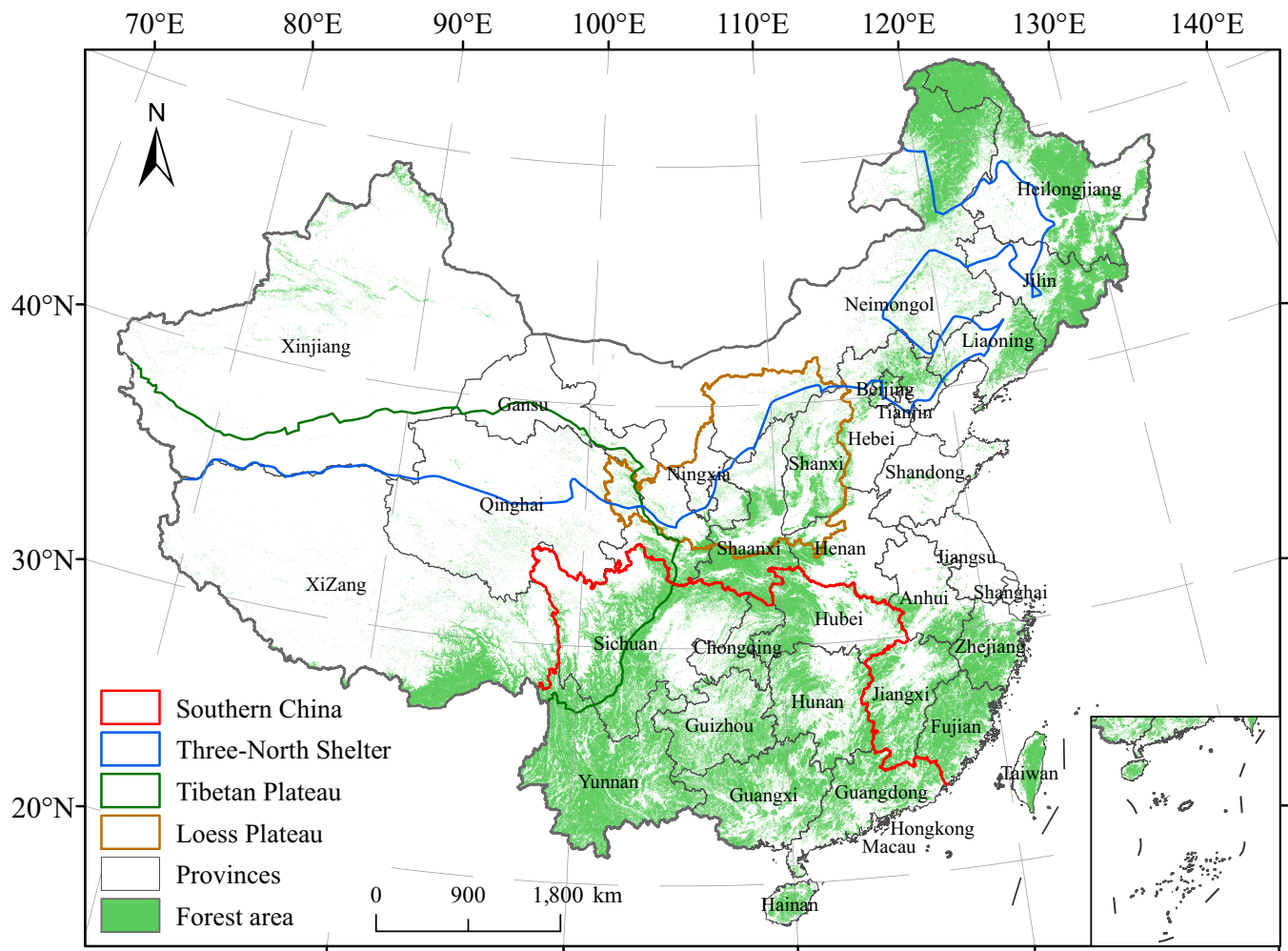


Fig. 1. Location of the provinces and 4 main ecological restoration regions (the Three-North Shelter, Tibetan Plateau, Loess Plateau, and Southern China) in China.

## Microwave VOD products

### L-VOD

L-VOD was retrieved from the Soil Moisture and Ocean Salinity (SMOS) satellite using the SMOS-IC algorithm (version 2) [41,42]. Version 2 of the L-VOD product provides daily global L-VOD and soil moisture (SM) retrieved from the ascending and descending data at 0.25° resolution for 2010 to 2020. The L-VOD product was retrieved by the L-band (1.4 GHz) microwave emission of the biosphere model, without any auxiliary hydrology or vegetation products as inputs [64]. In this study, the observations affected by RFI for which root mean square error (RMSE) associated with the L-VOD product is larger than 8 K were filtered out. In addition, observations affected by strong topography, frozen conditions, and the sum of urban, water and ice fractions higher than 10% were excluded [41]. The yearly L-VOD values for a given pixel were calculated as the medians of all available ascending and descending retrievals in the year. The SMOS-IC L-VOD product was selected here as it has been shown to exhibit a stronger relationship with AGC than other SMOS products [55].

### IB-VOD

The IB-VOD product was retrieved by the X-band microwave emission of the biosphere model [39]. IB-VOD was retrieved from

the Advanced Microwave Scanning Radiometer 2 (AMSR2) X-band (10.7 GHz, with a spatial resolution of 0.25°) descending data (local time = 1:30 AM) considering proximal conditions of the air, canopy, and soil at night. To avoid the ill-posed issue caused by retrieving both SM and VOD from mono-angular and dual-polarized observations (AMSR2), the X-band microwave emission of the biosphere model uses the ERA5 SM product [65] as input.

### LPDR-VOD

LPDR-VOD was derived from the land parameter data record (LPDR, version 2) [35]. LPDR provides a long-term (2002 to 2020) global record of the daily X-VOD products at 0.25° resolution. It is retrieved from the AMSR for Earth Observing System (AMSR-E) and AMSR2 sensors at 10.7 GHz based on both ascending (13.30) and descending (01.30) data. Grid cells affected by frost conditions, large water bodies, strong precipitation, or RFI were excluded using the quality assessment files [66].

### Optical VIs

We applied 3 optical VIs in this study that are commonly used to estimate AGC in previous studies (Table 1) [22,67], including:

1. NDVI, provided by the monthly MODIS product (MOD13A3) at 1-km spatial resolution [68]. NDVI is



**Table 1.** Overview of the VODs, VIs, and benchmark AGC datasets used in this study.

| Dataset    | Products/Sensor                                                  | Period    | Resolution | Temporal | Reference                |
|------------|------------------------------------------------------------------|-----------|------------|----------|--------------------------|
| L-VOD      | SMOS-IC V2                                                       | 2013–2019 | 0.25°      | Daily    | Fan et al. [6]           |
| IB-VOD     | AMSR2                                                            | 2013–2019 | 0.25°      | Daily    | Wang et al. [12]         |
| LPDR-VOD   | AMSE-E and AMSE2                                                 | 2013–2019 | 0.25°      | Daily    | Du et al. [35]           |
| NDVI       | MOD13A3                                                          | 2013–2019 | 1 km       | Monthly  | Didan. [68]              |
| LAI        | MOD15A2                                                          | 2013–2019 | 500 m      | 8 d      | Myneni et al. [71]       |
| TC         | MOD44B                                                           | 2013–2019 | 250 m      | Yearly   | Dimiceli et al. [73]     |
| Saatchi    | GLAS LiDAR, QSCAT radar, field data, NDVI, LAI, and SRTM         | 2015      | 1 km       | Yearly   | Saatchi et al. [7]       |
| CCI        | Sentinel-1, ALOS-2 PALSAR-2, field data, canopy height, and SRTM | 2017      | 100 m      | Yearly   | Santoro and Cartus. [76] |
| Saatchi-WT | Data fusion                                                      | 2010s     | 1 km       | Yearly   | Supplementary Text       |

widely used to monitor vegetation cover [69,70] and biomass [13,16].

2. LAI, derived from the 8-day MODIS product (MOD15A2) at 500-m spatial resolution [71]. LAI is an important indicator of plant growth and radiation use efficiency [72].
3. TC, derived from the yearly MODIS product (MOD44B) at 250-m spatial resolution [73], represents the percent tree canopy ( $\geq 5$  m) cover [74].

In this study, the pixels with snow or cloud were filtered out using the quality assessment data accompanying the product. Usually, only NDVI and EVI are referred to as vegetation indices. But for convenience, NDVI, EVI, LAI and TC products were referred to as vegetation indices as in previous studies [52]. All VI products were aggregated to 0.25° through simple averaging to match the spatial resolution of VODs.

### AGC benchmark maps

This study used 3 AGC benchmark maps to assess the performances of VIs and VODs to AGC (Table 1), including:

1. Saatchi map, estimating AGC in the pan-tropics at 1-km resolution by merging data from field measurements, GLAS LiDAR, MODIS, and Quick Scatterometer [7]. The Saatchi map used here is an updated version that represents global AGC for 2015 by combining data from the SRTM, Landsat, and Advanced Land Observing Satellite (ALOS) [75].
2. Climate Change Initiative (CCI) map, a global map of AGC at 100-m resolution for 2017 using multiple remote-sensing observations [76]. The CCI map was produced by combining data from the SAR C-band Sentinel-1 and L-band ALOS-2 Phased Array L-band SAR (PALSAR-2).
3. Saatchi-WT map, produced by merging the forest AGC map and nonforest AGC map [77]. The forest and nonforest regions were identified using the Liu et al. [78] land cover map [78]. The forest AGC map was estimated using a weighting technique (WT) method to merge the 5 published AGC products [5,7–9,19]. The WT method was trained using independent field observations across

China [79] (Supplementary Text). The nonforest AGC map was computed from the Saatchi map over nonforest regions.

The spatial resolution of all AGC maps was aggregated to 0.25° by simple averaging. The units of biomass density (Mg/ha) were converted to carbon density (Mg C/ha) by multiplying the original values by 0.5 [5].

### Methods

In this study, we focused on aboveground carbon stocks for live woody vegetation and masked the nonvegetated pixels dominated by the following land-cover classes: “settlement”, “water”, “bare and sparse vegetation”, and “wetland” using the CCI land cover dataset [80].

Considering the availability of the products, only L-VOD, LPDR-VOD, NDVI, LAI, and TC were used for estimating AGC. The VODs and VIs in the same year 2015 were chosen as the baseline for model calibration. We then estimated the yearly AGC dynamics across China for the period 2013 to 2019.

### Comparison of VIs and VODs to benchmark AGC maps

The relationship of VIs and VODs to AGC was indicated by the Pearson correlation coefficient  $R$  [81] (Eq. 1) by comparing the AGC datasets with VIs and VODs, as defined below:

$$R = \frac{\sum (VOD - \overline{VOD})(AGC - \overline{AGC})}{\sqrt{\sum (VOD - \overline{VOD})^2 \sum (AGC - \overline{AGC})^2}} \quad (1)$$

where  $VOD$  and  $AGC$  are the values of VOD and benchmark AGC, respectively.  $\overline{VOD}$  and  $\overline{AGC}$  are the average values of VOD and benchmark AGC, respectively. We used  $P$  value ( $P < 0.05$ ) to define the significance level of the correlation.

### Estimation of AGC using RF

The yearly AGC stocks were estimated by combining VODs and VIs with RF models [61]. Firstly, an AGC model hereafter  $AGC_{main}$  given by Eq. 2 was trained using all 5 predictors in 2015

(L-VOD ( $VOD_L$ ), LPDR-VOD ( $VOD_{LPDR}$ ), NDVI, LAI, and TC), and the benchmark AGC maps were the response variables. The  $AGC_{main}$  model was expressed as the linking model ( $f$ ) between benchmark AGC and the predictors in the framework of RF:

$$AGC_{main} = f(VOD_L, VOD_{LPDR}, NDVI, LAI, TC) \quad (2)$$

where  $AGC_{main}$  is the retrieved AGC value.

As there were missing values in the  $AGC_{main}$  estimates caused by the missing observations of L-VOD affected by RFI [41,47], an alternative AGC model ( $AGC_{noLVOD}$ , Eq. 3), as a supplementary model to  $AGC_{main}$ , was trained using the same predictors but without L-VOD. The supplementary AGC model (hereafter  $AGC_{noLVOD}$ ) was defined as:

$$AGC_{noLVOD} = f(VOD_{LPDR}, NDVI, LAI, TC) \quad (3)$$

$AGC_{noLVOD}$  was used to predict AGC estimates for pixels with missing estimates from  $AGC_{main}$ . In a final step a model ( $AGC_{merge}$ , Eq. 4) was defined to merge  $AGC_{main}$  and  $AGC_{noLVOD}$  to estimate the annual AGC density from 2013 to 2019.

$$AGC_{merge} = \text{merging}(AGC_{main}, AGC_{noLVOD}) \quad (4)$$

The  $AGC_{merge}$  model was then applied to the data for each year to derive maps of annual AGC density for 2013 to 2019.

To account for the uncertainties in the benchmark AGC maps, we used all 3 AGC maps to train the AGC model. Thus, 3 calibrated AGC models were obtained. Following Fan et al. [6], assuming that a good calibration result could be obtained, the AGC model adds a temporal dimension to static maps and provides a "space for time" replacement. The mean of these 3 AGC values was calculated to estimate the yearly AGC maps during 2013 to 2019 across China. The range (i.e., minima and maxima) of the 3 AGC values were also calculated, as they provide the uncertainty for the retrieved AGC estimates.

### Validation of AGC estimates

We used a bootstrap and cross-validation method to evaluate the performance of  $AGC_{main}$ ,  $AGC_{noLVOD}$ , and  $AGC_{merge}$ . Three sets of bootstrapped AGC estimates were computed on the basis of 3 calibrations of the AGC model. A bootstrap sampling method (sampling rate = 80%, iterations = 1000) was used to calibrate each AGC model (e.g.,  $AGC_{main}$ ,  $AGC_{noLVOD}$ , and  $AGC_{merge}$ ); the remaining (20%) was used for the validation of the AGC estimates. The uncertainties associated with the AGC estimates were evaluated using 2 metrics, including the adjusted coefficient of determination ( $R^2$ ) and RMSE between the benchmark AGC maps and the RF-derived AGC estimates. Also, the trends and the changes of  $AGC_{main}$  were compared with  $AGC_{noLVOD}$  for their overlap pixels, to confirm that  $AGC_{noLVOD}$  provide similar results to  $AGC_{main}$ .

The uncertainties in the  $AGC_{merge}$  estimates at pixel scale ( $Uncertainties_{pixel}(\%)$ ) were defined as the RMSE of the  $AGC_{merge}$  model ( $RMSE_{AGC_{merge}}$ ) relative to the AGC density from the corresponding benchmark maps ( $AGC \cdot density_{benchmark \cdot map}$ ), expressed as Eq. 5.  $RMSE_{AGC_{merge}}$  at the pixel scale was estimated using the bootstrap approach mentioned above.

$$Uncertainties_{pixel}(\%) = \frac{RMSE_{AGC_{merge}}}{AGC \cdot density_{benchmark \cdot map}} \quad (5)$$

## Results

### Comparison of VIs and VODs to the AGC benchmark maps

All VOD products showed similar spatial patterns with VIs and benchmark AGC maps (Fig. 2A to C): high VODs/VIs/AGC values can be observed in the forests of southern and northeastern China, and low VODs/VIs/AGC values can be observed in the temperate steppe and Qinghai-Tibet plateau alpine. However, large differences can be observed among the 3 benchmark AGC estimates (Fig. 2G to I). In consideration of the total carbon stock in the 3 benchmark AGC maps (Table S1), the values range from 10.23 Pg C (CCI), 12.57 Pg C (Saatchi-WT), to 14.61 Pg C (Saatchi). These differences in AGC stocks can be partly explained by the fact that these benchmark maps have different coverage, which were limited by the different AGC retrieval methods [67]. Another potential reason is that all AGC benchmark maps contain uncertainties and biases [18]. Thus, we used all the different maps to construct the RF model estimating the AGC dynamics over China in this study.

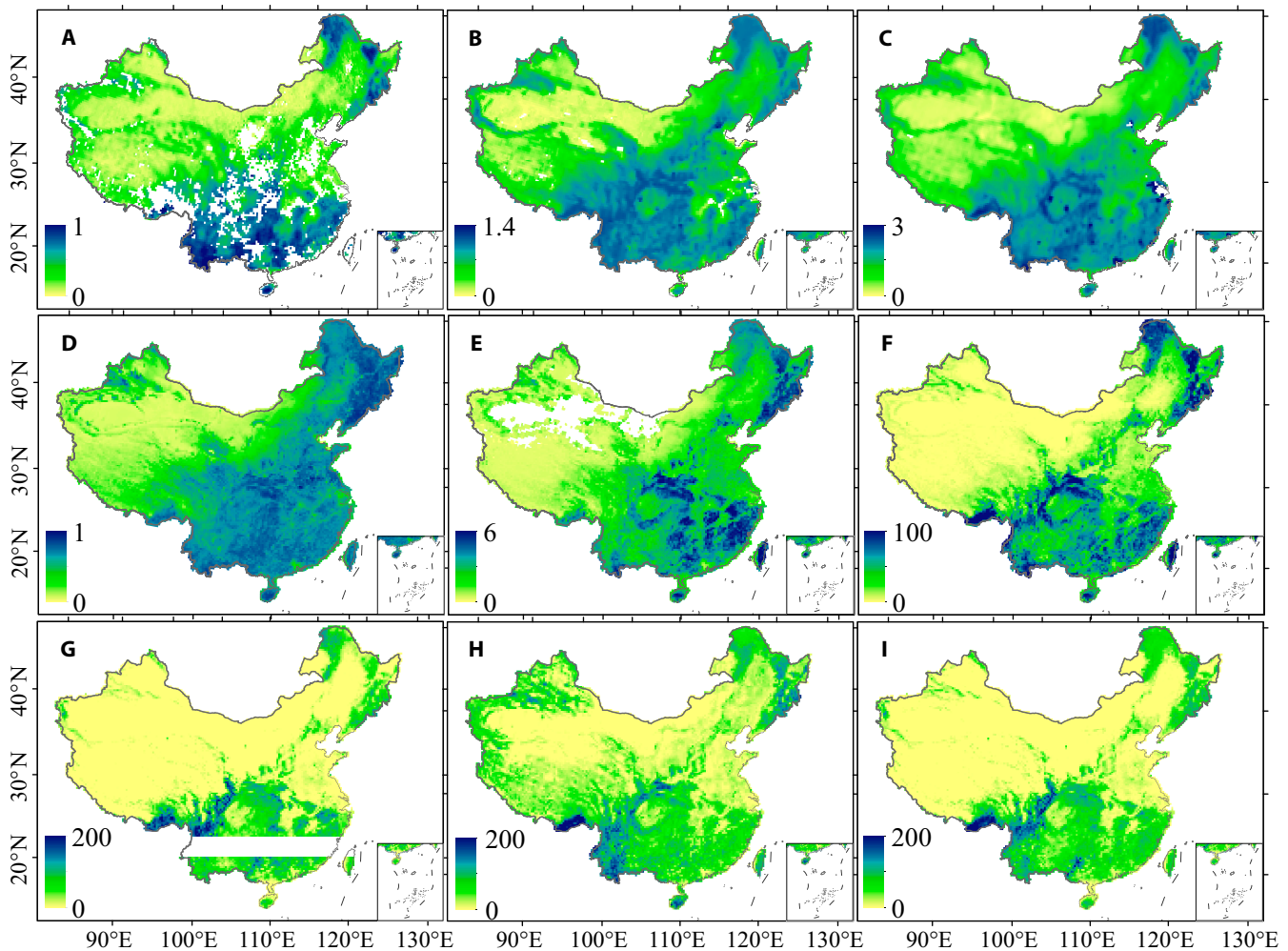
Before carrying out the retrievals of AGC, we firstly compared the VIs/VODs indices against the 3 benchmark AGC maps and found nonlinear relationships (Fig. 3 and Figs. S1 to S2). The highest correlation between the VIs and the benchmark AGC products was obtained for the TC product ( $R = 0.80$  to  $0.88$ ), followed by NDVI ( $R = 0.65$  to  $0.75$ ), while the lowest correlations were found for LAI ( $R = 0.60$  to  $0.68$ ) (Table 2). As an example, it can be seen that NDVI and LAI quickly saturate for Saatchi AGC values above 25 Mg C/ha (Fig. 3D and E). Similar results can be found for the comparison between VIs and other benchmark maps, e.g., CCI (Fig. S1) and Saatchi-WT maps (Fig. S2).

With respect to VODs, L-VOD showed the highest correlation with benchmark AGC maps ( $R = 0.80$  to  $0.85$ ), followed by LPDR-VOD ( $R = 0.72$  to  $0.80$ ) and IB-VOD ( $R = 0.67$  to  $0.74$ ) (Table 2). As an example, the scatterplots between L-VOD and Saatchi AGC showed noticeable variation when  $AGC > 75$  Mg C/ha (Fig. 3A). For IB-VOD and LPDR-VOD, both of them tended to saturate with AGC values greater than 50 Mg C/ha (Fig. 3B and C). Similar results can be seen for the comparison of VODs against the CCI (Fig. S1) and Saatchi-WT (Fig. S2) maps. In summary, IB-VOD and LPDR-VOD all display saturated signals when AGC values are greater than 50 Mg C/ha, whereas L-VOD shows a higher spatial correlation to AGC without strong signs of saturation.

The intercomparisons between VIs/VODs and benchmark AGCs were made to evaluate their expected performance for monitoring AGC, using the mean values of  $R$  as an indicator (Table 2). In summary, we found the highest correlation values were obtained by TC (with a mean value of  $R$  of 0.84), followed by L-VOD ( $R = 0.83$ ), LPDR-VOD ( $R = 0.76$ ), IB-VOD ( $R = 0.71$ ), NDVI ( $R = 0.70$ ), and LAI ( $R = 0.65$ ).

### Calibration and evaluation of the AGC estimation models

The  $AGC_{main}$  model was produced by merging optical and microwave indices with a RF algorithm. However, the missing estimates of the  $AGC_{main}$  model are found over southern and eastern China, limited by the availability of L-VOD data (Fig. 4A). As a supplementary model,  $AGC_{noLVOD}$  was used to provide AGC estimates for pixels with missing estimates from the  $AGC_{main}$  model. The  $AGC_{noLVOD}$  estimates almost covered the whole study region;



**Fig. 2.** Yearly VODs, Vis, and benchmark AGC maps. (A) L-VOD for 2015, (B) IB-VOD for 2015, (C) LPDR-VOD for 2015, (D) NDVI for 2015, (E) LAI for 2015, (F) TC for 2015, and benchmark AGC of (G) Saatchi, (H) CCI, and (I) Saatchi-WT.

thus, combining  $AGC_{main}$  and  $AGC_{noLVOD}$  provided a higher spatial coverage (81%) of China, relative to that obtained from the  $AGC_{main}$  estimates (63%) (Fig. 4B and C). The percentage increase of mean-squared error (%IncMSE) was calculated to evaluate the importance of each variable in RF model. As can be seen from Fig. S4, TC and L-VOD products were found the most important predictors in the RF model.

Both the  $AGC_{main}$  and  $AGC_{noLVOD}$  models nicely fit the spatial pattern of benchmark AGC, as indicated by the high  $R^2$  (0.90, 0.89 for  $AGC_{main}$  and  $AGC_{noLVOD}$  models) and low RMSE values (8.6 Mg C/ha and 9.3 Mg C/ha for  $AGC_{main}$  and  $AGC_{noLVOD}$  models) (Fig. 5). Taking the Saatchi map as an example, both  $AGC_{main}$  and  $AGC_{noLVOD}$  have overcome the saturations issues observed in relationships derived from single index:  $AGC_{main}$  and  $AGC_{noLVOD}$  can fit the spatial changes for AGC stocks larger than 100 Mg C/ha. A similar ability of  $AGC_{main}$  and  $AGC_{noLVOD}$  can be observed for low carbon density, while single index generally showed a high dispersion (e.g., L-VOD, Fig. 3A).

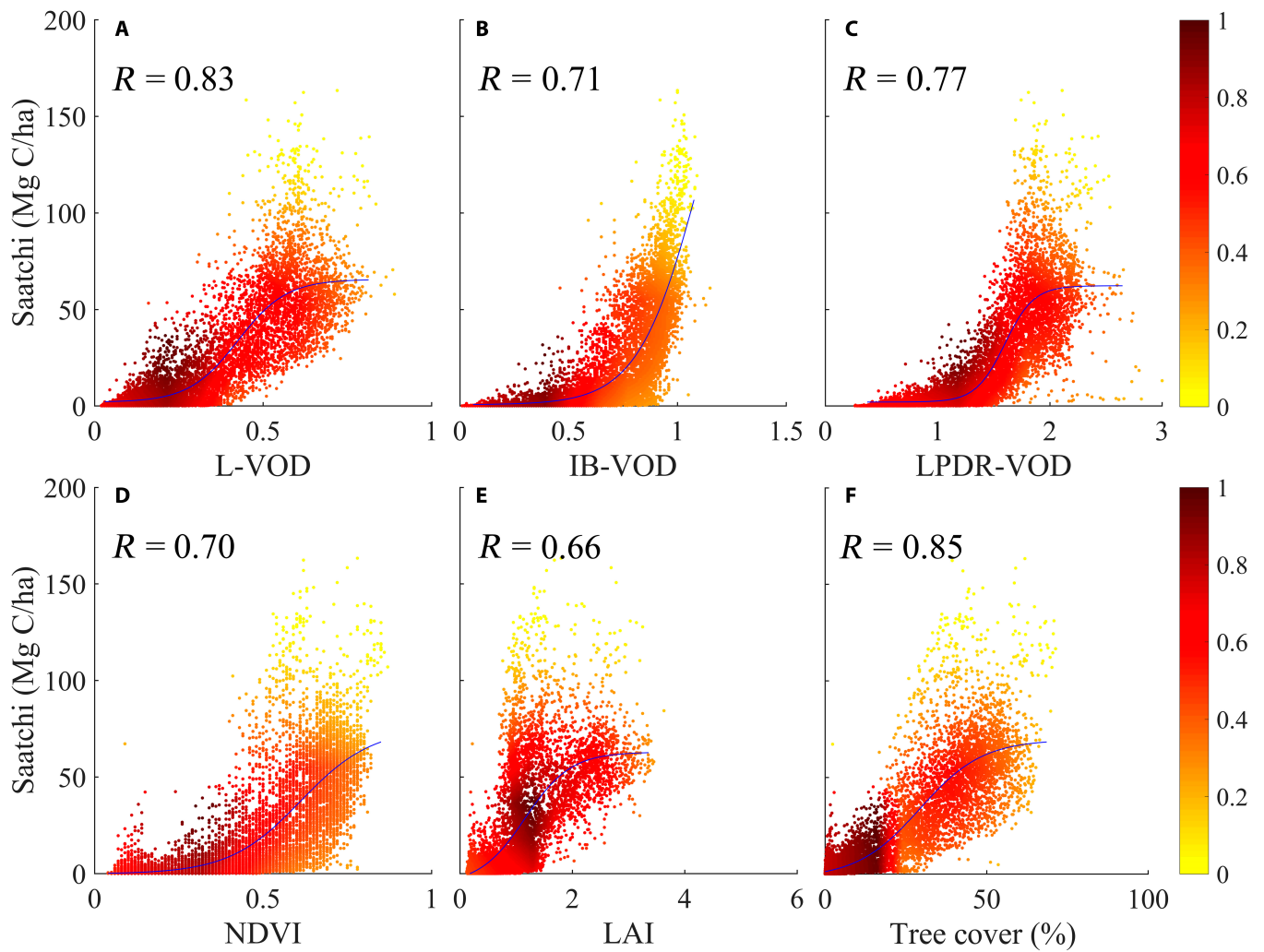
In the regions of overlap between the  $AGC_{main}$  and  $AGC_{noLVOD}$  estimates, the general trends and the magnitude of the changes in  $AGC_{main}$  are very similar to those in  $AGC_{noLVOD}$  (Fig. 6). For example, both the  $AGC_{main}$  and  $AGC_{noLVOD}$  estimates calibrated with the Saatchi map (Fig. 6A) showed generally a net increase in AGC stocks during 2013 to 2017 and an AGC decrease in

2017, in line with the AGC dynamics from the estimates calibrated by the other benchmark maps (Fig. 6B and C).

In terms of magnitude of AGC changes, the net AGC changes estimated from  $AGC_{noLVOD}$  are 12% to 16% higher than those estimated from  $AGC_{main}$  over the study period (Fig. 6). As the  $AGC_{noLVOD}$  model was only used over 26% of the study region, this difference will contribute to only a difference of 3% to 5% of the net AGC changes over China during 2013 to 2019. This suggests that the slight differences in the AGC estimates from the  $AGC_{noLVOD}$  and  $AGC_{main}$  models can be neglected when we consider the whole study period/area. In the following, we will focus on AGC estimates obtained from the model ( $AGC_{merge}$ ) merging the  $AGC_{noLVOD}$  and  $AGC_{main}$  models. The general trends and the magnitude of the  $AGC_{merge}$  dynamics using 3 sets of calibrated parameters are very similar for the 3 estimates (Fig. 7). The general trends of the changes from each variable are very similar to those from AGC (Fig. S5).

Furthermore, the cross-validation using a bootstrap method showed that the AGC stocks estimated by  $AGC_{merge}$  combining  $AGC_{main}$  and  $AGC_{noLVOD}$  have a high accuracy, as indicated by a RMSE of 0.05 Pg C. The good performance of  $AGC_{merge}$  is also indicated by the high  $R^2$  of 0.90 and low relative uncertainties (9.8%) in the pixel-scale AGC density estimates. Moreover, a better performance of  $AGC_{merge}$  can be observed at AGC density





**Fig. 3.** The spatial relationship ( $R$ ) between Saatchi AGC and VIs/VODs from (A) L-VOD, (B) IB-VOD, (C) LPDR-VOD, (D) NDVI, (E) LAI, and (F) TC. Solid blue lines are the fits by a logistic function following Rodríguez-Fernández et al. [55] with the parameters given in Table S1.

larger than 40 Mg C/ha (with relative uncertainties lower than 10%), relative to low AGC density (lower than 40 Mg C/ha) where the relative uncertainties are larger than 10% (Fig. 8B). The latter mainly correspond to the areas distributed across northern China (Fig. 8A). These results showed that the  $AGC_{merge}$  model has a satisfactory accuracy for estimating AGC stocks over China.

### AGC dynamics over the study period

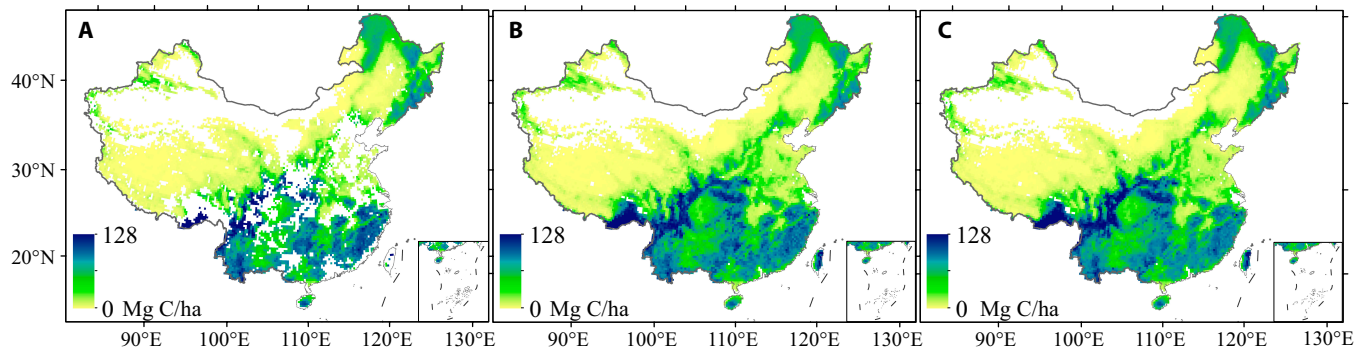
The yearly AGC across China estimated by the  $AGC_{merge}$  model showed a net AGC sink of  $+0.17 \text{ Pg C year}^{-1}$  during 2013 to 2019 (the value is a mean estimated from 3 calibrations; a positive value indicates a net sink of carbon in aboveground biomes; Fig. 9A). Rapid AGC increases can be observed in 2014 to 2016 ( $+0.26 \text{ Pg C year}^{-1}$ ) and 2018 to 2019 ( $+0.15 \text{ Pg C year}^{-1}$ ),

**Table 2.** Spatial correlation ( $R$ ) between VIs/VODs and benchmark AGCs using Eq. 1.

| Products   | L-VOD | IB-VOD | LPDR-VOD | NDVI | LAI  | TC   | Mean |
|------------|-------|--------|----------|------|------|------|------|
| Saatchi    | 0.83  | 0.71   | 0.77     | 0.70 | 0.66 | 0.85 | 0.75 |
| CCI        | 0.80  | 0.67   | 0.72     | 0.65 | 0.60 | 0.80 | 0.71 |
| Saatchi-WT | 0.85  | 0.74   | 0.80     | 0.75 | 0.68 | 0.88 | 0.78 |
| Mean       | 0.83  | 0.71   | 0.76     | 0.70 | 0.65 | 0.84 | --   |

Note: All the correlations are significant considering the criteria  $P < 0.05$ .





**Fig. 4.** The AGC estimates from (A)  $AGC_{main}$ , (B)  $AGC_{noLVOD}$ , and (C)  $AGC_{merge}$  models in 2015. The AGC estimates were computed as the median AGC values obtained from the 3 calibrated AGC models.

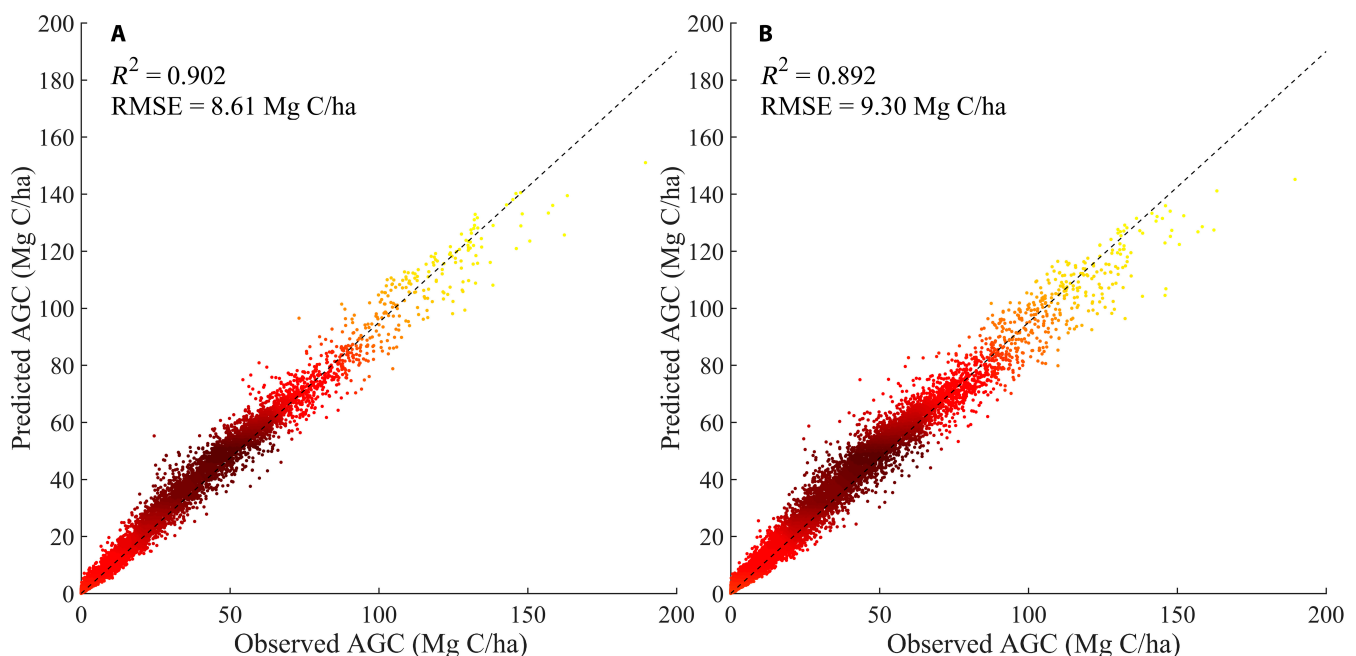
albeit with AGC losses of  $-0.14 \text{ Pg C}$  in 2017. Carbon losses ( $-0.02 \text{ Pg C year}^{-1}$ ) from areas between  $40^{\circ}\text{N}$  to  $50^{\circ}\text{N}$  of China (Northeast China) were offset by a carbon sink ( $+0.19 \text{ Pg C year}^{-1}$ ) in the rest of the regions. The largest carbon sink ( $+0.10 \text{ Pg C year}^{-1}$ ) was found around  $20^{\circ}\text{N}$  to  $35^{\circ}\text{N}$  of southern China (Fig. 10), which accounts for 56% of the total carbon sink in China. The carbon sink of southern China is mainly attributed to the Guizhou ( $+22.35 \text{ Tg C year}^{-1}$ ) and Sichuan ( $+14.49 \text{ Tg C year}^{-1}$ ), followed by the Hunan ( $+11.42 \text{ Tg C year}^{-1}$ ), Yunnan ( $+8.11 \text{ Tg C year}^{-1}$ ), Guangdong ( $+7.37 \text{ Tg C year}^{-1}$ ), and Guangxi ( $+5.22 \text{ Tg C year}^{-1}$ ) provinces. Guizhou is also the province having the largest carbon sink over China in the study period. The carbon sink of the Qinghai and Tibet provinces, as referred to as the “Tibetan Plateau”, ranked 4th ( $+9.79 \text{ Tg C year}^{-1}$ ) and 5th ( $+9.33 \text{ Tg C year}^{-1}$ ) among all provinces. Other regions supported by ecological restoration projects also showed a carbon uptake (Figs. 9 and 10): Loess Plateau (including the Shaanxi, Ningxia, and Shanxi provinces) has a net carbon sink of  $+0.03 \text{ Pg C year}^{-1}$ , and North

Shelter Forest (e.g., Xinjiang, Gansu, and Neimongol provinces) has a carbon increase of  $+0.01 \text{ Pg C year}^{-1}$ . It is worth noting that Northeast China (e.g., Heilongjiang, Jilin, and Liaoning provinces) showed a carbon losses of  $-0.02 \text{ Pg C year}^{-1}$  during 2013 to 2019.

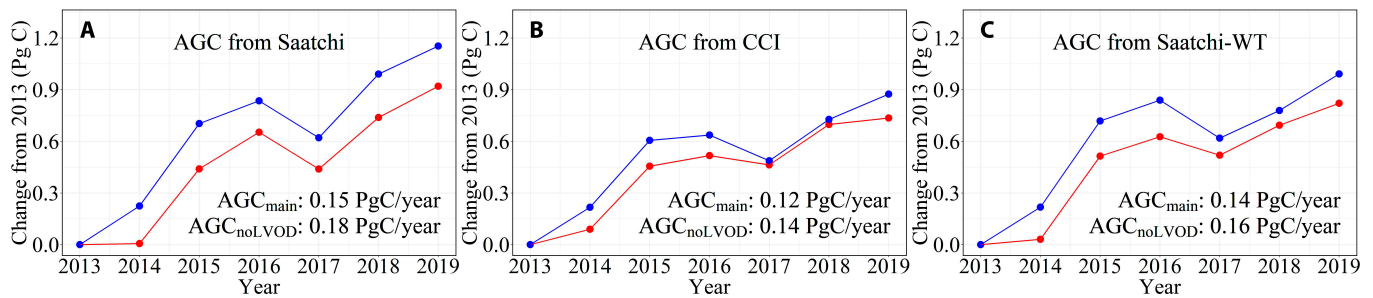
## Discussion

### Intercomparison between VIs/VODs and benchmark AGC maps

The highest correlation between VIs and AGC benchmark maps were found for the MODIS TC products (Fig. 3). The possible reasons could be that the TC product is not fully independent from the benchmark AGCs, given that the VIs such as NDVI and TC were used as inputs for the retrieval of benchmark AGCs [5,7,9], and these VIs are all derived from MODIS reflectance bands. Moreover, previous independent evaluations found errors in the MODIS TC product, such as overestimation in sparsely vegetated areas and underestimation in densely vegetated areas



**Fig. 5.** Performance of the 2 RF models using different predictors based on the Saatchi map, calculated using a bootstrap cross-validation method (sampling rate = 80%, iterations = 1000). (A)  $AGC_{main}$  using  $VOD_L$ ,  $VOD_{LPDR}$ , NDVI, LAI, and TC. (B)  $AGC_{noLVOD}$  using  $VOD_{LPDR}$ , NDVI, LAI, and TC. The colors show the density of scatter plots.



**Fig. 6.** Comparisons between  $AGC_{main}$  and  $AGC_{noLVOD}$  changes over the areas of overlap. The AGC changes were expressed by the AGC stocks relatively to 2013. AGC changes estimated by the benchmark AGC maps of (A) Saatchi, (B) CCI, and (C) Saatchi-WT.

[82,83]. Compared to ground reference data, previous studies showed that the MODIS TC product was affected by overestimations in croplands in regions such as South America and Southeast Asia [84] and was affected by severe underestimations in semiarid drylands [85].

Our results indicated that optical VIs such as NDVI and LAI have low correlations with benchmark AGCs and are saturated over dense vegetation with AGC values higher than 25 Mg C/ha (Figs. S1 and S2), in agreement with previous studies [52,55]. To compare the relationship linking VOD and AGC to the relationship between other VIs and AGC, scatter plots similar to those of Fig. 3 were computed using Saatchi's AGC with respect to NDVI and EVI (Fig. S3). It can be seen that there are no noticeable differences between NDVI and EVI, and both NDVI and EVI quickly saturate for AGC values above 50 Mg C/ha. These results are likely to be related to the fact that NDVI, EVI, and LAI are sensitive to the green vegetation cover, and not to the total AGC in dense vegetation [70]. However, VIs could provide better spatial coverage with high resolution and are available for long-term time series, which can provide effective global monitoring of vegetation trends [14]. A complementary use of VIs and VODs can thus allow for more complete monitoring of vegetation resources [51].

Among VODs, the spatial correlation between L-VOD and benchmark AGCs is higher than for the other VODs (Fig. 3A to C). This is due to the stronger penetration capacity of VOD retrieved at the L-band than at the C- and X-bands [46] and is thus more sensitive to the vegetation features of the whole canopy, including also the woody component [47,86]. This agrees with Li et al. [48] that showed that L-VOD maps are better correlated with AGC benchmark maps than higher-frequency VOD products (e.g., X-VOD and C-VOD) at the global scale.

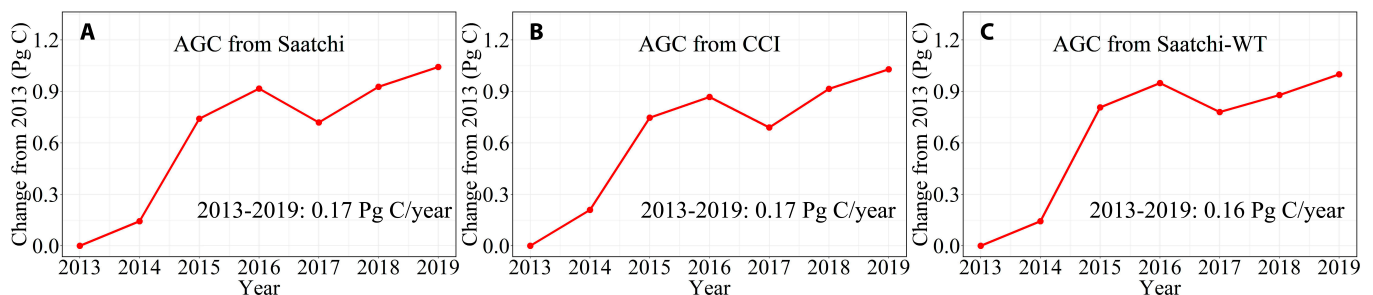
Note that both IB-VOD and LPDR-VOD tended to saturate for AGC values higher than 50 Mg C/ha (Figs. S1 and S2). This

result supports the findings of Chaparro et al. [54], who found that VOD at higher frequencies (e.g., X-band and C-band), as well as optical indices, saturates in dense vegetation in the tropics. This may lead to uncertainties in the AGC estimation using IB-VOD and LPDR-VOD in the tropical forests of southern China. Conversely, IB-VOD and LPDR-VOD products are more sensitive to the top of the canopy vegetation features, making them more suitable for monitoring temporal changes in the green vegetation components [46,48].

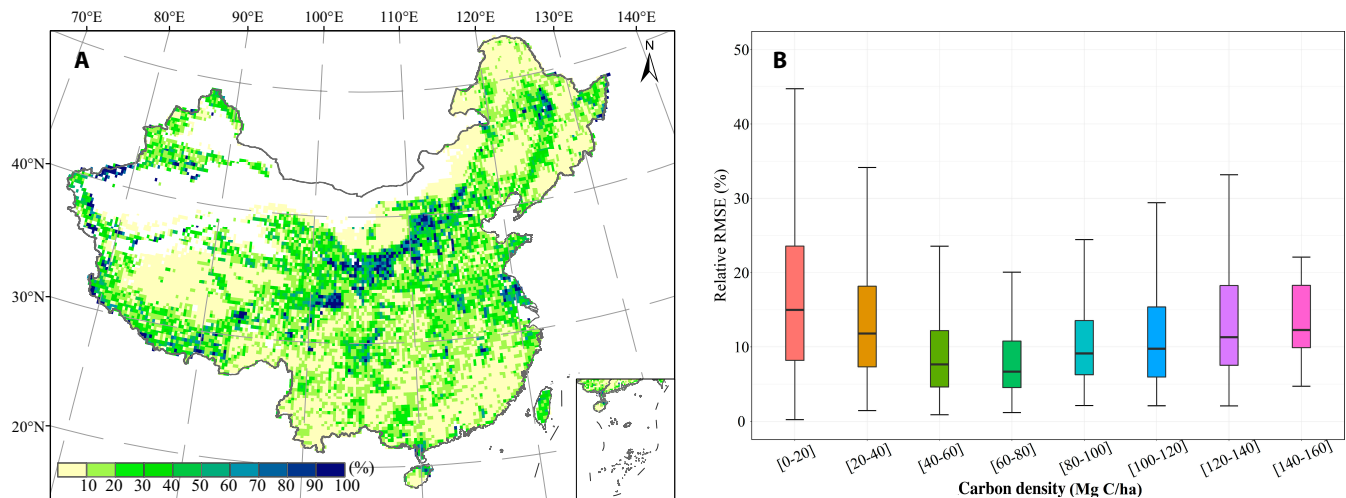
Regarding the different benchmark AGC maps, the best correlation values of benchmark AGCs with VIs and VODs were generally observed for the Saatchi-WT map (with a mean value of  $R = 0.78$ ), followed by the Saatchi map (with a mean value of  $R = 0.75$ ) (Table 2). This is in line with the findings of Chang et al. [77] showing that the Saatchi-WT map showed better accuracy than the other AGC products, when comparing them against in situ measurements. However, it is worth noting that all benchmark maps used information from MODIS VIs, due to the limited availability of independent sources of LiDAR, although GEDI can provide more observations for biomass in the following years [29]. The high correlation between the independent VODs and AGCs provide an alternative way to estimate AGC stocks, as proposed in this study.

### Continuous carbon sink over China

Our estimated carbon sink ( $+0.17 \text{ Pg C year}^{-1}$ ) during 2013 to 2019 are comparable to the values ( $+0.17 \text{ Pg C year}^{-1}$ ) obtained during the 2000s reported by Jiang et al. [11] (Table 3). Also, the recent estimates by Harris et al. [87] indicated that forests in China were a net carbon sink of  $+0.14 \text{ Pg C year}^{-1}$  during 2000 to 2019, which is close to our estimates for woody vegetation. Our estimates are comparable to recent estimates from Xu et al. [21] using multiple remote-sensing dataset, who reported a net carbon gain in live biomass of China of  $+0.17 \text{ Pg C year}^{-1}$  for



**Fig. 7.** AGC changes over China relatively to 2013 estimated by  $AGC_{merge}$  calibrated by different benchmark AGC maps. AGC changes estimated by the benchmark AGC maps of (A) Saatchi, (B) CCI, and (C) Saatchi-WT.



**Fig. 8.** Maps of the uncertainties of  $AGC_{merger}$  ( $Uncertainties_{pixel}$ ) in 2015 in terms of spatial patterns (A) and  $Uncertainties_{pixel}$  associated with AGC density (B).

2013 to 2019. These results confirmed that China has been a carbon sink in regards to biomass over the past 2 decades.

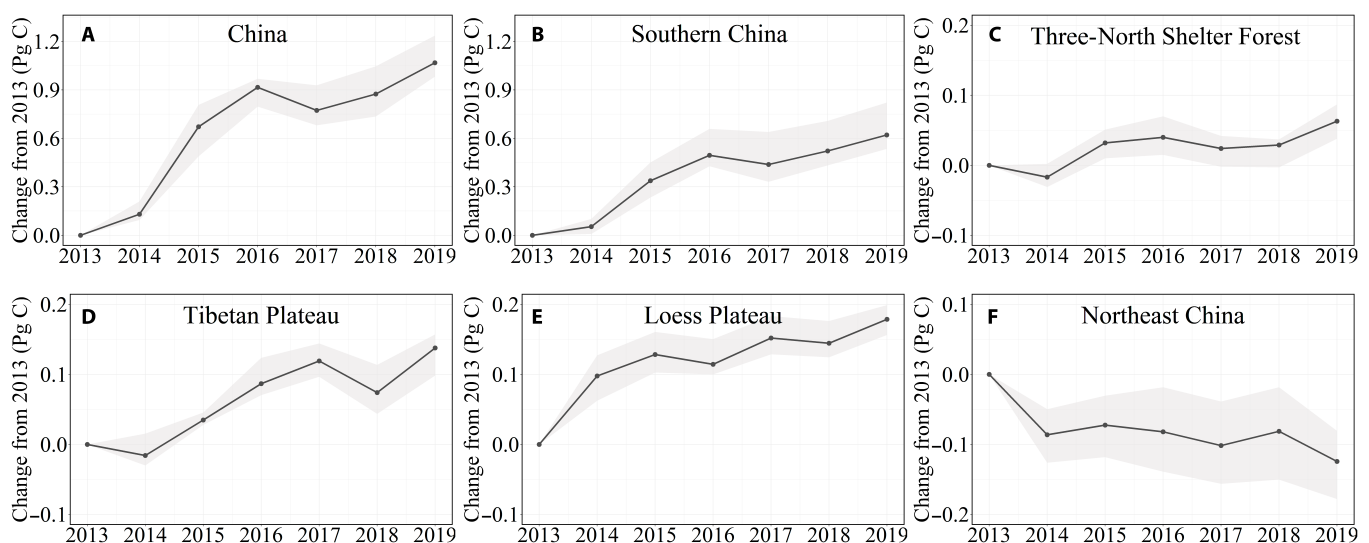
At the regional scale, in southern China, we found an increase of AGC of  $+0.10 \text{ Pg C year}^{-1}$ , which is quite similar to the estimate of  $+0.11 \text{ Pg C year}^{-1}$  obtained by Tong et al. [25] for 2002 to 2017 (Figs. 9 and 11). This also supports the results of Brandt et al. [88] and Tong et al. [89], which have found increased vegetation growth and carbon stocks in southern China during the recent decade. In southern China, our results suggested that the regions with the largest carbon sinks are the Guizhou, Hunan, and Guangdong provinces, in line with the findings of Wang et al. [12]. This can be explained primarily by increasing areas of newly planted forests, as a consequence of intensive national afforestation/reforestation programs implemented in the past few decades [63,90].

In northeast China, our results indicated a carbon loss of  $-0.02 \text{ Pg C year}^{-1}$ . In that region, Piao et al. [2] estimated a small carbon source ( $-0.004 \text{ Pg C year}^{-1}$ ) during the 1980s and 1990s using both inventory-satellite-based and atmospheric inversion

methods. This carbon loss was probably caused by overharvesting and degradation of forests [2,78].

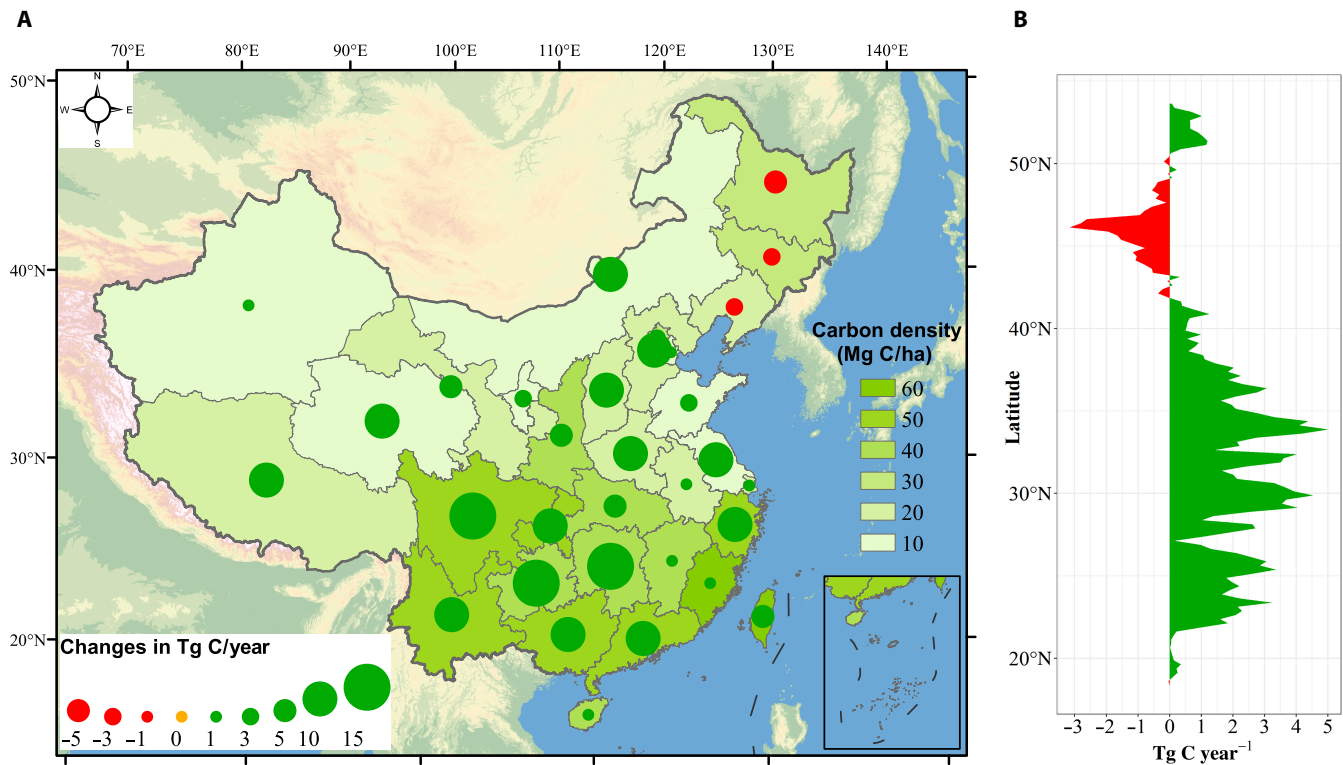
Our estimates of the carbon sink are lower than the estimates using inversion of atmospheric data. For example, the inversion results constrained by ground-based and space-based measurements of atmospheric  $\text{CO}_2$  indicated a net terrestrial  $\text{CO}_2$  sink in China of about  $0.30$  to  $0.43 \text{ Pg C year}^{-1}$  in 2010 to 2015 [91–93]. Relative to our results, this larger carbon sink values could be explained by the fact that our estimates do not account for the contribution for carbon sink from belowground vegetation biomass and soil organic matter.

Relative to the results for the 1980s and 1990s, the magnitude of the carbon stock increases presented in this study are higher than those reported by previous estimates (Table 3). For example, Fang et al. [1] estimated the carbon sink of vegetation biomass in China at  $[+0.096, +0.106] \text{ Pg C year}^{-1}$  between 1981 and 2000 using forest inventory data. Piao et al. [2] estimated the terrestrial carbon sink of China at  $+0.177 \text{ Pg C year}^{-1}$  for the same period, of which 58% ( $+0.105 \text{ Pg C year}^{-1}$ ) was attributed to



**Fig. 9.** Temporal variations in annual AGC (Pg C), expressed as the difference from 2013 values. (A) Annual variations in AGC in (A) China, (B) Southern China, (C) Three-North Shelter Forest, (D) Tibetan Plateau, (E) Loess Plateau, and (F) Northeast China.





**Fig. 10.** Spatial patterns of the net AGC dynamics across China during 2013 to 2019. (A) Average AGC density (in Mg C/ha) and net AGC change for 2013 to 2019 summed for each province (in Tg C year<sup>-1</sup>). (B) Net AGC change for 2013 to 2019 (summed per latitude). The corresponding pixel-scale AGC dynamics are shown in Fig. 11.

vegetation biomass. Our estimations of the biomass carbon stock increases are also higher than the estimates from Pan et al. [90], who reported a forest carbon sink of  $+0.12 \text{ Pg C year}^{-1}$  in China during 2000 to 2007 using inventory data. This difference suggests that forest biomass carbon gains in the 2010s substantially increased in comparison to the 1980s and 1990s in China.

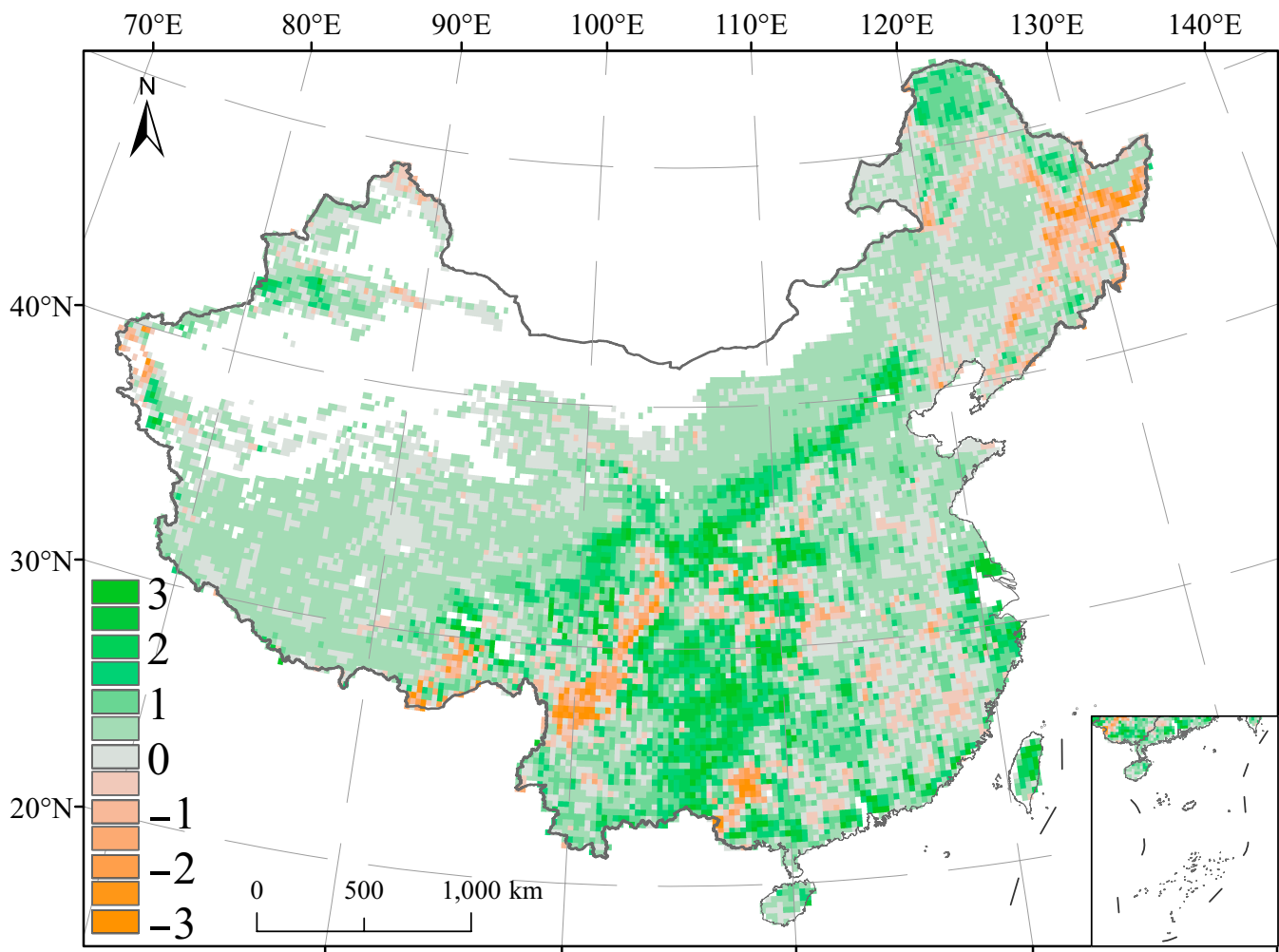
### Limitations of this study

High uncertainties in our AGC estimates were observed in regions with low AGC density. This is the case also in other

studies analyzing remote-sensing AGC retrievals. For example, AGC benchmark maps mainly derived from GLAS LiDAR observations have high uncertainties at low AGC density [7]. This could be explained by the fact that GLAS LiDAR shots only marked forest canopies as valid observations; thus, the regions with low AGC density (e.g., nonforest regions) could be inaccurately estimated [21]. However, considering the limited contribution of areas with low AGC density to the AGC dynamics over China, the impact of the low performance of AGC estimates at low AGC density is limited.

**Table 3.** Carbon sinks ( $\text{Pg C year}^{-1}$ ) for China documented in the scientific literature.

| Category   | Carbon sink       | Period      | Method        | Reference          |
|------------|-------------------|-------------|---------------|--------------------|
| Vegetation | 0.096–0.106       | 1980s–1990s | Inventory     | Fang et al. [1]    |
|            | $0.105 \pm 0.048$ | 1980s–1990s | Inventory     | Piao et al. [2]    |
|            | $0.092 \pm 0.074$ | 1980s–1990s | Process model | Piao et al. [2]    |
|            | $0.17 \pm 0.06$   | 2000s       | Inventory     | Jiang et al. [11]  |
|            | 0.17              | 2013–2019   | Maps          | Xu et al. [21]     |
| Forest     | $0.075 \pm 0.035$ | 1980s–1990s | Inventory     | Fang et al. [1]    |
|            | $0.13 \pm 0.04$   | 2000s       | Inventory     | Jiang et al. [11]  |
|            | 0.104             | 2000s       | Inventory     | Jiang et al. [11]  |
|            | 0.115             | 2000s       | Inventory     | Pan et al. [90]    |
|            | 0.14              | 2000–2019   | Maps          | Harris et al. [87] |
|            | 0.17              | 2013–2019   | VOD and VI    | Our study          |



**Fig. 11.** Net AGC change in China during 2013 to 2019 ( $\text{Mg C ha}^{-1} \text{ year}^{-1}$ ). Note: Negative values are carbon sources.

To gap-fill the missing observations of L-VOD (Fig. S6), a merging model was proposed to produce a more spatially complete coverage of the China AGC product. Effects of RFI on L-VOD, which vary in both space and time, could affect our estimates of the interannual variations of AGC. Meanwhile, the effects of RFI on L-VOD leads to missing observations and limits its applicability for monitoring seasonal variation of AGC over some regions in China [6]. However, the RFI effects on L-VOD have strongly decreased over China since 2016 [47].

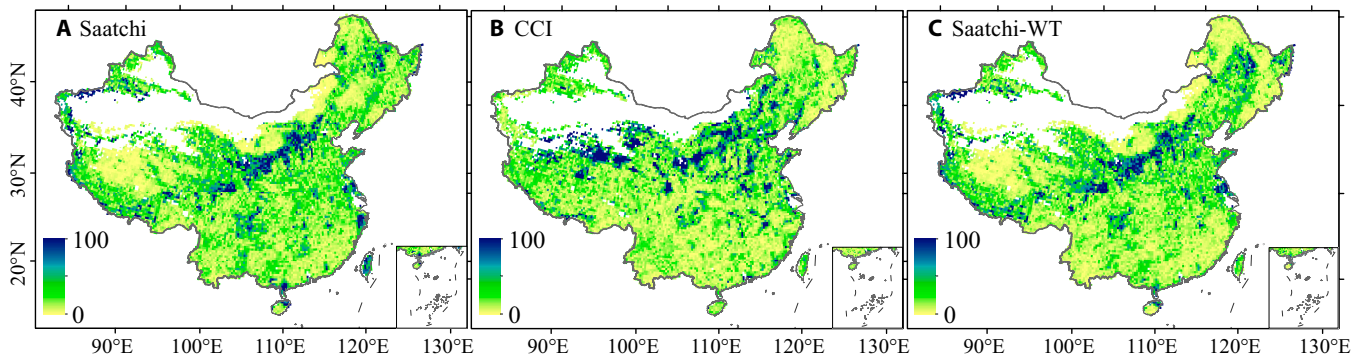
The coarse spatial resolution ( $0.25^\circ$ ) of our AGC product limits its applicability for monitoring fine-scale carbon losses and gains, e.g., related to disturbance regimes. As different land cover types are often mixed within a single pixel, the carbon change of AGC observed in a  $0.25^\circ$  grid cell comprises all land covers including forests and nonforest ecosystems, representing a mix of different processes (e.g., deforestation, forest degradation, and reforestation). Our pixel-level estimates of AGC sinks are admittedly conservative as a result of the coarse spatial resolution of the VOD product, which inevitably represents averaged estimates.

Considering the large spatial mismatch between AGC estimates ( $0.25^\circ$ ) and point-scale measurements, direct evaluation of the AGC estimates compared with in situ measurements will result in large bias, especially over strongly mixed  $0.25^\circ$  pixels, e.g., mixed pixels including dense and low-density forests. A

bootstrapped cross-validation method was used here to evaluate the uncertainties associated with the AGC estimates, but further independent evaluations using local carbon density estimated by airborne LiDAR are needed to better evaluate the performance of our AGC model. It is also noteworthy that the uncertainties of the AGC estimates computed from different calibration models were inconsistent (Fig. 12). For example, the calibrated AGC models have different uncertainties based on the Saatchi (Fig. 12A) and CCI benchmark maps (Fig. 12B), where uncertainties larger than 100% covered 6.6% and 10.9% of China, respectively. The higher uncertainties ( $>100\%$ ) in our AGC estimates were mainly observed in northern China and the eastern Tibetan Plateau and may be related to topography in those mountainous regions. Areas of high uncertainties in the AGC estimates calibrated with the CCI map displayed a larger coverage than for the others benchmark maps. In the most southern and southwestern regions of China, the majority of the uncertainties of the AGC estimates are smaller than 10%.

## Conclusions

This study provided the first comprehensive assessment of optical VIs and microwave VODs for monitoring spatial patterns of AGC in China, demonstrating the capabilities in the synergistic use of VODs and VIs to provide an explicit quantification



**Fig. 12.** Maps of the uncertainties associated with  $AGC_{merge}$  ( $Uncertainties_{pixel}$ ) relative to different benchmark AGC maps in 2015. Maps of  $Uncertainties_{pixel}$  relative to the benchmark AGC maps of (A) Saatchi, (B) CCI, and (C) Saatchi-WT.

of the AGC dynamics across large scales. It is worth noting that the validation for the estimated AGC product is not based on measured AGC, limited by the availability of in situ measurements at a large scale (e.g., larger than  $0.25^\circ$ ). Thus, further independent evaluations using local carbon density estimated by airborne LiDAR or forest inventory data are needed to validate our results. Especially, with the recent and future launches of Chinese HJ-1C [94], Britain's NovaSAR-S [95], NASA's GEDI [29], ICESat-2 [96], NISAR [97], and ESA's BIOMASS [75,98], more biomass products will be good candidates for improving the accuracy of continental AGC maps following the approach presented in this study. In future, the carbon dynamics found by this study will be analyzed using ancillary data on climate changes and forest disturbances, to better understand and reveal the drivers of China carbon dynamics.

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## Data Availability

The data that support the findings of this study are available at National Tibetan Plateau Data Center, DOI: 10.11888/terre.tpdc.272887.

## Supplementary Materials

Supplementary data associated with this article can be found in Supplementary Materials. Supplementary Text. Figs. S1 to S6. Tables S1 and S2.

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